

# Optimization of Temperature-separated Two-stage Anaerobic Fermentation Process Treating waste activated sludge and food waste

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## ABSTRACT

In this study, a temperature-separated two-stage anaerobic fermentation process of waste activated sludge (WAS) and food waste (FW) was performed to optimize the sludge retention time of methane producing stage (M-SRT) and sludge return ratios of the system (SRR). Results show that higher M-SRT improved the removal of organics and energy yield of the system but reduced the stability of hydrogen production in H<sub>2</sub>-reactor. In this study, all the M-SRT conditions were below the threshold level. The highest total energy yield and volatile solid (VS) removal efficiency of the two-stage system of 8.98 kJ/g-VS and 63.8% was observed under the optimized M-SRT of 12 d. Return sludge could complement alkalinity for acidogenic stage especially with higher SRR, but the methanogens in it inhibited hydrogen production and caused stage transfer eventually. The optimized SRR was 1:1 with comprehensive consideration of energy yield, VS removal efficiency and alkali dosage. The dominant bacterial families in acidogenic stage were *Ruminococcaceae* and *Clostridiaceae* in this study.

## KEYWORDS

Two-stage anaerobic fermentation; waste activated sludge; food waste; hydrogen production; methane production.

## 1. INTRODUCTION

Anaerobic fermentation can realize the waste reduction and bioenergy production from organic solid wastes[1]. In recent years, co-digestion has been widely used to improve reactor performance[2]. Meanwhile, the amount of food waste (FW) generated in 2015 was estimated to 56.57 million tons in mainland China[3]. However, traditional solid wastes disposal methods cannot efficiently transform the valuable energy in WAS and FW biomass. Compared with the single stage system, two-stage anaerobic fermentation could improve the energy recovery efficiency and stability of the system[4]. In this study, a temperature-separated two-stage anaerobic fermentation process for the treatment of waste activated sludge (WAS) and FW was performed to optimize the sludge retention time of methane producing stage (M-SRT) and sludge return ratios of the system (SRR).

## 2. MATERIALS AND METHODS

### 2.1 Seed sludge and substrates

The hydrogen-producing and methane-producing seed sludge were both obtained from the anaerobic digestion reactors with mixed substrates of WAS and FW. The hydrogen-producing seed sludge was base treated for 24 h, and then cultured for 7 days at 55°C, under anaerobic condition with FW as substrates. During base treatment, pH value of the sludge was adjusted to 12.0 with 1.0 mol/L NaOH, and then adjusted to 7.0 with 1.0 mol/L HCl after 24 h[5].

Substrates in this study included WAS and FW. Based on previous research work[6], the volatile solids (VS) of the substrates was 43.5g/L, among which the FW accounted for 54% (VS). The WAS was collected from the secondary sedimentation tank of Zhangguizhuang wastewater treatment plant in Tianjin, China. The FW was collected from the cafeteria of Tianjin University. The characteristics of substrates were shown in Table 1.

**Table 1**

Characteristics of waste activated sludge and food waste in feedstock.

Parameters	Waste activated sludge	Food waste	Feedstock
pH	7.1	5.5	6.6
VS/TS (%)	59.1	97.7	75.8
Total COD (mg/g-TS)	1009.7	1189.9	1080.0
Soluble carbohydrate (mg/g-TS)	0.4	289.8	146.7
Soluble protein (mg/g-TS)	0.2	18.8	10.2
Alkalinity (mg-CaCO <sub>3</sub> /g-TS)	8.1	0.2	3.8

### 2.2 Experimental conditions

Schematic of the temperature-separated two-stage anaerobic fermentation process (BIOTECH-5BG, Shanghai Baoxing Bio-engineering Equipment Co. Ltd, China) was shown in Fig.1. The hydrogen producing stage (H<sub>2</sub>-reactor)

was thermophilic fermentation with operating temperature of 55°C, while the methane producing stage (CH<sub>4</sub>-reactor) was mesophilic fermentation with operating temperature of 37°C. The effective volume of two reactors were 1.5 L and 5 L, with fermentation broth volume of 1 L and 4 L respectively. Both the two reactors were stirred completely with stirring speed of 70 rpm and 300 rpm respectively. During the start-up period, the hydrogen-producing seed sludge and mixed substrates with volume ratio of 2:1 were added into H<sub>2</sub>-reactor, and only seed sludge was added into CH<sub>4</sub>-reactor. Both the two reactors were purged by nitrogen gas for 10 min before sealing. Before formal operation, mixed substrates were added into the reactors for continuous acclimation for a month. The pH value of the two reactors were controlled at 5.0-5.5 and 6.7-7.4 respectively by the addition of K<sub>2</sub>HPO<sub>4</sub>. The two reactors were fed twice a day.

