Microbiological and Physico-chemical Characteristics of Municipal 1 Wastewater at Treatment Plants, province Sharkia, Egypt 2 (Case study) 3

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ABSTRACT

15 The present study was conducted to evaluate the microbiological and physico-16 chemical characteristics of effluent produced from 17 Wastewater Treatment Plants (WTPs) distributed in province Sharkia before discharge in the drainage to control the 17 pollution and their disposal options. Total bacterial count (TBC), total yeasts & 18 19 moulds count (TYMC), total Candida count (TCC), total coliform count (TCFC), 20 Escherichia coli (EC), Salmonella and Shigella (SS) count) were analyzed for 21 untreated wastewater (UW), aeration treatment wastewater (ATW), oxidation 22 treatment wastewater (OTW), anaerobic treatment wastewater (ATW) and treated 23 wastewater (TW). Physicochemical parameters (Temperature, pH, Total Suspended 24 Solids (TSS), Total dissolved Solids (TDS), Biochemical Oxygen Demand (BOD), 25 Chemical Oxygen Demand (COD), Nitrate (NO₃-), sulphite (SO₄-) and oil contents of UW and TW in the WTPs were examined in different seasons. The treatment plant 26 27 received the municipal wastewater from various Sewage Pumping Stations being 28 treated through different stages viz. primary (physical), secondary (chemical) and 29 tertiary (disinfection) treatments. The results revealed that the wastewater was heavily 30 contaminated with cultivable bacteria and inorganic & organic pollution. The 31 coliform bacteria were correlated indicators of a reduction in pathogenic bacteria 32 concentrations during the wastewater treatment, but were not correlated to Candida 33 contamination of wastewater and drainage samples. The TBC, TYMC, TCC, TCFC, 34 EC and SS were significantly (p < 0.05) decreased in treated water. The maximum removal of TBC (60%), TYMC (59 %), TCC (75 %), TCFC (77%), EC (75%) and SS 35 36 (74%) of treated wastewater were observed after the finally treated. The TSS, TDS, 37 BOD, COD, sulphite, nitrate and oil levels were significantly (p <0.05) decreased in 38 wastewater after the finally treated. The maximum removal of pH (6%), BOD (90%), 39 COD (89%), TSS (88%), and SO₄- (86%), of treated wastewater were recorded after 40 finally treatment. The results indicated that the treatment plants had a significant role 41 in the control of pollution load from microbial, organic and inorganic pollution at 42 province Sharkia, Egypt. The results conclude that microbiological parameters are 43 essential to monitor the correct WTP operation and we propose quantification of 44 *Candida* as indicator of wastewater microbiological quality.

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46 Keywords: Microbiological, *Salmonella*, *Candida*, sewage water, TDS, BOD, COD.

48 **1. Introduction**

49 Actually, there remain major uncertainties about the implications of possible wastewater safety under a changing climate. Climate change can be expected to 50 51 present a variety of new challenges in the area of wastewater treatment on middle and long-term timeline. Evidence of the impact of climate change on the transmission of 52 53 waterborne diseases has become clear [1]. Climate change is a significant impact on 54 land surface water availability, decreasing by 20% according to Mariotti et al., [2]. 55 The Mediterranean is a 'hot spot' for climate change, an increase in the average 56 annual temperature between +3.5 °C and +3.9 °C[3]. The Intergovernmental Panel on 57 Climate Change scenario, the average global air temperature should increase between 58 1.8 and 4.0 °C [4] during the 21st century and this increase might effect on wastewater 59 treatment. Moreover, a drying tendency in summer is expected, particularly in subtropics, low and mid-latitudes, in addition with an extreme events increase in 60 61 general. The vulnerability assessment of water resources in Egypt to climatic change 62 in the Nile Basin was reported [5]. Irrigation of agricultural lands with wastewaters, 63 following varying levels of treatment, is increasing around the world [6, 7, 8, 9].

64 Pollution from wastewater is currently the greatest threat to the sustainable use of 65 surface and groundwater in the megacity. Household, commercial, and industrial effluents and raw untreated sewage are often discharged into the open and fresh-water 66 sources such as the majority of villages and rural areas discharge their raw domestic 67 68 wastewater directly into the waterways in the most of developing countries. The wastewater eventually percolates or is washed into the water bodies by rainstorms. 69 70 The stagnating pools of wastewater in the open gutters and on the roads often provide 71 the breeding grounds for mosquitoes and habitat for several bacteria and viruses. In 72 addition, wastewater pools contain hazardous contaminants such as oil and grease, 73 pesticides, ammonia, and heavy metals [10]. When point source pollution is reduced 74 in many countries (even if wastewater treatment plants begin to reach their capacity 75 limits), climate (global) change impacts could tend to increase the diffuse pollution 76 with for example urban or agricultural runoff. The climate change determinants 77 affecting water quality are mainly the ambient (air) temperature and the increase of 78 extreme hydrological events. Soil drying-rewetting cycles and solar radiation increase 79 may also be considered.

