

BIOLOGICAL POST HYDROLYSIS OF DIGESTATE ENHANCES THE BIOGAS PRODUCTION IN ANAEROBIC DIGESTION OF AGRO-WASTE

David Bolzonella, Federico Battista*, Andrea Mattioli, Cristina Nicolato, Silvia Lampis

Department of Biotechnology, University of Verona, Strada Le Grazie 15, 37134 Verona, Italy

*Corresponding author email: federico.battista@univr.it

ABSTRACT

This work simulates a post-thermal treatment to the recirculated fraction of digestate, still rich in recalcitrant organic materials. By this way, the digestate can be further exploited for methane production, increasing the substrates exploitation's grade. The influence of the digestate post treatment has been studied by the conduction of a thermal treatment on a mesophilic digestate, considered exhausted. Three different duration of the thermal treatment have been considered. Then the digestate has been processed once at mesophilic condition AD. It was found that HRT of 3 days of the thermal treatment assures an increasing of the 25% and 65% of the soluble organic matter and VFA concentrations, respectively. As consequence biogas production increased of the 30% respect to the case without thermal post treatment.

Keywords: Anaerobic digestion, biogas, thermophilic, post-hydrolysis, thermal treatment

1. Introduction

The worldwide production of solid waste has reached the alarming level of 17 billion tons and it has been estimated it will be around of 27 billion tons within 2050 as consequence of the human population growth, especially in Africa and Asia continents [1]. The increasing awareness of the problem is leading the Western countries' governments to legislate in favor of more sustainable waste management based on the circular economy concept [2, 3]. Circular economy considers waste as raw materials for the production of new environmental friendly and biofuels or for the recovery of high added value molecules and nutrients. Consequentially waste disposal on soil or landfill must be avoided.

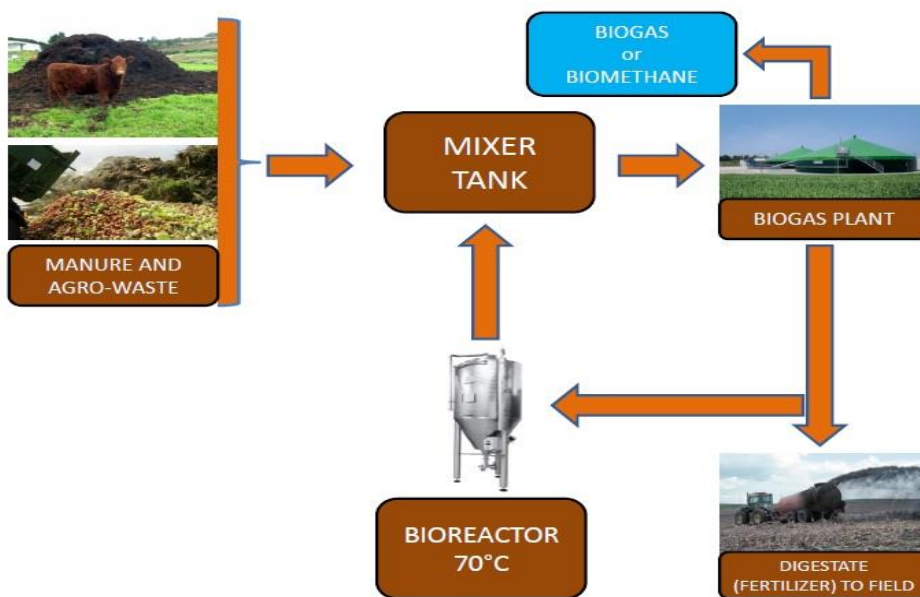
Organic residual from farm and agricultural activities are among the most abundant wastes in Europe with more than 1,500 million tons/year and 250 million tons/year, respectively [4, 5]. Circular economy model encouraged the exploitation of these wastes streams close to farms where they are produced to reduce the energetic and economic costs for transport [6]. Anaerobic digestion (AD) is the most adopted technology for the management of great amount of agricultural wastes. Some of the benefits of this technology are the production of biogas rich in methane to be used for heat and electricity cogeneration or as automotive fuels. Moreover, AD reduces the odor, the amount of the organic wastes stabilizing the wastes, contributes to the reduction of the Greenhouse Gases (GHG) emissions and allows the production of a high value fertilizer to be commercialized or applied directly in the farm [7]. One of the major disadvantages of AD from agricultural wastes consist in the recalcitrant nature (essentially lignin and crystalline cellulose) of these substrates. When feeding as manure or straw to anaerobic digesters, the inherent resistance of these lignocellulosic fraction in the raw material limit the convertibility of the materials. As a result only between 40% and 50% of the feed stock will be converted to biogas and the rest leaves the reactor unused, reducing the profitability of the AD system [8]. Consequentially, the output digestate is still rich in not stabilized organic matter and can be not applied as soil improvers [9].

The pretreatment stage becomes fundamental to increase the lignocellulosic materials degradation, to increase the porosity and consequentially the specific surface available to microorganisms involved in the AD. Pretreatments allow also the optimization of the biogas production and the stabilization of the organic substrates at the end of the AD. The combination of physical and chemical pretreatments are the most adopted for substrates with high content of lignin and cellulose. The wet explosion (WEx) pretreatment showed a high efficiency in the solubilization of cellulose in hexose sugars favoring their following conversion in methane by AD [8]. WEx contemplates the use of oxygen as oxidizing agent to destroy the cellulose polymers in glucose and to degrade the lignin in lower molecular weight aliphatic acids and phenols. Steam explosion is an alternative pretreatment where cellulose and lignin solubilization is performed by the substrates soaking in an acid solution, usually sulfuric acid, at high temperature (150-200°C) and high pressure (10-15 bar) for some minutes (5-15 minutes) [10]. Good performances have been also obtained by the steam explosion in an alkali solution (NaOH), which was able to hydrolyze 69 and 38% of the cellulose and hemicellulose, respectively, and favour the partial solubilization of the 75% of the lignin [11]. An alternative approach is the ammonia fiber expansion (AFEX) pretreatment. Biomass enters in contact with concentrated ammonia at temperatures of 70-180 °C and pressure ranges between 200 and 1000 psi. After a brief residence time, the pressure is explosively released, effectively disrupting the structure of the biomass. AFEX decrystallizes cellulose, partially hydrolyzes hemicellulose, and depolymerizes lignin [12], inducing a better methane production.

All these pretreatments are characterized by high use of reagents and energy. Thermal pretreatments can be used as valid processes to avoid the use of reagents resulting is particularly well suited to locations where there is a supply of waste

54 heat, for example from a nearby factory or power plant [13]. The thermal treatments require heating the substrates between
 55 130-200°C for about 30 minutes and under high pressure. This action results in a better degradation of the proteins and
 56 of the fibers, and in general improves the volatile solids removal by 20% of the conventional AD process. In addition,
 57 thermal treatment contributes to reduce the reaction medium viscosity improving the heat and mass transfer [14]. By this
 58 way, thermal pretreatments is able to impact the AD kinetics reducing the hydraulic retention time (HRT) and the methane
 59 production [15]. Recently a new process has been tested at laboratory/pilot scale, the Intermediate Thermal Hydrolysis
 60 Process (ITHP) in which thermal hydrolysis is not a pretreatment but an intermediate process [16]. In this process, the
 61 substrates are treated by AD before the ITHP. The resulting digestate from ITHP has still adopted for a final stage of AD
 62 step. This process improved organic matter removal by 15% as compared to the conventional thermal treatment with a
 63 consequent methane increasing of about 10%. It was demonstrated that post-treatment of digestate improved volumetric
 64 methane yields by 7% and the COD-reduction increased from 68% to 74% in a mesophilic (37 °C) semi-continuous
 65 system [17]. Moreover, the HRT was compared to a conventional system with pre-treatment of feed substrates at 70 °C
 66 [18].