The operation process for optimizing M-SRT was divided into three periods, with different M-SRT of 4 d, 8 d and 12 d. VS of the feedstock substrates in H<sub>2</sub>-reactor was constant at 43.5g/L in each period. SRT and organic loading rate (OLR) was determined by changing feedstock volume of CH<sub>4</sub>-reactor. Sludge from CH<sub>4</sub>-reactor was recirculated into H<sub>2</sub>-reactor to complement its alkalinity. The SRR (volume ratio of return sludge from CH<sub>4</sub>-reactor to mixed substrates in H<sub>2</sub>-reactor) under condition C-1 and C-2 was 1:1. Considering the insufficient sludge from CH<sub>4</sub>-reactor under condition C-3, all the sludge was recirculated and the SRR was 2:3. The operation process for optimizing SRR was divided into three periods with different SRR at 1:1, 1:2 and 2:1. Total feedstock volume, SRT of hydrogen producing stage (H-SRT) and total VS of the mixed substrates in feedstock were constant in each period. SRR were determined by changing volume ratio of return sludge to mixed substrates. The actual SRT, OLR and operating time of different periods considering return sludge were shown in Table 2.

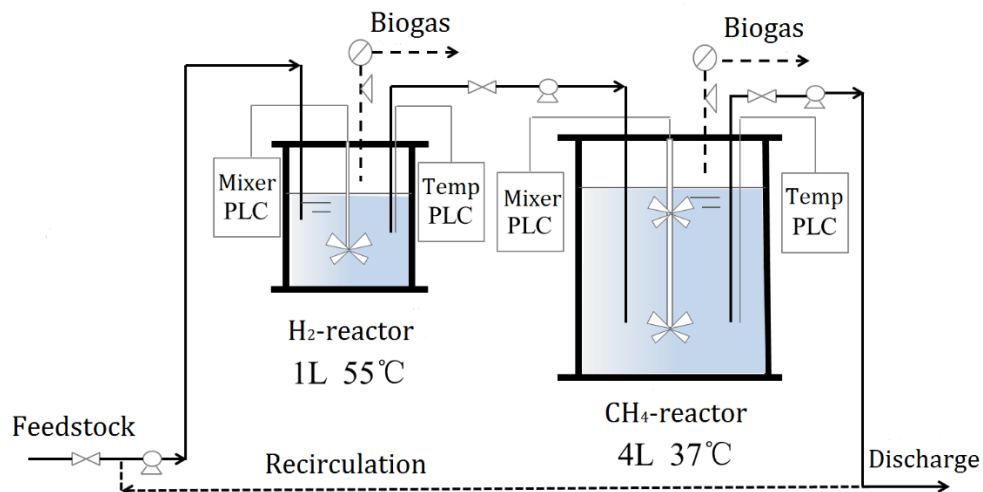


Fig. 1. Schematic of the temperature-separated two-stage anaerobic process.

Table 2

Operating conditions of different periods in the two-stage anaerobic fermentation process.

Operating condition	H <sub>2</sub> -reactor		CH <sub>4</sub> -reactor		Operating time (d)
	SRT (d)	OLR (g-VS/L/d)	SRT (d)	OLR (g-VS/L/d)	
C-1	1	30.8	4	5.7	1~11
C-2	1	30.3	8	2.9	12~32
C-3	1.2	27.4	12	1.9	33~56
SRR=1:1	1	30.6	4	5.5	1~16
SRR=1:2	1	26.9	4	4.7	17~38
SRR=2:1	1	34.3	4	6.4	39~58