80 Waterborne pathogens could be spread within the freshwater after a contamination by 81 animal or human waste due to heavy rainfall discharge in combined sewer systems 82 (CSS). When the flow exceeds the CSS capacity, the sewers overflow directly into 83 surface water body [11]. Coliform load in a tidal embayment was studied and shown 84 that storm water coming from the surrounding watershed is a primary source of coliform [12]. Moreover, higher water temperatures will probably lead to a pathogen 85 86 survival increase in the environment, although there is still no clear evidence [13]. 87 Half of the waterborne disease outbreaks in the US during the last half century

88 followed a period of extreme rainfall [14]. Even though the risk of diseases outbreaks linked to mains drinking waters is low in developed countries, private supplies would 89 90 be at risk [13], and even properly constructed onsite wastewater treatment systems may cause a waterborne outbreak [15]. In addition, an increase in temperature threats 91 92 water quality with regard to waterborne diseases especially cholera disease in Asia, 93 Africa and South America [13]. Lastly, it was shown that with increased UV radiation 94 due to ozone layer depletion, NOM trap higher levels of UV energy and breaks down 95 to more bioavailable organic compounds, minerals and micronutrients. All these 96 processes could stimulate bacterial activity in aquatic ecosystems [16]. The prevalence 97 of pathogenic microbes in treated wastewater has raised concerns about the capacities 98 of existing treatment to remove theses microbes [17]. In Egypt, according to FAO 99 [18], at present, wastewater is estimated at 4930 billion m³/yr. However, the total capacity of the installed treatment plants amounts to about 1.752 billion m³/yr. The 100 101 whole wastewater reuse in agriculture is about 0.2 billion m3/year [19]. At present, 102 there are more than 239 wastewater treatment plants (WTPs) in Egypt in 2012 and 35 103 WTPs of the total in province Sharkia. Urban coverage with improved sanitation 104 gradually increased from 45% in 1993 to 56% in 2004. The WTPs are treating an 105 average of 10.1 million cubic meters per day [20], serving more than 18 million 106 people. The number has increased 10 times between 1985 and 2005 [21]. The amount 107 of water which is released into the Nile is 3.8 billion m³ per year, out of which only 108 35% was treated properly as of 2004.

109 Recently, the average log removals after treated wastewater by three different pilot-110 scale sand filters were 2.2-3.5 for pathogenic human noro- and adenoviruses and 4.3-5.2 and 4.6-5.4 log CFU/ml for indicator viruses and bacteria, respectively. The 111 112 system that effectively removed microbes was also efficient at removing nutrients 113 [22]. Thus, this study assessed the performance of 17 Wastewater Treatment Plants 114 (WTPs) in cold and hot climate over a one-year period from April 2012 to March 115 2013 in province Sharkia, Egypt. The WTPs, traditionally, rural household wastewater has been treated by two or three separate septic tanks that provide primary 116 treatment, secondary treatments and a soil absorption field for further treatment. The 117 118 main purpose was to examine the changes in physical, chemical and microbiological 119 quality in wastewater during treatment operations. The function of the systems 120 evaluated in terms of their ability to remove nutrients, organic and microbial loads. 121 The second aim was to evaluate the quality of wastewater in drainages that discharges 122 from the WTPs.

123 **2. Materials and methods**

124 2.1. Study area

The area of case study is located in Sharkia governorate, Egypt. The Sharkia area locates at latitude 30.7 °N and 31.63 °E longitudinal at an evaluation of 10 m above the mean sea level. The TW discharges into the Belbeis Drain (BD) and then into Bahr El Baqar Drain (BEBD), which in turn drains to Lake Manzala 170 km away from Cairo. The drain and Lake Manzala had been identified as "black spots" by theEgyptian Environmental Action Plan back in 1992.

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132 2.2. Water sampling

133 The study was conducted in 17 WTPs located in the megacities in Sharkia, Egypt. They served about 6.884.000 populations, receive about 387.000 m³/y and indirect 134 135 discharge about 138.000 m³/y. Other 10 sites, the wastewater samples were collected 136 from BEBD located in Sharkia, Egypt during seasons of 2012-2013. The samples were 137 collected each month at the same location between 8 am and 11 am in a sterile Schott 138 glass bottle from the TMW, BD and BEBD at the same point. Water was obtained 139 from areas of fast flow at a depth half that of the total in order to avoid debris and 140 collecting exclusively surface water. The samples were placed in a container filled 141 with ice, then transported to the microbiological laboratory, and stored at 4 °C prior to 142 analysis.