67 This research has the aim to simulate a 70°C thermal treatment to the recirculated fraction of digestate, derived from a
 68 full-scale digester, treating manure and straw. After the thermal post-treatment, the digestate is mixed with new fresh
 69 organic matter for a mesophilic AD process (Figure 1). By this way, the digestate, still rich in recalcitrant lignocellulosic
 70 materials, can be further hydrolyzed for methane production, increasing the substrates exploitation's grade. The influence
 71 of the digestate post treatment has been studied by the conduction of a mild thermal treatment on a mesophilic digestate,
 72 considered exhausted. Three different duration of the thermal treatment have been considered. Then the digestate has
 73 been processed once at mesophilic condition AD. The influence of thermal treatment on the microbial community
 74 composition has been also studied through PCR-DGGE analysis carried out on the bioreactor under thermophilic
 75 conditions.



76
 77 Figure 1. Concept of the thermal treatment to the recirculation fraction of digestate.

78 **2. Materials and Methods**

79 The research activity is composed by two experimental parts: i) the Biochemical Methane Potential (BMP) tests conducted
 80 in batch and ii) the semi-continuous tests.

81 **2.1 Characterization of the digestates**

82 The two experimental parts were conducted adopted two digestates: the first digestate is represented by a mesophilic
 83 digestate from a full scale reactor located at Isola della Scala (Italy) working at 37°C and used to simulate a typical
 84 mesophilic AD process. The second is constituted by a thermophilic digestate from a full scale reactor located at Treviso
 85 working at 65°C. Table 1 shows the chemical characterization of the both the digestates.

	Mesophilic Inoculum	Thermophilic inoculum
TS (% w/w)	67.18 ± 0.52	62.40 ± 1.27
TVS (% w/w)	43.48 ± 0.37	44.40 ± 0.61

pH	8.36 ± 0.02	5.30 ± 0.01
Alkalinity (g CaCO ₃ /L)	1.49 ± 0.26	2.94 ± 0.17
COD (mg O ₂ /gTS)	758.93 ± 2.60	706.56 ± 5.54
sCOD (mg O ₂ /L)	7,110.39 ± 34.91	14,131.00 ± 42.31
VFA (mg O ₂ /L)	675.48 ± 4.73	6,889.00 ± 10.45
TKN (mg/gTS)	21.91 ± 1.67	33.49 ± 1.57
NH ₃ -N (mg N/L)	3,190.00 ± 19.23	662.00 ± 12.48
NH ₃ (mg N/L)	367.39 ± 4.95	0.96 ± 0.04
P (mg/gTS)	16.41 ± 0.91	14.32 ± 0.60

86 Table 1. Chemical characterization of the digestates used for the tests

87 2.2 BMP tests

88 BMP tests have been conducted to determine the methane production from two agricultural digestates, representing the
89 output of poultry and bovine manure and rice straw, treated by two different anaerobic full scale digesters in Italy. In
90 particular, the first digestate is represented by a mesophilic digestate from a full scale reactor located at Isola della Scala
91 (Italy) working at 37°C and used to simulate a typical mesophilic AD process. The second is constituted by a thermophilic
92 digestate from a full scale reactor located at Treviso working at 65°C.

93 BMP tests were based on the procedure described in Angelidaki et al. [19] and Holliger et al. [20]. Two typologies of
94 BMP tests were performed: the first to simulate a mesophilic AD process to use as control test, the second to simulate the
95 thermophilic post treatment on a mesophilic digestate (PTMD). The difference in methane production has been used to
96 evaluate the performance of thermal treatment.

97 The tests were conducted at mesophilic temperature (37 ± 1°C). 1000 mL glass bottles were used as reactors. The control
98 tests bottle were filled with the mesophilic digestate till to reach the working volume of 500 mL. The PTMD tests were
99 performed according to an inoculum to substrate ratio of 2.5:1 on a VS-basis, where the term inoculum is referred to the
100 digestate from thermophilic AD process, while substrate is represented by the digestate from mesophilic process. Finally,
101 the bottles were flushed with N₂ to guarantee anaerobic conditions, closed and incubated in a temperature-controlled
102 chamber. A SMC pressure Switch manometer (1 bar, 5% accuracy) was used to measure biogas production in the
103 headspace of the bottles, up to the depletion of the biogas production. Accumulated volumetric biogas production was
104 calculated considering the pressure increase in the headspace volume. Periodically, gas samples were taken from the
105 reactors to analyse biogas composition by the portable biogas analyzer BIOGAS5000 (Geotech, United Kingdom). Assays
106 were carried out in triplicate, and the results were expressed as the biogas volume produced per gram of volatile solids
107 under standard ambient temperature conditions (L biogas or methane kg VS⁻¹ at T=298 K, P=1.0133 bar).
108 Kinetic parameters were calculated using the modified Gompertz model described in Eq. 1 [21].

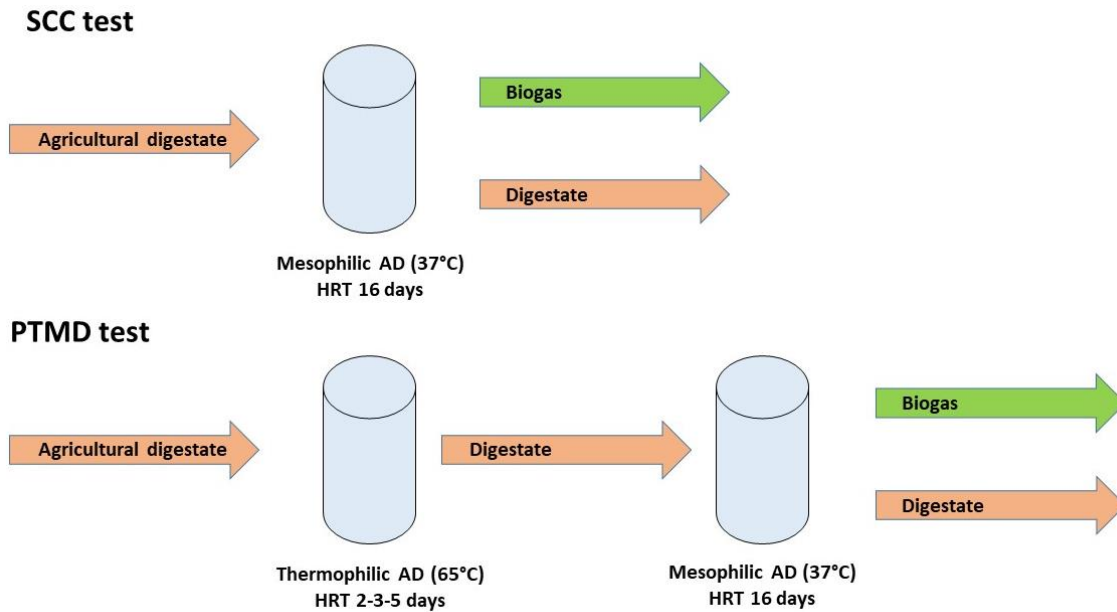
$$109 P_{\text{net}}(t) = P_{\text{max}} \exp \left\{ -\exp \left[\frac{R_{\text{max}} \cdot e}{P_{\text{max}}} (\lambda - 1) + 1 \right] \right\} \quad \text{Eq. 1}$$

110 Where $P_{\text{net}}(t)$ is the net accumulated methane production (NL_{CH₄}/kg_{TVS}) at time t, P_{max} is the methane potential production
111 (NL_{CH₄}/kg_{TVS}), R_{max} is the maximum methane production rate (NL_{CH₄}/kg_{TVS} d), and λ is the lag phase (d).