### 2.3 Analytical methods

The volume of biogas produced in the reactors were measured by gas meter calibration device with U-tube. Compositions of biogas were analysed by a gas chromatograph (BEIFEN 3040, China) equipped with a thermal conductivity detector and a stainless-steel packed column (TDX-01, 2 m). Argon was used as the carrier gas at a flow rate of 35 ml/min. The operation temperatures of the injection port, oven and detector were 100°C, 100°C and 130°C,

respectively. Ethanol and total volatile fatty acids (TVFAs) including acetate, propionate, butyrate, i-butyrate, valerate and i-valerate in the mixed liquor were analysed by another gas chromatograph (SP6890, China) equipped with a flame ionization detector and a fused-silica capillary column (HP-FFAP, 0.53 mm×10 m×1 μm). Nitrogen was used as the carrier gas with a flow rate of 6 ml/min and split ratio of 10:1. The temperature of injection port and detector were 200°C and 250 °C respectively. Soluble carbohydrate was analysed using anthrone-sulfuric acid method with glucose as standard[7], and soluble proteins was analysed by Lowry method[8]. Chemical oxygen demand (COD) was measured by HACH method (HACH, USA) and pH was measured by glass electrode method (PB-10 Sartorius, Germany). The samples for analysis of soluble carbohydrate and proteins were filtrated by 0.45 μm membrane before detecting. TS, VS, alkalinity and ammonium were determined according to Standard Methods (APHA, 2005).

#### 2.4 Microbial analysis

Discharged sludge samples from H<sub>2</sub>-reactor were collected after each period of operation for microbial community analysis. High-throughput sequencing was carried out on Illumina HiSeq platform (Novogene Co. Ltd., Beijing, China). DNA was extracted using the PowerSoil DNA Isolation Kit (MoBio Laboratories Inc., USA) according to the manufacturer's instructions. The primer set 515F (5'-GTGCCAGCAGCCGCGGTAA-3') and 806R (5'-GGACTACCAGGGTATCTAAT-3') was used in the amplification of the V4 region of the bacterial 16S rRNA gene.

### 3. RESULTS AND DISCUSSION

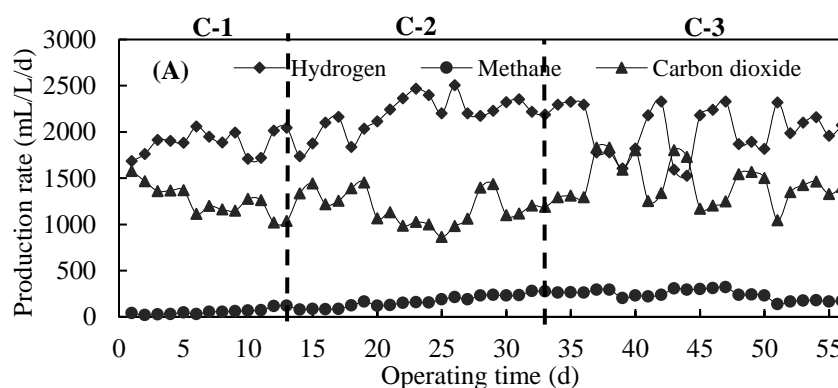
#### 3.1 Biogas production and energy yields

##### 3.1.1 Hydrogen and methane production in H<sub>2</sub>-reactor

The production rate of H<sub>2</sub>, CO<sub>2</sub> and CH<sub>4</sub> in H<sub>2</sub>-reactor with different M-SRT and SRR are shown in Fig. 2. The biogas composition and H<sub>2</sub> yields are shown in Table 3. The biogas in H<sub>2</sub>-reactor was mainly composed of H<sub>2</sub> and CO<sub>2</sub> and H<sub>2</sub> yields were close to the result of earlier research (76.8 mL-H<sub>2</sub>/g-VS<sub>added</sub>)[6].