143 2.3. Microbiological analysis

144 Wastewater samples (10 ml) were aseptically pipetted into a sterile Erlenmeyer flask 145 and diluted tenfold by adding 90 ml of in sterile buffered peptone water (BPW: 146 peptone 1 g/l, pH 7.4) followed by subsequent decimal dilution (up to 10^{-7}) using the 147 BPW. Total bacterial counts (TBC) for wastewater samples were conducted in 148 triplicate according to the American Public Health Association [23]. using plate count 149 agar and incubated at 30 °C for 48 h. Results are expressed as the mean (log₁₀) with the calculated standard error indicated. Total coliforms, 1.0 ml from dilution sample 150 was poured in sterile Petri dish then poured 10 ml of Violet Red Bile Dextrose Agar 151 152 (Biolife 402188). After solidifying media, a 10 ml overlay of the same molten 153 medium was added. The incubation was carried out at 37 °C for 24 h. For Escherichia 154 *coli*, the detection was done by using the selective Chromo Cult Coliform agar (Merck 155 KGaA, Germany) according to the manufacturer's instructions and confirmed with 156 Kovac's indole reagent. Yeasts and moulds were detected onto Rose Bengal 157 Chloramphenicol Agar (Lab M, 36, supplemented with chloramphenicol, X009) at 25 °C for 5 days; Candida counts were counted on Candida Agar (Biolife, 4012802, 158 159 Milano, Italy) by spreading 0.1 ml of sample onto media and incubated at 37 °C for 48 160 h. All plates were examined for typical colony types and morphological 161 characteristics associated to each culture medium. Salmonella and Shigella were 162 counted on Salmonella & Shigella Agar (SS Agar, LAB052, UK) after incubation for 163 24 h at 37 °C.

164 **2.4.** Data collection and statistical analysis

In this research, some physicochemical and bacteriological data, routinely 165 166 experimented each week by Holding Company for Water and Wastewater that used to 167 evaluate the treated wastewater quality at the Wastewater treatment plants. These 168 parameters include: Temperature, pH, Total Suspended Solids (TSS), Turbidity, Total dissolved Solids (TDS), Biochemical Oxygen Demand (BOD), Chemical Oxygen 169 170 Demand (COD), Nitrate (NO₃-), sulphite (SO₄-) and oil [24]. Holding Company for 171 Water and Wastewater have established 35 stations along the entire megacity e.g. 172 Zagazig city has 3 stations. Wastewater quality data interpretations of these stations 173 and drains were conducted in a period of one year from April 2012 to March 2013. 174 The removal efficiency of each treated wastewater sample in the wastewater treatment 175 plants was calculated as [(influent- effluent)/influent x 100]

176 All analyses were performed in three replicates. The results were expressed by 177 the mean of the two samples plus the standard error. Data were statistically analyzed 178 using ANOVA through the general linear models (GLM) procedure of the statistical 179 analysis system software (SAS version 9.1, SAS Institute, Inc., 2003). Least 180 significant differences were used to separate means at p < 0.05.

181 **3. Results and discussions**

182 3.1. Influent, effluent characteristics and microbial indicators removal efficiency

183 The level of total bacterial count (TBC), total yeasts & moulds count (TYMC), total 184 Candida count (TCC), total coliform count (TCFC), Escherichia coli (EC), 185 Salmonella and Shigella (SS) counts in wastewater samples collected from different 186 treatment processes in 17 WTPs are presented in Figs 1&2 and Tables 1&2. The 187 pathogenic bacteria and microbial indicators are used to evaluate WTPs through one 188 year 2012-2013. The results revealed that the influent in the WTPs was heavily 189 contaminated with cultivable bacteria and yeasts. The TBC, TYMC, TCC, TCFC, EC 190 and SS were significantly (p < 0.05) decreased in the TW during all the period of 191 study. Among the many kinds of wastewater disinfection, chlorination has gained wide 192 acceptance commercially, because of its simple application and moderate cost [25]. 193 However, the counts were varying from a minimum of 3.1 log CFU/ml to maximum 194 9.2 log/CFU/ml, from 2.1 to 5.76 log CFU/ml, from 1.0 to 4.47 log CFU/ml, from 1.2 195 to 5.86 log CFU/ml, from 1.2 to 5.1 log CFU/ml and from 1.5 to 5.71 log CFU/ml 196 (Figs. 1&2), respectively. The average log removals after treated wastewater by WTP 197 systems were 4.71 (58.08%), 2.87 (56%), 3.20 (57.87%), 2.33 (49.44%), 3.55 (66.03) 198 and 1.97 (59.51%) log CFU/ml for TBC, TYMC, TCFC, EC, SS and TCC counts, respectively (Table, 1). The maximum removal of TBC was (60%), TYMC (59 %), 199 200 TCC (75 %), TCFC (77%), EC (75%) and SS (74%) of TW in August, October and September 2012. Coliforms, E. coli and Salmonella spp. have been accepted as 201 202 contamination indicator bacteria in treated wastewater [26]. Moreover, seasonal 203 conditions appear to have a clear effect on purification efficiencies, emphasising the 204 strongest of these systems especially in hot climates [22]. However, the capacity of 205 wastewater treatment plants has not enough to counteract increased of domestic 206 wastewater. The reduction in microbial groups may have been influenced by the 207 seasonal changes and the volume of receiving stream [27]. The average log removals 208 after treated wastewater by three different pilot-scale sand filters were 4.6-5.4 log 209 CFU/ml for bacteria. The system that effectively removed microbes was also efficient 210 at removing nutrients. The coliform bacteria were correlated (r= 0.83) indicators of a 211 reduction in pathogenic bacteria concentrations during the wastewater treatment, but 212 were not correlated (r=-0.33) to Candida contamination of wastewater (Table, 2). 213 Total coliforms load in a tidal embayment was studied and shown that storm water 214 coming from the surrounding watershed is a primary source of coliform [28]. The 215 presence of coliforms is usually assumed to indicate the potential presence of other fecal pathogens such as Salmonella spp., Shigella spp. or pathogenic strains of 216 217 Escherichia coli [29]. These organisms can cause gastroenteric illnesses via the 218 fecal/oral route through the consumption of raw produce irrigated with contaminated 219 water. Moreover, higher water temperatures will probably lead to a pathogen survival 220 increase in the environment, although there is still no clear evidence [13]. Its logic 221 after tertiary treatment, the pathogens could be absent but in this study the wastewater 222 treatment received domestic water more than its capacity and the conventional tertiary 223 treatment do not applied correctly. In Europe, Salmonella spp. was more rarely