112 2.3 Semi-continuous tests

113 The performances of the PTMD were evaluated in semi-continuous mode too and compared with a semi-continuous
114 control test (SCC). Figure 2 summarizes the SCC and PTMD processes. In the PTMD test the mesophilic digestate was
115 digested at 70°C and then digested for a second time at mesophilic temperature (37°C). On the contrary SCC was simply
116 digested a second time at 37°C.

117



118

119 Figure 2. Schemes of SCC and PTMD tests.

120 SCC and PTMD tests were conducted at lab scale reactors, with a total volume and a working volume of 4.5 L and 4.0 L,
 121 respectively. The PTMD experimental campaign included a 14 days start-up phase where the digestates have been
 122 gradually added until to reach the working volume and three phases corresponding to three different Hydraulic Retention
 123 Time (HRT) of the thermophilic reactor set up at 2.0; 3.0 and 5.0 days. The HRT of the mesophilic reactor was keep
 124 constant at 16 day; the duration of the mesophilic AD was established at 56 days corresponding to 3.5 HRT to assure the
 125 reaching of steady state methane production. Every day SCC and PTMD were fed according the Organic Load Rate
 126 (OLR) reported in Table 2. It is possible to observe that the OLR depends on the HRT of the reactors. Consequentially,
 127 SCC reactor and mesophilic phase of PTMD process were characterized by constant OLR along the duration of the semi-
 128 continuous tests, while the thermophilic phase of PTMD test' OLR decreased with the HRT augmentation. The working
 129 volumes of the reactors were kept constant by the daily discharged of the same digestate amounts.

Reactor	OLR of the different phase of PTMD test (kg TVS / m3 d)		
	HRT (2 days)	HRT (3 days)	HRT (5 days)
SCC process	3.64 ± 0.24	3.25 ± 0.30	3.54 ± 0.11
Thermophilic reactor of PTMD	24.78 ± 0.84	17.33 ± 1.60	11.34 ± 0.34
Mesophilic reactor of PTMD	3.49 ± 0.28	3.20 ± 0.18	3.18 ± 0.14

130 Table 2. OLR values for the different HRT of the thermophilic reactor in PTMD process.

131 **2.4 Evaluation of the tests' performances and analytical methods**

132 BMP tests were evaluated simply considering the biogas production and the methane content, according the methods
 133 previously described.

134 The performances of the semi-continuous tests were evaluated considering the difference of: i) the methane productions
 135 between SCC test and the mesophilic reactor of PTMD tests; ii) soluble Chemical Oxygen Demand (sCOD) and Volatile
 136 Fatty Acids concentrations between SCC and PTMD tests. These latter parameters are considered as indicators of the
 137 hydrolysis efficacy of the thermal treatment.

138 Several parameters were measured using samples taken at the outlet stream from SCC and PTMD processes: pH,
 139 alkalinity, Volatile Fatty Acids (VFAs), total Chemical Oxygen Demand (COD), soluble COD (sCOD), Total Solids (TS),
 140 Total Volatile Solids (TVS), and the concentrations of total nitrogen compounds (TKN), of ammonium (NH₃-N) and free
 141 ammonia (NH₃) and phosphorus (P). They were determined using the standard methods described in the scientific
 142 literature [22]. In addition, the biogas produced was measured both quantitatively and qualitatively, using a TG1PP gas
 143 meter (RITTER, Germany) and a portable biogas analyzer, the BIOGAS5000 (Geotech, United Kingdom).

144 **2.5 Total DNA extraction, PCR amplification, and DGGE analysis**

145 DNA extractions and PCR setups were performed under microbiological safety cabinet (SafeFAST Elite class II, Carlo
 146 Erba). Total DNA was extracted using the FastDNA® SPIN Kit for soil (MP Biomedicals, USA) according to the
 147 manufacturer's instructions. Samples collected from the hyperthermophilic reactor were extracted in duplicate, starting
 148 from 0.5 g of biomass for each extraction. PCR comprised about 50 ng of template DNA, 0.8 μ M of each primer, 0.4 mM
 149 of the four dNTPs, 1 unit of GoTaq™ DNA Polymerase and 5 μ l of 5X PCR buffer. Eubacterial 16S rRNA genes were
 150 amplified through primers fd1 and rp2 [23]. The PCR reaction was performed with the following program: initial
 151 denaturation at 94 °C for 5 min; 30 cycles of denaturation at 94 °C for 30 s, annealing at 45 °C for 30 s, extension at 72
 152 °C for 2 min; and final extension at 72 °C for 5 min. The following nested PCR was performed on the hypervariable V3
 153 region using primers p2/p3 [24], with a GC-clamp. The nested PCR was as follows: initial denaturation at 94 °C for 5
 154 min; 35 cycles of denaturation at 94 °C for 30 s, annealing at 57 °C for 30 s, extension at 72 °C for 45 s; and final extension
 155 at 72 °C for 5 min. Archaeal 16S rRNA genes were amplified using primers A109-f [25] and 1510-r [26]. Afterward, a
 156 nested PCR was performed on the hypervariable V2-V3 region using primers A109(T)-f and 515-GCr with a GC clamp.
 157 The first archaeal PCR reaction was performed with the following thermocycle program: initial denaturation at 94 °C for
 158 5 min; 38 cycles of denaturation at 94 °C for 30 s, annealing at 52 °C for 30 s, extension at 72 °C for 1 min; and final
 159 extension at 72 °C for 5 min. The nested PCR was as follows: initial denaturation at 94 °C for 5 min; 35 cycles of
 160 denaturation at 94 °C for 30 s, annealing at 52 °C for 30 s, extension at 72 °C for 45 s; and final extension at 72 °C for 5
 161 min. All primers were purchased from Sigma-Genosys (Milan, Italy). The PCR products were quantified using Low DNA
 162 Mass™ Ladder (Euroclone, Italy) in a 2.0 % agarose gel. DGGE analyses were performed on amplicons obtained for
 163 V3 regions for Eubacteria and V2-V3 region for Archea as previously reported [27, 28]. The gel (8%
 164 acrylamide/bisacrylamide 19:1, BioRad) was cast using a denaturing gradient of 30–60%, with 100% denaturant defined
 165 as 7 M urea and 20% (v/v) formamide.