Both average production rate of H<sub>2</sub> and H<sub>2</sub> yield increased with higher M-SRT of 8 d. A small amount of methane was also observed in biogas of H<sub>2</sub>-reactor, especially with M-SRT of 8 d and 12 d. The production of CH<sub>4</sub> may be contributed to methanogens carried by return sludge from CH<sub>4</sub>-reactor. With synergistic interaction between fermenting bacteria and methanogenic archaea[9], interspecies hydrogen exchange can happen in the co-culture system, which means fermentation of glucose to acetate and H<sub>2</sub> in syntrophic association with H<sub>2</sub>-consuming microorganisms that keep low partial pressure of H<sub>2</sub>[10]. Meanwhile, high partial pressure of hydrogen may inhibit H<sub>2</sub> production[11]. Hence when the hydrogen produced by acetogenic bacteria was consumed by methanogens, especially hydrogenotrophic methanogens, the production of VFAs could be proceeded smoothly. Promotion of CH<sub>4</sub> production with higher M-SRT indicated better activity of methanogens from return sludge. Therefore, H<sub>2</sub> production in H<sub>2</sub>-reactor could be promoted by properly prolonging M-SRT. However, H-SRT was relatively extended caused by insufficient return sludge when M-SRT was 12 d. Though with higher activity, the quantity of methanogens from return sludge decreased and thus the CH<sub>4</sub> production of H<sub>2</sub>-reactor in period C-3 kept almost the same level with the end of period C-2. Variance of production rate of H<sub>2</sub> and CO<sub>2</sub> increased with higher M-SRT, indicating worse stability of H<sub>2</sub> and CO<sub>2</sub> production. Meanwhile, H<sub>2</sub> yield decreased slightly when M-SRT was 12 d. Possibly the methanogens affected the activity of hydrogen producing acetogenic bacteria and reduced stability of acidogenic stage in the long run. Nevertheless, the H<sub>2</sub> yield was not influenced severely with M-SRT of 12 d.

The production rate of H<sub>2</sub>, H<sub>2</sub> content in biogas and H<sub>2</sub> yield both decreased with higher SRR. More seriously, after running for several days when SRR was 2:1, the production rate of H<sub>2</sub> dropped sharply while the production rate of CH<sub>4</sub> increased on the contrary. With the addition of sodium 2-bromoethanesulphonate (BESA, a kind of methanogenic inhibitor) into H<sub>2</sub>-reactor, the production rate of H<sub>2</sub> became even higher than that when SRR was 1:2 and the production rate of CH<sub>4</sub> dropped back to original level. The recovery seemed to imply that the collapse of H<sub>2</sub>-reactor was due to methanogens from return sludge, especially with higher SRR. The production rate of H<sub>2</sub> and CO<sub>2</sub>, mainly substrates for synthesis of CH<sub>4</sub>, both decreased corresponding to higher CH<sub>4</sub> production when the collapse happened. Therefore, it might indicate the stage in H<sub>2</sub>-reactor transferred from acidogenic stage to methanogenic stage when SRR was 2:1.



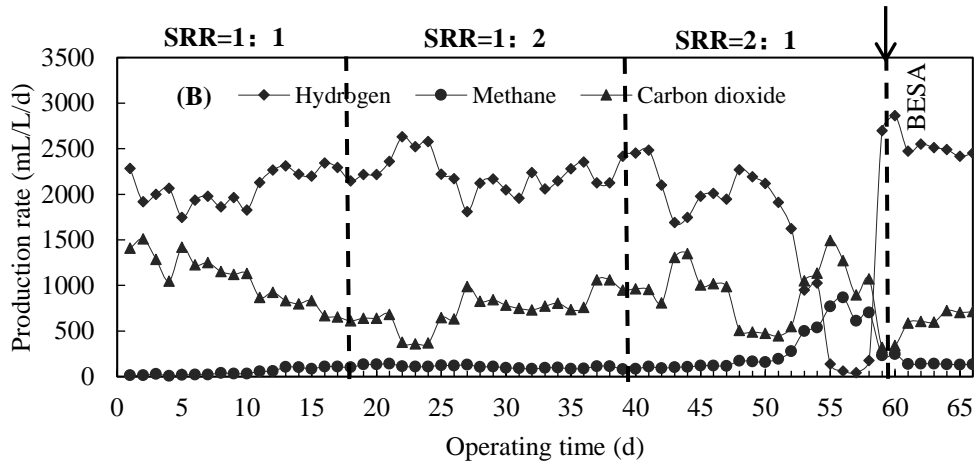


Fig. 2. Production rate of H<sub>2</sub>, CO<sub>2</sub> and CH<sub>4</sub> in H<sub>2</sub>-reactor with different M-SRT (A) and SRR (B).