detected (16.3%) of the reclaimed wastewater and *Campylobacter* cells were only found in 2% of samples [30].

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227 *3.2. Microbial indicators in drainages*

228 The levels of bacterial indicators in 10 sites on BEBD located in Sharkia were more less varying from a minimum of 6.8 log CFU/ml to maximum 9.7 log/CFU/ml, from 229 4.6 to 6.06 log CFU/ml, from 4.0 to 5.49 log CFU/ml, from 4.52 to 5.66 log CFU/ml, 230 from 4.32 to 5.46 log CFU/ml and from 5.15 to 5.85 log CFU/ml, respectively (data 231 232 not shown). This high levels of microbes resulted in the majority of villages and rural 233 areas discharge their raw domestic wastewater directly into the waterways. The 234 discharges are increasing year after year due to the population growth as well as the 235 rapid implementation of water supply networks in many villages without the parallel 236 construction of sewage systems. In Egypt, the increasing population, urbanization and 237 industrialization has resulted in a large proportion of mostly rural communities 238 lacking adequate sanitation, waste disposal and access to safety wastewater. When the 239 flow exceeds the combined sewer systems capacity, the sewers overflow directly into 240 surface water body [11].

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242 3.2. Influent, effluent characteristics and nutrients removal efficiency

243 The concentrations of COD, BOD and TSS contents in the wastewater samples 244 collected from different treatment processes in the WTPs are shown in Fig. 3. All 245 targeted parameters were detected in influent and effluent samples during at least two 246 of the sampling months throughout the year which they showed higher concentrations 247 in cold seasons than hot seasons. In the stage for treating aeration caused a significantly (p<0.05) reduction of COD and BOD contents between influent and 248 249 effluent. The targeted parameters (minimum to maximum, mg/l: COD (441 to 541), 250 BOD (367 to 421) and TSS (289 to 320) were detected in all influent samples (Fig. 3). 251 These targeted parameters were decreased significantly (p<0.05) in all effluent 252 samples, mg/l: COD (56-62), BOD (34-46) and TSS (52-82). The removal efficiency 253 was of COD (86.95- 89.22%)), BOD (88.83-91.56%) and TSS (74.83- 82.73%). 254 These results indicated that the elimination of organic compounds in WTPs was 255 incomplete and that more than of effluent are discharge to waterways. It was shown that with increased UV radiation due to ozone layer depletion, NOM trap higher 256 257 levels of UV energy and breaks down to more bioavailable organic compounds, 258 minerals and micronutrients in water. All these processes could stimulate bacterial 259 activity in aquatic ecosystems [16].

260 The concentrations of temperature, pH, TDS, nitrate (NO₃), sulphite (SO₄) and oil contents in the wastewater samples collected from different treatment processes in the 261 262 WTPs are shown in Table 3. These values were used to assess treated wastewater 263 characteristics before discharge in waterways. Seasonal trends in mass removals were 264 observed for all parameters. Monthly average influent temperature and pH ranged 265 from 18 to 29 °C and 7.6 to 7.9, while average effluent temperature ranged from 17 to 266 28 °C and 7.4 to 7.6, respectively (Table 3). There was no significant difference (p> 267 0.05) in the level of TDS and NO₃- in influent compared to the initial values recorded during all the period of study. However, the slight decrease in the NO⁻³ and TDS 268 269 levels in effluent and this might be WTPs are lack to the efficiency of removal nitrate 270 and TDS during tertiary treatments. The SO⁻⁴ and oil levels were significantly (p 271 <0.05) decreased in treated wastewater. The maximum removal of pH (3.02%), TDS 272 (2.76%), NO-3 (2%) and SO4- (89.85%) of treated wastewater compared to the initial 273 values. Seasonal differences were observed in effluent NO₃- rates and COD, with the 274 highest values in cold calamite than hot climate. The targeted removal rate 275 efficiencies greater than 75% were achieved for TSS, COD, BOD, SO₄- and oil levels. 276 The efficiency of aeration on total carbon conversion rates depends on the 277 bioavailability of easily degradable organic substances, on the abundance, 278 composition and activity of microbial groups involved in degradation processes and 279 on pre-existing environmental conditions such as oxygen supply, pH, temperature, 280 water and nutrient concentrations. A higher carbon conversion rate under aerated than 281 anaerobic conditions is attributed to aerobic microbial groups being able to convert 282 semi-degradable and hardly-degradable organic substances such as lignin which are 283 resistant to anaerobic microbial break-downs [31].