166 2.6 Cloning, sequencing, and taxonomical analysis

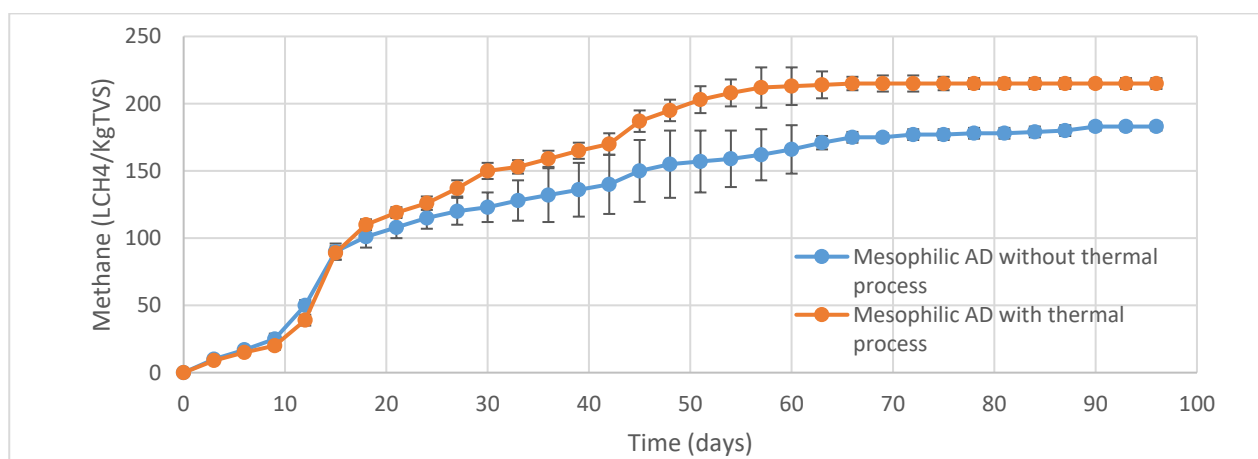
167 Major bands in the DGGE-gel were cut off. Twenty-five μ l of sterilized water was added to the excised DGGE bands and
 168 incubated on a rotary shaker (200 rpm) at 37°C for 4 h. Afterwards, 5 μ l eluate was used as DNA template for re-
 169 amplification. PCR amplification was carried out as described before, except for the use of non-GC-clamped primers.
 170 PCR products were transformed in *Escherichia coli* X11blue using the pGEM-T vector system according to the
 171 manufacturer's instructions (Promega, Italy), sequenced on both strands (GATC Biotech; Cologne, Germany), and finally
 172 searched for identity using the NCBI [29] and EzBioCloud [30] databases.

173 3. Results and discussions

174 3.1 BMP tests

175 As reported above, two BMP tests have been carried out to show the methane production's difference from a mesophilic
 176 AD process with and without a thermal process. By this way, the performances a thermal post treatment and the
 177 recirculation of the digestate from a mesophilic AD process, may be predicted.

178 Figure 3 shows the Gompertz curves of the BMP tests. The first 15 days of the tests have been characterized by a very
 179 similar trend of methane production: the two tests had almost the same exponential phase growth. The lag phases of both
 180 the BMP tests was inferior to 24 hours, demonstrating the agricultural digestate are still rich in edible organic matter.
 181 After the 20th day, the kinetics of the BMP tests decreased and methane production recorded smaller and smaller daily
 182 rates. The mesophilic AD with and without thermal pretreatment BMP tests reached the definitive methane production
 183 after about two and three months, respectively.



184
 185 Figure 3. The Gompertz curve of the BMP tests

186 The trend of the BMP tests, having similar exponential growth in the first 15 days, demonstrate that thermal treatment
187 seems to be efficacy on the degradation of more recalcitrant and slowly degradable organic compounds. The kinetic of
188 residual simple compounds (sugars and carbohydrates) was not affected by thermal treatment. Instead, it improved the
189 solubilization of more complex molecules, such as lipids and proteins [31], present in high concentration in digestate
190 from manure AD [9]. These compounds, and in particular proteins, are thermos-labile structures whose hydrogen and
191 covalent bonds can be definitively compromise by heat, for temperatures higher to 60-65°C [32]. Moreover, previous
192 studies on thermal treatment at 70°C before AD showed that heat is able to hydrolyze the lignocellulosic materials. In
193 particular, the reduction rates of cellulose were 15.7%, 23.5%, 24.9%, and 44.3% and those of hemicellulose were 1.8%,
194 13.2%, 19.0%, and 32.5% for a duration of the thermal process of 1 d, 2 d, 3 d, and 4 d, respectively [33]. This effect is
195 helpful in both alleviating the binding force and facilitating solubilization allowing a higher biogas production [34].

196 BMP tests confirmed the beneficial effect of thermal treatment on digestate in the increasing of the more complex organic
197 matter solubilization, with a consequent optimization of the methane production, which passed from 183 LCH₄/kg_{TVS} for
198 mesophilic AD without thermal treatment of digestate to 215 LCH₄/kg_{TVS} for the mesophilic AD with thermal treatment
199 of digestate. It corresponded to a methane increment of about 15%, which encouraged the conduction of new tests at
200 bigger scale and semi-continuous mode.

201 3.2 Semi continuous tests

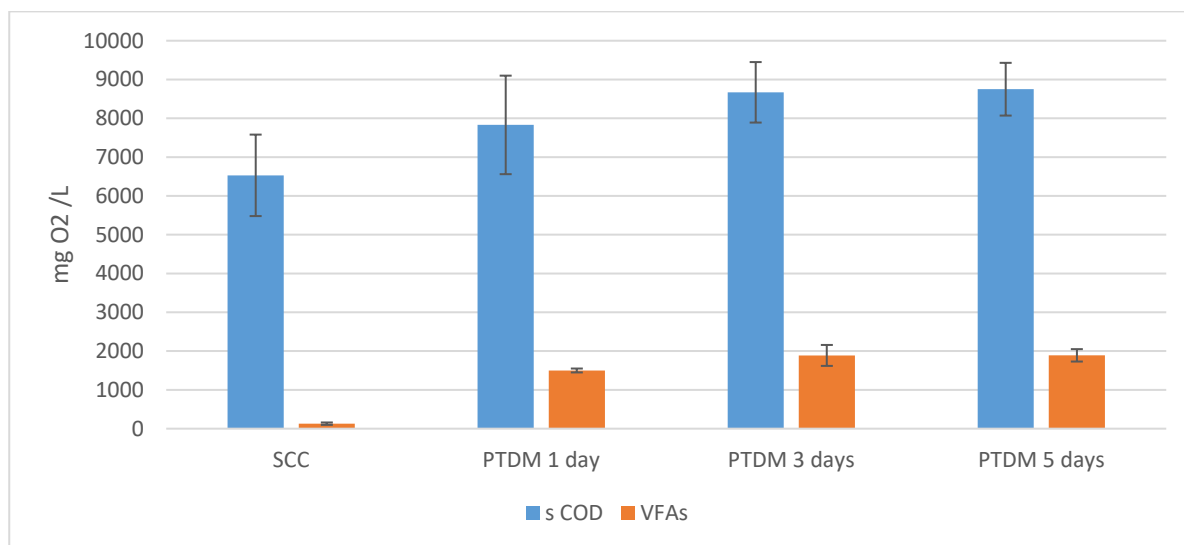
202 Table 3 summarizes the chemical parameters of the digestates of the semi-continuous SCC and of the PTDM at the three
203 different thermophilic stage's HRTs.