**Table 3**

Biogas composition and hydrogen yields in H<sub>2</sub>-reactor under different operating conditions.

Operating condition	Biogas composition (%)			H <sub>2</sub> yield (mL-H <sub>2</sub> /g-VS <sub>added</sub> )
	H <sub>2</sub>	CO <sub>2</sub>	CH <sub>4</sub>	
C-1	51.0-64.3	34.8-47.8	0.6-2.3	85.5±5.7
C-2	54.9-67.7	26.6-42.4	2.4-7.5	99.8±9.3
C-3	43.0-66.2	29.9-48.7	3.9-8.3	93.1±11.8
SRR=1:1	54.9-75.1	21.4-44.6	0.3-3.6	95.0±8.6
SRR=1:2	61.9-84.4	12.0-33.7	2.8-4.4	102.0±8.8
SRR=2:1	2.7-77.0	17.2-62.2	2.4-39.4	79.5±33.5

### 3.1.2 Methane production in CH<sub>4</sub>-reactor

The sludge from H<sub>2</sub>-reactor was discharged into CH<sub>4</sub>-reactor. Production rate of CH<sub>4</sub> and CO<sub>2</sub> in CH<sub>4</sub>-reactor with different SRR are shown in Fig. 3. The biogas composition and CH<sub>4</sub> yields are shown in Table 4.

The M-SRT were determined by changing the feedstock volume of CH<sub>4</sub>-reactor and thus production rate of CH<sub>4</sub> decreased with increasing M-SRT. CH<sub>4</sub> content in biogas and CH<sub>4</sub> yields both increased with higher M-SRT. The highest CH<sub>4</sub> yield, 218.0 mL-CH<sub>4</sub>/g-VS<sub>added</sub>, was observed when M-SRT was 12 d. Possibly higher M-SRT benefitted the growth of methanogens with low growth rates, such as acetotrophic methanogens[12]. Hence higher activity of methanogens promoted CH<sub>4</sub> production. The change of sludge in H<sub>2</sub>-reactor with higher M-SRT might also improve the dissolution and hydrolysis of organic matters, which further boosted CH<sub>4</sub> production in CH<sub>4</sub>-reactor.

The stage transfer in H<sub>2</sub>-reactor promoted activity of methanogens in H<sub>2</sub>-reactor. Then the feedstock of CH<sub>4</sub>-reactor, which was from H<sub>2</sub>-reactor, promoted CH<sub>4</sub> production in CH<sub>4</sub>-reactor and further inhibited H<sub>2</sub> production in H<sub>2</sub>-reactor. For the mixed substrates added to H<sub>2</sub>-reactor in this study, the volume decreased with higher SRR while the total VS was constant, as mentioned in *Experimental conditions*. Therefore, VS concentration of the mixed substrate increased with higher SRR, leading to higher VS concentration of feedstock in CH<sub>4</sub>-reactor. Production rate of CH<sub>4</sub> in CH<sub>4</sub>-reactor remained stable with different SRR and only increased when H<sub>2</sub> production was inhibited in H<sub>2</sub>-reactor. Possibly the inhibition of acidogenesis process caused lower degradation rate of organic matters in sludge of H<sub>2</sub>-reactor. It provided more biodegradable substrates in feedstock of CH<sub>4</sub>-reactor and led to the temporary improvement. The lower CH<sub>4</sub> yields with higher SRR indicated that inhibition of H<sub>2</sub> production in H<sub>2</sub>-reactor was unfavourable for CH<sub>4</sub> production in CH<sub>4</sub>-reactor. Namely, the hydrogen producing process played an important role in subsequent CH<sub>4</sub> producing stage.