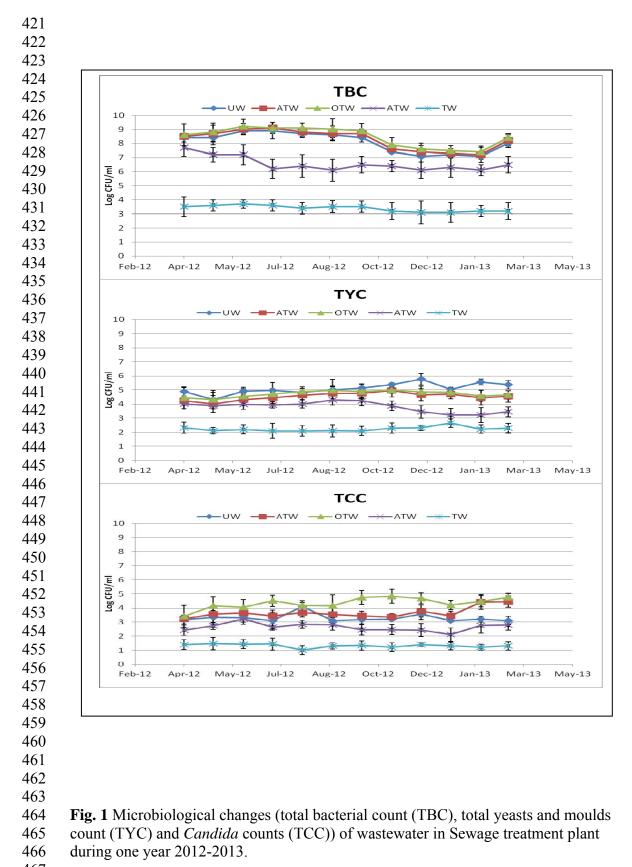
- 284 **References**
- [1] A.J. McMichael, D.H. Campbell-Lendrum, C.F. Corvalan, K.L. Ebi, A. Githeko,
 J.D. Scheraga *et al.* Climate change and human health WHO, WMO & UNEP,
 Geneva (2008).
- [2] A. Mariotti, N. Zeng, J-H. Yoon, V. Artale, A. Navarra, P. Alpert, L.Z.X. Li, Mediterranean water cycle changes: transition to drier 21st century conditions in observations and CMIP3 simulations Environmental Research Letters, 3 (2008), p. 044001.
- [3] F. Giorgi, Climate change hot-spots. Geophysical Research Letters, 33 (2006), p.
 L08707.
- [4] B.C. Bates, Z.W. Kundzewicz, S. Wu, J.P. Palutikof Climate change and water
 Technical paper of the Intergovernmental Panel on Climate change. IPCC
 Secretariat, Geneva (2008).
- [5] M.S. Kenneth, N. David, D.E.D. El Quosy, Vulnerability assessment of water
 resources in Egypt to climatic change in the Nile Basin. Clim Res, 6 (1996) 89–
 95.
- 300 [6] G. Barkle, R. Stenger, P. Singleteon, D. Painter, Effect of regular irrigation with
 301 dairy farm effluent on soil organic matter and soil microbial biomass Aust. J. Soil
 302 Res., 38 (2000), pp. 1087–1097.
- F. Papadopoulos, G. Parissopoulos, A. Papadopoulos, A. Zdragas, D. Ntanos, C.
 Prochaska, I. Metaxa, Assessment of reclaimed municipal wastewater application on rice cultivation. Environ. Manage., 43 (2009), pp. 135–143.
- 306 [8] M. Arienzo, E.W. Christen, W. Quayle, A. Kumar, A review of the fate of
 307 potassium in the soil-plant system after land application of wastewaters J. Hazard.
 308 Mater., 164 (2009), pp. 415–422
- 309 [9] K.P.M. Mosse, A.F. Patti; R.J. Smernik; E.W. Christen; T.R. Cavagnaro,
 310 Physicochemical and microbiological effects of long- and short-term winery waste
 311 wastewater application to soils.
 312 Journal of Hazardous Materials, 201–202,(30) 2012, 219–228.
- 313 [10] J.K. Saliu and O.J. Eruteya, Biodiversity of gutters in Lagos Metropolis,
 314 Nigeria. Journal of Biological Sciences, 6 (5) 2006, 936–940.
- [11] D.F. Charron, M.K. Thomas, D. Waltner-Toews, J.J. Aramini, T. Edge, R.A.
 Kent et al, Vulnerability of waterborne diseases to climate change in Canada: a
 review. J Toxicol Environ Health Part A, 67 (2004), pp. 1667–1677.