204 SCC test showed high residual concentration of TS and TVS, respectively at the 75.80 and 53.20% w/w. The COD
205 content, parameter usually used to measure the organic matter concentration, is almost of the 800 mg O₂/gTS, while the
206 sCOD is lower than all the other semi-continuous tests (6.53 gO₂/L), where the thermal process were applied. It
207 demonstrated that agricultural digestate is still rich in recalcitrant organic matter, not degradable by a mesophilic AD even
208 at long HRT [35]. Consequentially, digestate cannot be considered stabilized and adapted for fertilizer application. These
209 criticisms justify the need of a mild thermal treatment.

	SCC	HRT 2 days		HRT 3 days		HRT 5 days	
		Thermophilic reactor of PTDM	Mesophilic reactor of PTDM	Thermophilic reactor of PTDM	Mesophilic reactor of PTDM	Thermophilic reactor of PTDM	Mesophilic reactor of PTDM
TS (% w/w)	75.80 ± 2.06	67.54 ± 3.13	65.04 ± 3.05	62.80 ± 1.94	53.74 ± 1.64	62.62 ± 3.37	54.20 ± 2.52
TVS (% w/w)	53.20 ± 2.51	46.40 ± 2,58	44.62 ± 2.37	42,02 ± 1.93	35.17 ± 2.96	40.24 ± 2.25	36.59 ± 1.52
pH	8.08 ± 0.13	8.46 ± 0.15	8.23 ± 0.10	8.34 ± 0.08	8.20 ± 0.05	8.45 ± 0.15	8.22 ± 0.17
COD (mg O ₂ /gTS)	790.91 ± 29.34	847.84 ± 29.91	843.54 ± 30.36	850.38 ± 33.91	803.51 ± 28.40	795.85 ± 27.34	772.20 ± 44.70
sCOD (g O ₂ /L)	6.53 ± 1.05	7.83 ± 2.27	7.21 ± 1.98	8.67 ± 0.78	7.51 ± 0.26	8.75 ± 0.68	7.39 ± 0.81
VFA (mg O ₂ /L)	129.02 ± 30.92	1500 ± 50	409.05 ± 5.52	1888 ± 270	196.07 ± 31.20	1890 ± 160	129.20 ± 30.57
P (mg/gTS)	13.50 ± 1.10	13.20 ± 0.91	13.00 ± 0.43	13.66 ± 1.57	13.71 ± 0.89	13.68 ± 0.55	13.72 ± 0.08
TKN (mg/gTS)	18.04 ± 3.47	19.90 ± 1.20	23.98 ± 4.05	19.73 ± 2.11	19.38 ± 1.07	20.36 ± 3.20	20.06 ± 2.14

210 Table 3. Chemical characterization of the digestate at the end of the semi-continuous tests

211 The thermal treatment was tested with three different HRT of 2, 3 and 5 days. In all the cases, it was always efficacy in
 212 the degradation of the organic matter, as showed in Figure 4 where the thermal treatments' effects were illustrated in
 213 terms of sCOD and VFAs increasing.



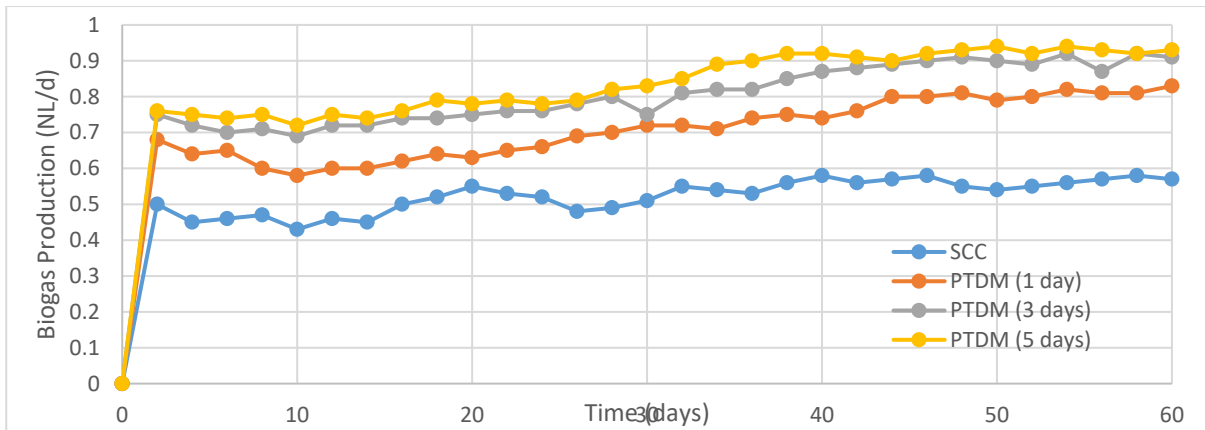
214
 215 Figure 4. sCOD and VFA concentration of the semi-continuous tests.

216 The concentration of sCOD passed from 653 mgO₂/L for the SCC test to 783, 867 and 875 mgO₂/L for the thermal
 217 treatment conducted with HRT at 2, 3 and 5 days, respectively. On the contrary, the TS concentrations decreased from
 218 75% w/w of the SCC to about 50-55% of the PTDM tests having the HRT at 3 and 5 days, respectively. Similarly, the
 219 TVS dropped from about 55% w/w of the SCC tests to about 35% w/w of the PTDM tests with HRT at 3 and 5 days,
 220 respectively (Table 3). These results confirmed the efficacy of the heat application in the degradation of thermos-labile
 221 organic structures, as proteins and -lipids, which start to lose their natural conformation for temperature higher than 55-
 222 60°C [36]. Thermal treatment has also recognized for its ability to solubilize long polymers chains, such as cellulose
 223 hemicellulose and partially of the lignin [37].

224 The parameter which was more affected by the thermal treatment has been the VFA concentration. PTMD tests showed
 225 very high VFA concentration after the thermophilic stage: 1,500 mgO₂/L and almost 1,900 mgO₂/L for the tests having
 226 HRT 2 and 3-5 days, respectively. The control SCC tests, where thermophilic treatment was not applied had lower VFAs
 227 concentrations of about 670 mgO₂/L and 129 mgO₂/L at the beginning and at the end of the mesophilic AD, respectively.
 228 It was demonstrated the improving of thermal application on VFA concentrations, testing different operative
 229 temperatures. They showed how mild temperature (60-100°C) contributed to the higher VFA production derived from
 230 the degradation of proteins, carbohydrates and cellulose. At these temperatures, the formation of inhibiting compounds,
 231 which occurs for higher values (150-200°C), can be avoided [37].