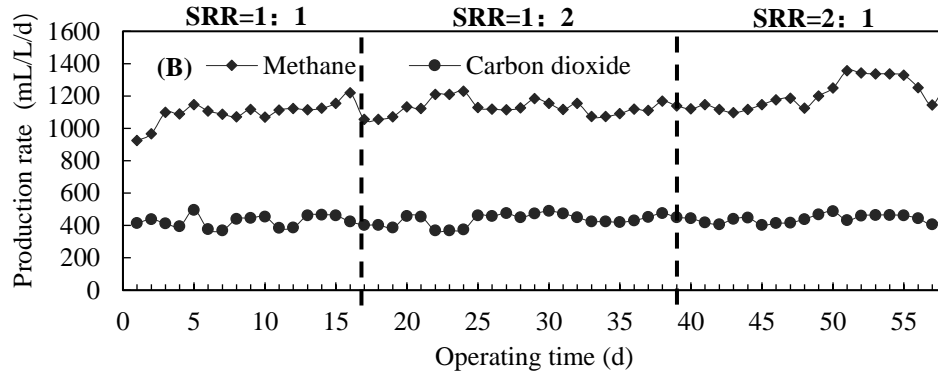


Fig. 3. Production rate of CH<sub>4</sub> and CO<sub>2</sub> in CH<sub>4</sub>-reactor with different SRR.

Table 4

Biogas composition and methane yields in CH<sub>4</sub>-reactor under different operating conditions.

Operating condition	Biogas composition (%)		Methane yield (mL-CH <sub>4</sub> /g-VS <sub>added</sub> )
	CH <sub>4</sub>	CO <sub>2</sub>	
C-1	66.4-71.4	28.6-33.6	187.4±12.5
C-2	67.7-72.0	28.0-32.4	204.4±7.6
C-3	68.6-73.2	26.9-31.4	218.0±13.2
SRR=1:1	68.8-74.6	25.4-31.2	200.2±14.1
SRR=1:2	70.1-76.7	23.3-29.9	240.3±12.8
SRR=2:1	71.4-75.8	24.2-28.6	188.4±18.5

### 3.1.3 Energy yields

Energy yields of H<sub>2</sub>-reactor, CH<sub>4</sub>-reactor and total energy yields of the two-stage system under different operating conditions are shown in Table 5. The combustion heat of H<sub>2</sub> and CH<sub>4</sub> are 285.84 kJ/mol and 890.35 kJ/mol respectively[13]. Results show that the main energy yields of the two-stage system came from CH<sub>4</sub>. Energy yields of the two reactors were consistent with their respective tending of H<sub>2</sub> and CH<sub>4</sub> production. The total energy yields of the two-stage system increased with higher M-SRT but decreased with higher SRR, reaching maximum value of 8.98 kJ/g-VS<sub>added</sub> when M-SRT was 12 d and 9.66 kJ/g-VS<sub>added</sub> when SRR was 1:2 respectively. H<sub>2</sub> production was inhibited slightly while CH<sub>4</sub> production in CH<sub>4</sub>-reactor was enhanced with higher M-SRT and thus the total energy yields increased. To some extent, with appropriate concentration of methanogens in H<sub>2</sub>-reactor, the energy yields of the two-stage system would be promoted. The energy yield of CH<sub>4</sub> in H<sub>2</sub>-reactor increased with higher SRR due to stage transfer. However, with worse production of H<sub>2</sub> in H<sub>2</sub>-reactor and CH<sub>4</sub> in CH<sub>4</sub>-reactor, the total energy yields decreased with higher SRR. Therefore, excessive methanogens may affect the stability of H<sub>2</sub>-reactor and lead to stage transfer, which would inhibit the operation of two-stage system at last.

Table 5

Energy yields of H<sub>2</sub>-reactor, CH<sub>4</sub>-reactor and total energy yields of the two-stage system under different operating conditions.

Operating condition	H <sub>2</sub> -reactor			CH <sub>4</sub> -reactor		Total energy yield (kJ/g-VS <sub>added</sub> )
	Energy yield of H <sub>2</sub> (kJ/g-VS <sub>added</sub> )	Energy yield of CH <sub>4</sub> (kJ/g-VS <sub>added</sub> )	Proportion of total energy yield (%)	Energy yield (kJ/g-VS <sub>added</sub> )	Proportion of total energy yield (%)	
C-1	0.91±0.06	0.07±0.03	12.9	6.56±0.44	87.1	7.53±0.48
C-2	1.06±0.10	0.24±0.09	15.4	7.15±0.27	84.6	8.46±0.29
C-3	0.99±0.13	0.37±0.08	15.1	7.63±0.46	84.9	8.98±0.46
SRR=1:1	1.01±0.09	0.07±0.05	13.3	7.01±0.49	86.7	8.09±0.55
SRR=1:2	1.08±0.09	0.17±0.02	12.9	8.41±0.45	87.1	9.66±0.51
SRR=2:1	0.90±0.35	0.33±0.37	15.7	6.59±0.65	84.3	7.82±0.68

### 3.2 VS removal efficiency

The VS concentration of discharged sludge from H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor, VS removal efficiency of the two reactors and total VS removal efficiency of the two-stage system under different operating conditions are shown in Fig.4 and Table 6.