- [12] A.M. Pednekar, S.B. Grant, Y. Jeong, Y. Poon, C. Oancea, Influence of climate change, tidal mixing, and watershed urbanization on historical water quality in Newport Bay, a saltwater wetland and tidal embayment in southern California.
 Environ Sci Technol, 39 (2005), pp. 9071–9082.
- [13] P.R. Hunter, Climate change and waterborne and vector-borne disease J Appl
 Microbiol, 94 (2003), pp. 37S–46S.
- [14] F.C. Curriero, J.A. Patz, J.B. Rose, S. Lele, The association between extreme
 precipitation and waterborne disease outbreaks in the United States, 1948–1994
 Am J Public Health, 91 (2001), pp. 1194–1199.
- 327 [15] M.A. Borchardt, K.R. Bradbury, E.C. Alexander, R.J. Kolberg, S.C. Alexander,
 328 J.R. Archer, L.A. Braatz, B.M. Forest, J.A. Green, S.K. Spencer, Norovirus
 329 outbreak caused by a new septic system in a dolomite aquifer. Ground Water 49
 330 (2011) 85-97.
- [16] Y.C. Soh, F. Roddick and J. van Leeuwen, The future of water in Australia: the
 potential effects of climate change and ozone depletion on Australian water
 quality, quantity and treatability. Environmentalist, 28 (2008) 158–65.
- [17] L. Maunula, P. Klemola, A. Kauppinen, K. Soderberg, T. Ngujen, T. Pitkänen,
 S. Kaijalainen, M.L. Simonen, I.T. Miettinen, M. Lappalainen, J. Laine, R.
 Vuento, M. Kuusi and M. Roivainen, Enteric viruses in a large waterborne
 outbreak of acute gastroenteritis in Finland. Food Environ. Virol. 1(2009) 31-36.
- [18] FAO. Water Quality Management and Pollution Control in the Near East: An
 Overview. Regional Workshop on Water Quality Management and Pollution
 Control in the Near East. Cairo, Egypt, 2000.
- [19] M.N. Allam and G.I. Allam, (Water resources in Egypt: future challenges and opportunities. Int Water Resour Assoc Water Int 32(2)2007,205–218.
- 343 [20] Global Water Intelligence (November 2012). "Wstewater focus moves out of the
 344 city".
- 345 [21] World Bank (2008), Arab Republic of Egypt: Urban Sector Update, Retrieved
 346 on 2009-12-15
- [22] A. Kauppinen, K. Martikainen, V. Matikka, A-M. Veijalainen, T. Pitkänen, H.
 Heinonen-Tanski, I. T. Miettinen, (2014). Sand filters for removal of microbes
 and nutrients from wastewater during a one-year pilot study in a cold temperate
 climate. Journal of Environmental Management 133 (2014) 206-213.
- [23] American Public Health Association, APHA. 1998. Standard method for the
 examination of Water and wastewaters 20th ed., New York.
- 353 [24] American Public Health Association, APHA, Standard Methods for the
 354 Examination of Water and Wastewater 21th ed. Washington, D.C. 2005.
- [25] P. Rusin and C. Gerba, Association of chlorination and UV irradiation to
 increasing antibiotic resistance in bacteria, Rev. Environ. Cont. Toxicol., 171
 (2001) 1-52.
- [26] L.A. Marcos, P. Yi, A. Machicado, R. Andrade, S. Samalvides, J. Sánchez et al.
 Hepatic fibrosis and Fasciola hepatica infection in cattle. J Helminthol, 81 (2007),
 pp. 381–386.
- [27] S. George, V. Raju, M.R.V. Krishnan, T.V. Subramanian, K. Jayaraman,
 Production of protease by Bacillus amyloliquefaciensin solid-state fermentation

- 363 [and its application in the unhairing of hides and skins. Process Biochem, 30,364 (1995) 457–462.
- 365 [28] A. Pednekar, S. Grant, Y. Jeong, Y. Poon, C. Oancea, Influence of climate change, tidal mixing, and watershed urbanization on historical water quality in
 367 Newport Bay, a saltwater wetland and tidal embayment in southern California.
 368 Environ. Sci. Technol. 39 (23), (2005), 9071–9082.
- [29] A. Maimon, A. Tal, F. Friedler and A. Gross, Safe on-site reuse of greywater for
 irrigation A critical review of current guidelines. Environmental Science and
 Technology, 44(9), 2101, 3213–3220.
- [30] C. Levantesi, R. La Mantia, C. Masciopinto, U. Böckelmann, M. N. AyusoGabella, M. Salgot, V. Tandoi, E. Van Houtte, T. Wintgens and E. Grohmann,
 Quantification of pathogenic microorganisms and microbial indicators in three
 wastewater reclamation and managed aquifer recharge facilities in Europe.
 Science of the Total Environment 408 (2010) 4923–4930.
- [31] M. Ritzkowski and R. Stegmann, Emission behaviour of aerated landfills: results
 of laboratory scale investigations. In: Proceedings of Sardinia 2003 Ninth Waste
 Management and Landfill Symposium, Cagliari, Italy (2003).