232 Another interesting consideration regards the optimal HRT for the mild thermal treatment. Figure 2 shows how the 1 day
 233 - HRT contributed to an important optimization of sCOD and VFA concentrations compared to the SCC test, with an
 234 increasing of more than the 15% and 55%, respectively. Anyway, longer HRT of 3 days brought to further sCOD and
 235 VFA productions' increasing at 25% and 65%, respectively. Instead, a HRT of 5 days did not lead to better performances.
 236 Consequentially, 3 days seemed to be the ideal HRT for thermal treatment conducted at mild temperatures. This result is
 237 coherent with previous studies, which observed the best VFA productions for thermal treatment having a duration of 3
 238 days [38, 39]. In particular, Zhang et al. [39] explained that if carbohydrates acidification needs of some hours, the
 239 solubilization of proteins and cellulose reach the maximum yield after 3 days. Proteins contribution in the VFA increasing
 240 is particularly important in agriculture digestate, containing high percentage of animal manure.

241 The methane production from the AD follows the trend of the sCOD and VFA concentrations. Figure 5 shows the daily
 242 biogas production for the SCC control test and for the PTDM tests at the different HRT values. As for the sCOD and
 243 VFA, thermal treatment allowed an optimization of biogas production from a daily production of 0.5 NL/d of the SCC
 244 test to about 0.9 NL/d of the PTDM tests with HRT of 3 and 5 days when the steady state was reached (Figure 5).



245

246

Figure 5. The daily biogas production for the semi-continuous tests.

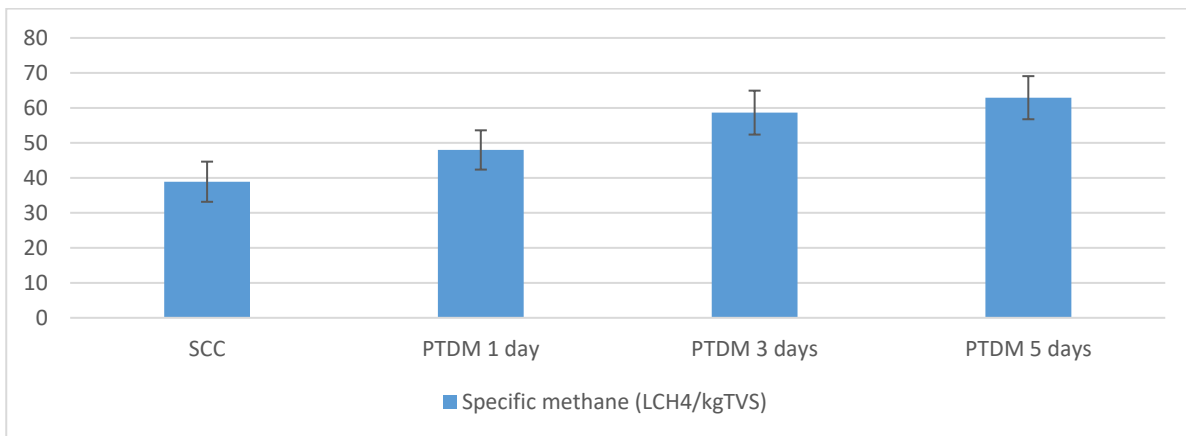
247

248

249

250

Lastly, Figure 6 shows the specific methane production for the semi-continuous tests: SCC reached a methane production of 38.90 LCH₄/kgTVS, while PTDM at 1, 3, 5 days HRT recorded a methane production of 47.98, 58.65 and 62.92 LCH₄/kgTVS, respectively. It means that a post-thermal treatment of digestate with a duration of 1, 3 and 5 days followed by a new mesophilic AD process, could allow to an increasing of methane production of about 20, 33 and 38%.



251

252

Figure 6. The specific methane production of the semi-continuous tests.

253

3.3 PCR-DGGE analysis on Eubacteria and Archea communities

254

255

256

The composition of eubacterial and archaeal communities enriched in the hyperthermophilic reactor was analyzed by means of PCR-DGGE molecular technique. DGGE patterns showed a well defined bacterial community for either eubacterial and archeal populations (Fig. 7).

257

258

259

260

261

262

263

264

265

266

267

268

269

270

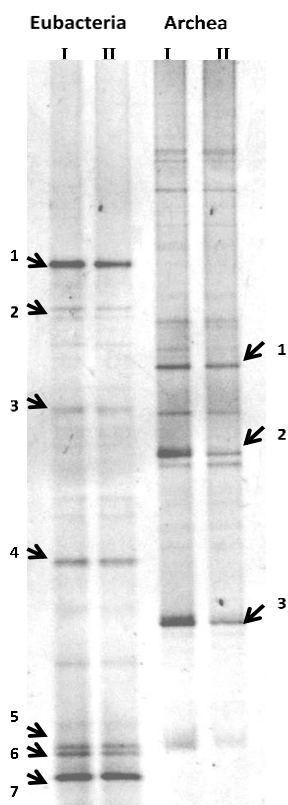
Major bands in DGGE gels were excised, cloned, and sequenced (Tab. 4). Sequencing analyses revealed that the main phyla for Eubacteria were Firmicutes, Bacteroidetes, Actinobacteria and Thermotogae. This findings is in accordance with previous studies which evidenced Firmicutes and Thermotogae as the major phyla in thermophilic digesters [40]. Moreover, members of the Bacteroidales, Clostridiales, and Thermotogales orders were reported to dominate the microbial community of biogas reactor once temperature increased from 55 to 75 °C [41]. In fact, it is worth noting that sequences we retrieved from our reactor have been already described as phylotypes in microbial communities of thermophilic reactor for anaerobic digestion of agricultural wastes. Among Firmicutes phylotypes, we found both Clostridiales and Hydrogenispora families, whereas among Bacteroidetes clones we identified members of Rikenellaceae and Porphyrimonadaceae family. Of particular interest is the band 5 in the Eubacteria DGGE pattern which corresponds to *Thermotoga neapolitana* species. *Thermotoga* is a genus of (hyper)thermophilic bacteria of the phylum *Thermotogae* [42], with optimum growth temperatures up to 80 °C. Members of the genus *Thermotoga* are anaerobic, rod-shaped bacteria encapsulated by a unique ‘toga’- like outer membrane. Members of this genus are able to use wide range of carbon sources (hexoses, pentoses, disaccharides, glucans, xylans, glucomannan, galactomannan, pectin, chitin and amorphouse found *Methanosarcina* cellulose).

271

272

273

Regarding Archeal community, we found members of *Methanosarcina*, *Methanobacterium* and *Methanothermobacter* genera (Tab. 4). Also in the case of Archea, these results are in accordance with previous studies which showed an increase of *Methanosarcina* and *Methanothermobacter* genera under thermophilic conditions in anaerobic digesters [41].



274

275 Figure 7. DGGE profiles of eubacterial and archeal communities in the iperthermophilic reactor. Arrows and letters in
 276 the gel indicate bands that have been excised, cloned, and sequenced.