VS removal efficiency of H<sub>2</sub>-reactor with M-SRT of 8 d was lower than that with M-SRT of 4 d, while the production of H<sub>2</sub> and CH<sub>4</sub> was promoted. Possibly it was caused by more degradation of carbohydrates and less degradation of proteins and lipids. H<sub>2</sub> yields from carbohydrates were much higher than that from proteins and lipids[14]. The promotion of methanogens in H<sub>2</sub>-reactor may affect metabolic pathway of different organics. However, the highest VS removal efficiency in H<sub>2</sub>-reactor was 53.3% with M-SRT of 12 d. This result could be caused by the insufficient return sludge from CH<sub>4</sub>-reactor mentioned in *Experimental conditions*. The decrease of feedstock volume led to longer H-SRT, and thus more organic matters could be consumed in H<sub>2</sub>-reactor. The VS removal efficiency of CH<sub>4</sub>-reactor (21.0%~27.0%) was at the same level of previous research (23.7%-25.1%)[6]. The trend of VS removal efficiency in CH<sub>4</sub>-reactor with different M-SRT was contrary to that in H<sub>2</sub>-reactor. It might indicate the organics which had not been consumed in H<sub>2</sub>-reactor would be used in CH<sub>4</sub>-reactor, especially those hardly degraded substrates, such as proteins and lipids. However, most of the biodegradable organic matters in feedstock had been degraded in H<sub>2</sub>-reactor and thus limited the utilization of organics in CH<sub>4</sub>-reactor. Nevertheless, the VS removal efficiency of the two-stage system increased with increasing M-SRT and reached highest value of 63.8% when M-SRT was 12 d. This result was 17.7% higher than the VS removal efficiency under relatively equal OLR condition (29.3 g-VS/L/d)[6]. It seemed that higher M-SRT promoted the degradation of organics in the two-stage system.

As mentioned in *Experimental conditions*., VS concentration of the mixed substrate increased with higher SRR, leading to higher VS concentration of feedstock in both reactors. The total VS removal efficiency of two-stage system decreased with higher SRR and the highest total VS removal efficiency of 63.3% was obtained when SRR was 1:2. This trend was consistent with CH<sub>4</sub> production of CH<sub>4</sub>-reactor, as well as energy yield of CH<sub>4</sub>-reactor and the two-stage system. It might suggest that the temperature-separated two-stage anaerobic fermentation system performed better than temperature-separated methanogenic stage system.

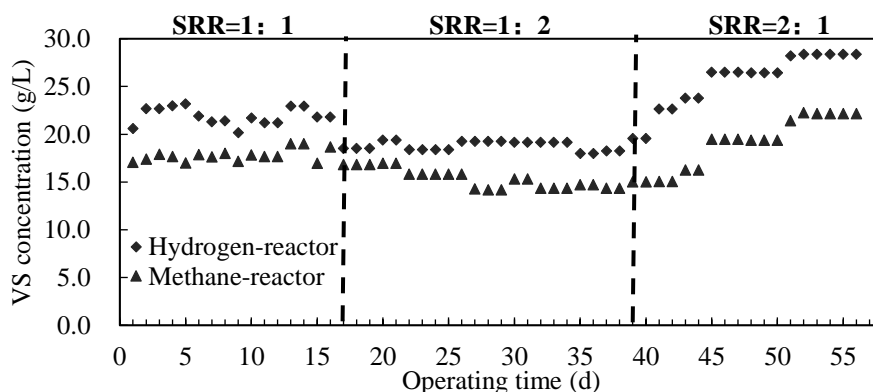


Fig. 4. VS concentration of discharged sludge from H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor with different SRR.