381 Caption of Figures and Tables

- Fig. 1 Microbiological changes (total bacterial count (TBC), total yeasts and moulds
 count (TYC) and *Candida* counts (TCC)) of wastewater in Sewage treatment plant
 during one year 2012-2013.
- 385 Fig. 2 Microbiological changes (total coliform count (TCFC), total E. coli (TEC) and
- total *Salmonella* and *Shigella* counts (TSS)) of wastewater in Sewage treatment plant
 during one year 2012-2013.
- **Fig.3**: Total Suspended Solids (TSS, mg/l), Biochemical Oxygen Demand (BOD),
- and Chemical Oxygen Demand (COD) of Untreated Wastewater (UW) and Treated
- Wastewater (TW) in Sewage Water Treatment Plants (SWTPs) during one year 2012 2013.
- **Table 1:** Removal efficiency (RE) of total bacterial count (TBC), total yeasts & moulds count (TYMC), total Candida count (TCC), total coliform count (TCFC),
- *Escherichia coli* (EC), *Salmonella* and *Shigella* (SS) count) in Sewage Water Treatment Plant during 2012-2013.
- **Table 2:** Correlations between microbial groups (total bacterial count (TBC), total
- 397 yeasts & moulds count (TYMC), total *Candida* count (TCC), total coliform count
- 398 (TCFC), Escherichia coli (EC), Salmonella and Shigella (SS) count) in Sewage Water
- 399 Treatment Plant during 2012-2013
- 400
- 401 **Table 3:** Temperature, pH, Total Dissolved Solids (TDS, mg/l), SO⁻⁴, NO⁻³ and oil
- 402 levels of Untreated Wastewater (UW) and Treated Wastewater (TW) in Sewage
- 403 Water Treatment Plants (SWTPs) during one year 2012-2013
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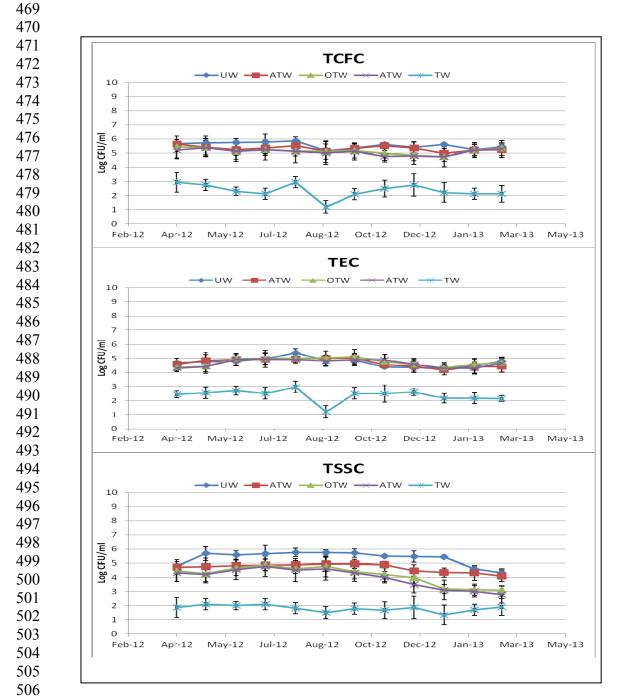


Fig. 2 Microbiological changes (total coliform count (TCFC), total *E. coli* (TEC) and
total *Salmonella* and *Shigella* counts (TSS)) of wastewater in Sewage treatment plant
during one year 2012-2013.

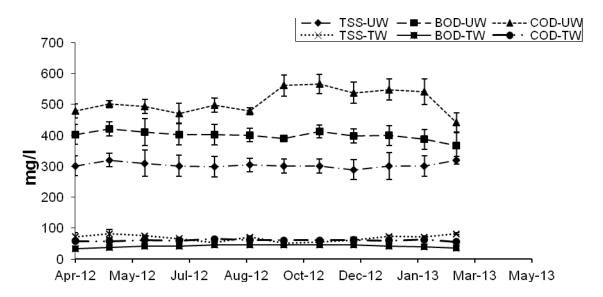




Fig.3 : Total Suspended Solids (TSS), Biochemical Oxygen Demand (BOD), and

516 Chemical Oxygen Demand (COD) of Untreated Wastewater (UW) and Treated

517 Wastewater (TW) in Sewage Water Treatment Plants (SWTPs) during one year 2012-518 2013.

Table 1: Removal efficiency (RE) of total bacterial count (TBC), total yeasts & moulds
count (TYC), total *Candida* count (TCC), total coliform count (TCFC), *Escherichia coli*(EC), *Salmonella* and *Shigella* (SS) count) of wastewater in Sewage Water Treatment Plant
during 2012-2013.