277

Band	Taxon	Accession number	Percentage of identity (%)	Phylogenetic group
EU-1	Uncultured Bacteroidales	FN436068	99,47	Bacteroidales f
EU-2	Uncultured Hydrogenispora	DQ887962	100	Firmicutes, Hydrogenispora f
EU-3	Uncultured Porphyromonadaceae	FN436026	98,94	Bacteroidetes, Porphyromonadaceae f
EU-4	Uncultured <i>Catonella</i>	DQ394697	98,35	Firmicutes, Clostridiales
EU-5	<i>Thermotoga neapolitana</i>	DSM 4359	98,96	Thermotogaceae;
EU-6	<i>Cellulosilyticum lentocellum</i>	DSM 5427	97,04	Firmicutes, Clostridiales
EU-7	<i>Bifidobacterium pseudolongum</i>	DSM 20092	100	Actinobacteria, Bifidobacteriaceae
ARC-1	<i>Methanosarcina flavescens</i>	E03.2	99,43	Methanosarcinales
ARC-2	<i>Methanobacterium kanagiense</i>	169	97,42	Methanobacteriales
ARC-3	<i>Methanothermobacter wolfeii</i>	DSM 2970	99,72	Methanobacteriales

278

279 Table 4. Taxonomic characterization of the major bands cloned and sequenced from the DGGE profiles

280 **Conclusions**

281 The efficacy of the post treatment on an agricultural digestate was demonstrated. In particular, a considered exhausted
282 digestate is still rich in recalcitrant organic matter, potentially exploitable for methane production after an adequate
283 hydrolysis step. A thermophilic treatment conducted at 70°C with a HRT of 3 days assured an increasing of the soluble
284 organic matter and VFA concentrations of 25% and 65%, respectively. As consequence, the following AD, conducted in
285 mesophilic condition and in semi-continues mode, resulted in a specific methane production of about 60-65 LCH₄/Kgvs,
286 corresponding to a 30% more than the case without thermal treatment.

287 **References**

288 [1] Laurent, A., Bakas, I., Clavreul, J., Bernstad, A., Niero, M., Gentil, E., Hauschild, M.Z., Christensen, T.H.: Review
289 of LCA studies of solid waste management systems - Part I: Lessons learned and perspectives. *Waste Management* 34
290 (3), 573-588 (2014).

291 [2] US EPA: Sustainable Materials Management: Non-Hazardous Materials and Waste Management Hierarchy (2016).
292 Available online: [https://www.epa.gov/smm/sustainable-materials-management-non-hazardous-materials-and-waste-](https://www.epa.gov/smm/sustainable-materials-management-non-hazardous-materials-and-waste-management-hierarchy)
293 [management-hierarchy](https://www.epa.gov/smm/sustainable-materials-management-non-hazardous-materials-and-waste-management-hierarchy).

294 [3] European Commission: Waste directive 2008/98/EC. *Off. J. Eur. Union* 312, 3-30 (2008).

295 [4] Scarlat, N., Martinov, M., Dallemand, J.F.: Assessment of the availability of agricultural crop residues in the European
296 Union: potential and limitations for bioenergy use. *Waste Manage* 30 (10), 1889 – 1897 (2010).

297 [5] Nielsen, J.B.H., Oleskowicz-Popiel, P.: The future of biogas in Europe: Visions and targets until 2020. *The Future of*
298 *Biogas in Europe III* (2017).
299 [http://213.229.136.11/bases/ainia_probiogas.nsf/0/271DB55383A6DD0AC12575C400595F99/\\$FILE/Jens_Bo_Holm-](http://213.229.136.11/bases/ainia_probiogas.nsf/0/271DB55383A6DD0AC12575C400595F99/$FILE/Jens_Bo_Holm-NielsenPROBIOGASperspectivas%20ue.pdf)
300 [NielsenPROBIOGASperspectivas%20ue.pdf](http://213.229.136.11/bases/ainia_probiogas.nsf/0/271DB55383A6DD0AC12575C400595F99/$FILE/Jens_Bo_Holm-NielsenPROBIOGASperspectivas%20ue.pdf)

301 [6] Djuric Ilic, D., Eriksson, O., Ödlund L., Åberg, M.: No zero burden assumption in a circular economy. *Journal of*
302 *Cleaner Production* 182, 352-362 (2018).

303 [7] Mirtsou-Xanthopoulou, C., Jurado, E., Skiadas, I.V., Gavala, H.N.: Effect of aqueous ammonia soaking on the
304 methane yield and composition of digested manure fibers applying different ammonia concentrations and treatment
305 durations. *Energies* 7, 4157-4168 (2014).

306 [8] Ahring B.K., Biswas, R., Ahamed, A., Teller, P.J., Uellendahl, H.: Making lignin accessible for anaerobic digestion
307 by wet-explosion pretreatment. *Bioresource Technology* 175, 182–188 (2015).

308 [9] Battista, F., Bolzonella D.: Exploitation of Solar Energy for Ammonium Sulfate Recovery from Anaerobic Digestate
309 of Different Origin. *Waste and Biomass Valorization* (2019). doi:10.1007%2Fs12649-019-00597-x.

310 [10] Battista, F., Gomez Almendros, M., Rousset, R., Bouillon, P.A. : Enzymatic hydrolysis at high lignocellulosic
311 content: Optimization of the mixing system geometry and of a fed-batch strategy to increase glucose concentration.
312 *Renewable energy* 131, 152-158 (2019).

313 [11] Murnen, H.K., Balan, V., Chundawat, S.P.S., Bals, B., da Costa Sousa, L., Dale, B.E.: Optimization of Ammonia
314 Fiber Expansion (AFEX) Pretreatment and Enzymatic Hydrolysis of *Miscanthus x giganteus* to Fermentable Sugars.
315 *Biotechnol. Prog.* 23, 846-850 (2007).

316 [12] Mosier, N., Wyman, C., Dale, B., Elander, R., Lee, Y., Holtzapple, M., Ladisch, M.: Features of promising
317 technologies for pretreatment of lignocellulosic biomass. *Bioresour. Technol.* 96, 673-686 (2005).

318 [13] Montgomery, L.F.R., Bochmann, G.: Pretreatment of feedstock for enhanced biogas production. *IEA Bioenergy*
319 (2014). ISBN 978-1-910154-05-2.

320 [14] Barber, W.P.F.: Thermal hydrolysis for sewage treatment: A critical review, *Water Research* 104, 53-57 (2016).

321 [15] Li, Y., Jin, Y., Li, J., Li, H., Yu, Z., Nie, Y.: Effects of thermal pretreatment on degradation kinetics of organics
322 during kitchen waste anaerobic digestion, *Energy* 118, 377-386 (2017).

323 [16] Shana, A., Ouki, S.K. Asaadi, M. Pearce, P.: The impact of intermediate thermal hydrolysis on the degradation
324 kinetics of carbohydrates in sewage sludge, *Biores. Tech.* 137, 239–244 (2013).