Time (month)	RE-TBC	RE-TYC	RE-CFC	RE-EC	RE-SS	RE-TCC
Apr-12	58.33	52.46	47.88	47.65	60.96	55.66
May-12	57.14	51.27	51.75	46.11	63.09	55.86
Jun-12	58.43	55.19	59.83	43.31	64.09	56.80
Jul-12	59.55	57.66	63.26	49.40	62.79	53.70
Aug-12	60.92	56.52	49.66	44.96	68.75	75.61
Sep-12	59.30	57.97	76.70	75.21	73.91	57.88
Oct-12	58.33	59.14	60.63	48.25	68.83	57.86
Nov-12	56.76	57.33	55.26	43.05	69.58	62.31
Dec-12	56.34	59.90	48.98	40.27	65.81	60.50
Jan-13	56.94	47.52	60.43	49.30	75.37	57.88
Feb-13	54.93	59.89	59.46	51.12	63.20	62.19
Mar-13	60.00	57.17	60.59	54.72	55.92	57.88
Total mean of log						
removals CFU/ml	4.71	2.87	3.20	2.33	3.55	1.97
Total mean of						
*RE %	58.08	56.00	57.87	49.44	66.03	59.51

524	*RE, Removal	Efficiency = [(influent-effluer	nt)/influentx100
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Table 2: Correlations between microbial groups (total bacterial count (TBC), total yeasts &

532 moulds count (TYMC), total *Candida* count (TCC), total coliform count (TCFC), *Escherichia*

533 coli (EC), Salmonella and Shigella (SS) count) of wastewater in Sewage Water Treatment

534 Plant during 2012-2013

	TBC	TYMC	TCFC	EC	TSSC	TCC
TBC		0.020189*	0.188167*	0.283812*	-0.10849	0.19645*
TYMC			0.188741*	0.110162*	-0.19269	0.225603*
TCFC				0.833005*	0.341162*	-0.32795
EC					0.252595*	-0.19008
TSSC						0.243344*
*P< 0.05						

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537 **Table 3:** Temperature, pH, Total Dissolved Solids (TDS, mg/l), SO-4, NO-3 and oil

538 levels of Untreated Wastewater (UW) and Treated Wastewater (TW) in Sewage

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	Untreated Wastewater (UW)						Treated Wastewater (TW)					
Time (Month)	Temp UW	pH- UW	TDS-UW	SO4 UW	NO3- UW	Oil-UW	Temp TW	pH- TW	TDS-TW	SO4TW	NO3-TW	Oil-TW
Apr-12	23±3.1	7.6±0.08	1179±31.2	71±3.9	10±1.5	66±6.1	23±3.5	7.4±0.14	1108±37.2	6.7±0.41	7.9±0.77	7.3±0.31
May-12	26±3.2	7.6±0.08	1190±32.3	75±3.8	11±1.9	62±6.4	26±3.5	7.5±0.13	1087±33.1	6.7±0.31	8.6±0.72	7.2±0.36
Jun-12	28±3.3	7.7±0.07	1166±34.1	66±3.7	11±1.8	69±6.5	28±3.3	7.5±0.12	1149±43.1	6.7±0.41	8.7±0.73	7.4±0.35
Jul-12	29±3.5	7.7±0.07	1156±23.9	67±3.9	12±1.3	70±6.4	27±3.2	7.4±0.13	1148±34.3	6.8±0.41	9.5±0.76	7.4±0.37
Aug-12	24±3.2	7.8±0.07	1194±23.7	70±3.9	12±1.5	55±6.3	24±3.3	7.5±0.14	1193±43.3	7.8±0.43	9.6±0.75	7.3±0.38
Sep-12	27±3.2	7.8±0.09	1197±31.5	77±3.8	13±1.4	75±6.3	27±3.2	7.4±0.12	1190±34.3	7.3±0.32	10.3±0.71	7.4±0.38
Oct-12	27±3.6	7.9±0.09	1107±34.1	75±3.7	12±1.6	70±6.1	27±3.2	7.6±0.15	1118±32.1	7.8±0.33	10.1±0.74	7.7±0.37
Nov-12	22±3.4	7.7±0.07	1190±23.7	75±3.8	11±1.7	74±6.7	22±3.2	7.4±0.14	1144±34.3	7.8±0.31	8.4±0.77	7.8±0.35
Dec-12	18±3.2	7.7±0.08	1106±32.9	75±3.8	10±1.7	74±6.1	18±3.6	7.5±0.18	1143±34.1	7.6±0.34	7.8±0.74	8.1±0.33
Jan-13	19±3.4	7.7±0.08	1190±22.8	75±3.8	13±1.4	77±6.4	17±3.6	7.6±0.19	1162±32.1	7.7±0.45	10.2±0.75	7.5±0.32
Feb-13	19±3.3	7.8±0.09	1184±34.6	75±3.9	13±1.4	70±6.3	19±3.7	7.4±0.11	1061±33.4	7.7±0.43	9.5±0.77	6.6±0.34
Mar-13	24±3.2	7.7±0.06	1177±32.7	70±3.9	12±1.3	71±6.2	24±3.1	7.4±0.21	1146±34.2	7.8±0.31	9.8±0.72	7.2±0.34
Average	23.67	7.73	1169.67	72.58	11.67	69.42	23.67	7.49	1137.42	7.37	9.19	7.41
*RE								3.02	2.76	89.85	2.00	89.33

541 *RE, Removal Efficiency = [(influent-effluent)/influentx100]