- 325 [17] Svensson, K., Kjølraug, O., Higgins, M.J., Linjordet, R., Horn, S.J.: Post-anaerobic digestion thermal hydrolysis of
 326 sewage sludge and food waste: Effect on methane yields, dewaterability and solids reduction. *Water Research* 132, 158-
 327 166 (2018).
- 328 [18] Yuan, T., Cheng, Y., Zhang, Z.: Comparative study on hydrothermal treatment as pre- and post-treatment of anaerobic
 329 digestion of primary sludge: Focus on energy balance, resources transformation and sludge dewaterability. *Applied*
 330 *Energy* 239, 171-180 (2019).
- 331 [19] Angelidaki, I., Alves, M., Bolzonella, D., Borzacconi, L., Campos, J.L., Guwy, A.J., Kalyuzhnyi, S., Jenicek, P., van
 332 Lier, J.B.: Defining the biomethane potential (BMP) of solid organic wastes and energy crops: a proposed protocol for
 333 batch assays. *Water Sci. Technol.* 59(5), 927-934 (2009).
- 334 [20] Holliger, C., Alves, M., Andrade, D., Angelidaki, I., Astals, S., Baier, U., Bougrier, C., Buffière, P., Carballa, M., de
 335 Wilde, V., Ebertseder, F., Fernández, B., Ficara, E., Fotidis, I., Frigon, J.C., de Lacroix, H.F., Ghasimi, D.S., Hack, G.,
 336 Hartel, M., Heerenklage, J., Horvath, I.S., Jenicek, P., Koch, K., Krautwald, J., Lizasoain, J., Liu, J., Mosberger, L., Nistor,
 337 M., Oechsner, H., Oliveira, J.V., Paterson, M., Pauss, A., Pommier, S., Porqueddu, I., Raposo, F., Ribeiro, T., Rüsç
 338 Pfund, F., Strömberg, S., Torrijos, M., van Eckert, M., van Lier, J., Wedwitschka, H., Wierinck, I.: Towards a
 339 standardization of biomethane potential tests. *Water Science & Technology* 74 (11), 2515-2522 (2016).
- 340 [21] Nielfa, A., Cano, R., Fdz-Polanco, M.: Theoretical methane production generated by the co-digestion of organic
 341 fraction municipal solid waste and biological sludge. *Biotechnol. Reports* 5, 14-21 (2015).
- 342 [22] APHA/AWWA/WEF: Standards Methods for the Examination of Water and Wastewater, 20th ed., United Book
 343 Press, Inc., Baltimore, Maryland (1998).
- 344 [23] Weisburg, W.G., Barns, S.M., Pelletier, D.A., Lane, D.J.: 16S ribosomal DNA amplification for phylogenetic study.
 345 *J. Bacteriol.* 173, 697-703 (1991).
- 346 [24] Muyzer, G., De Waal, E.C., Uittierlinden, A.G.: Profiling of complex microbial populations by denaturing gradient
 347 gel electrophoresis analysis of polymerase chain reaction- amplified genes coding for 16S rRNA. *Am. Soc. Microbiol.*
 348 59, 695-700 (1993).
- 349 [25] Grosskopf, R., Janssen, P.H., Liesack, W.: Diversity and structure of the methanogenic community in anoxic rice
 350 paddy soil microcosms as examined by cultivation and direct 16S rRNA gene sequence retrieval. *Appl Environ Microbiol*
 351 64(3), 960-969 (1998).
- 352 [26] Lane, D.J.: 16S/23S rRNA sequencing. *Nucleic Acid Techniques in Bacterial Systematics*. Wiley, Chichester (1991).
- 353 [27] Lampis, S., Ferrari, A., Cunha-Queda, A. C., Alvarenga, P., Di Gregorio, S., Vallini, G.: Selenite resistant
 354 rhizobacteria stimulate SeO₃²⁻ phytoextraction by *Brassica juncea* in bioaugmented water-filtering artificial beds. *Environ*
 355 *Sci. Pollut. Res. Int.* 16, 663-670 (2009).
- 356 [28] Zeppilli, M., Villano, M., Aulenta, F., Lampis, S., Vallini, G., Majone, M.: Effect of the anode feeding composition
 357 on the performance of a continuous-flow methane-producing microbial electrolysis cell. *Environ Sci Pollut Res Int.*
 358 22(10): 7349-60 (2015).
- 359 [29] Altschul, S.F., Madden, T.L., Schaffer, A.A., Zhang, J., Zhang, Z., Miller, W., Lipman, D.J.: Gapped BLAST and
 360 PSI-BLAST: a new generation of protein database search programs. *Nucleic Acids Res* 25(17), 3389-3402 (1997).
- 361 [30] Yoon, S. H., Ha, S. M., Kwon, S., Lim, J., Kim, Y., Seo, H. and Chun, J.: Introducing EzBioCloud: A taxonomically
 362 united database of 16S rRNA and whole genome assemblies. *Int J Syst Evol Microbiol.* 67, 1613-1617 (2017).
- 363 [31] Pedizzi, C., Lema, J.M., Carballa, M.: A combination of ammonia stripping and low temperature thermal pre-
 364 treatment improves anaerobic post-digestion of the supernatant from organic fraction of municipal solid waste treatment.
 365 *Waste Management* 78, 271-278 (2018).
- 366 [32] Battista, F., Bolzonella, D.: Some critical aspects of the enzymatic hydrolysis at high dry-matter content: a review.
 367 *Biofuels, Bioprod. Bioref.* 12 (4), 711-723 (2018).
- 368 [33] Golkowska, K., Greger M.: Anaerobic digestion of maize and cellulose under thermophilic and mesophilic conditions
 369 – a comparative study. *Biomass Bioenergy*, 545-554 (2013).
- 370 [34] Wu, J., Hu, Y.-Y., Wang, S.-F., Cao, Z.-P., Li, H.-Z., Fu, X.-M., Wang, K.-J., Zuo, J.-E.: Effects of thermal treatment
 371 on high solid anaerobic digestion of swine manure: Enhancement assessment and kinetic analysis. *Waste management*
 372 62, 69-75 (2017).

- 373 [35] Pecchi, M., Baratieri, M.: Coupling anaerobic digestion with gasification, pyrolysis or hydrothermal carbonization:
374 A review. *Renewable and Sustainable Energy Reviews* 105, 462-475 (2019).
- 375 [36] Kim, D., Lee, K., Park, K.Y.: Enhancement of biogas production from anaerobic digestion of waste activated sludge
376 by hydrothermal pre-treatment. *Int Biodeter Biodegr* 101, 42–6 (2015).
- 377 [37] Yuan, T., Cheng, Y., Zhang, Z., Lei, Z., Shimizu, K.: Comparative study on hydrothermal treatment as pre- and post-
378 treatment of anaerobic digestion of primary sludge: Focus on energy balance, resources transformation and sludge
379 dewaterability. *Applied Energy* 239, 171–180 (2019).
- 380 [38] Ariesyady, H.D., Ito, T., Okabe, S.: Functional bacterial and archaeal community structures of major trophic groups
381 in a full-scale anaerobic sludge digester. *Water Res* 41, 1554–1568 (2007).
- 382 [39] Zhang, D., Fu, X., Jial, S., Dai, L., Wu, B., Dai, X.: Excess sludge and herbaceous plant co-digestion for volatile
383 fatty acids generation improved by protein and cellulose conversion enhancement. *Environ Sci Pollut Res* 23, 1492–1504
384 (2016).
- 385 [40] Zamanzadeh, M., Hagen, L.H., Svensson, K., Linjordet, R., Horn, S.J.: Anaerobic digestion of food waste - Effect
386 of recirculation and temperature on performance and microbiology. *Water Res.* 96: 246-254 (2016).
- 387 [41] Rademacher, A., Nolte, C., Schönberg, M., Klocke, M.: Temperature increases from 55 to 75 °C in a two-phase
388 biogas reactor result in fundamental alterations within the bacterial and archaeal community structure. *Appl Microbiol*
389 *Biotechnol.* 96(2), 565-576 (2012).
- 390 [42] Frock, A.D., Notey, J.S., Kelly, R.M.: The genus *Thermotoga*: recent developments. *Environ Technol.* 31(10), 1169-
391 81 (2010).