Table 6

VS removal efficiency and concentration of discharged sludge from H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor, total VS removal efficiency of the two-stage system under different operating conditions.

Operating condition	VS concentration of discharged sludge (g/L)		VS removal efficiency (%)		
	H <sub>2</sub> -reactor	CH <sub>4</sub> -reactor	H <sub>2</sub> -reactor	CH <sub>4</sub> -reactor	Two-stage system
C-1	22.7±0.8	18.0±0.5	50.7±1.4	21.0±2.7	61.1±1.2
C-2	23.1±0.7	16.8±0.7	48.4±2.0	27.0±3.5	62.1±1.9
C-3	22.4±0.7	16.7±0.5	53.3±2.9	25.2±2.2	63.8±1.5
SRR=1:1			52.1±1.7	18.9±4.3	61.2±1.4
SRR=1:2	Shown in Fig. 4		52.2±3.6	22.7±7.7	63.3±1.5
SRR=2:1			51.4±1.0	18.4±2.0	60.6±1.2

### 3.3 pH and alkalinity

The pH, average alkali dosage and alkalinity of sludge in H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor under different operating

conditions are shown in Table 7. In two-stage anaerobic fermentation process, return sludge from methanogenic stage can complement alkalinity for acidogenic stage. However, with SRR of 1:1, alkalinity of return sludge cannot meet the demand of acidogenic stage to maintain appropriate pH. In this research,  $K_2HPO_4$  was added into  $H_2$ -reactor to maintain pH value of 5.14-5.21, which is in effective range of 5.0-6.5 for hydrogen producing acetogenic bacteria[15]. In  $CH_4$ -reactor, pH maintained a suitable range of 7.03-7.06 for methanogens without external alkalinity[16]. All the alkalinity in two reactors were in favourable range of 1000-5000 mg- $CaCO_3$ /L[17].

The volume of return sludge decreased with M-SRT of 12 d as mentioned above and the reduction of supplementary alkalinity caused more alkali dosage into  $H_2$ -reactor. More acidic substrates such as VFAs produced together with  $H_2$  and more ammonium decomposed from proteins may explained the lower pH and higher alkalinity in  $H_2$ -reactor with M-SRT of 8 d compared to that of 4 d. Alkalinity in  $CH_4$ -reactor increased with higher M-SRT. Possibly it was caused by the promotion of  $CH_4$  production, which provided better degradation of acidic substrates and more ammonium production.

The average alkali dosage reduced nearly by half with increased alkalinity in  $H_2$ -reactor when SRR increased from 1:2 to 1:1. The highest pH value and alkalinity in  $H_2$ -reactor was obtained when SRR was 2:1 even with no external alkalinity. Obviously, the return sludge made a great contribution to supplementary alkalinity in acidogenic stage. However, the pH value continued to accelerate over the operating time, up to 5.68 when  $H_2$  production was inhibited apparently. It seems that with higher SRR, return sludge carried not only excessive alkalinity but also more methanogens into  $H_2$ -reactor. They promoted  $CH_4$  production and thus caused higher pH value and alkalinity in  $H_2$ -reactor, which in turn inhibited  $H_2$  production. Considering the economic cost of external alkalinity as well as running effect, SRR of 1:1 was considered a favorable ratio in this study.

**Table 7**  
pH, average alkali dosage and alkalinity in  $H_2$ -reactor and  $CH_4$ -reactor under different operating conditions.

Operating condition	$H_2$ -reactor			$CH_4$ -reactor	
	pH	Average alkali dosage (g/L/d)	Alkalinity (mg- $CaCO_3$ /L)	pH	Alkalinity (mg- $CaCO_3$ /L)
C-1	5.21±0.06	0.791	1951.6±222.5	7.03±0.02	3904.1±124.3
C-2	5.18±0.08	0.922	2103.2±300.9	7.03±0.03	4307.6±225.0
C-3	5.14±0.07	1.640	2376.3±207.4	7.06±0.05	4781.5±255.2
SRR=1:1	5.17±0.04	1.331	2183.7±290.8	7.05±0.03	3771.5±254.6
SRR=1:2	5.11±0.04	2.206	1720.1±68.1	7.01±0.02	3830.2±169.0
SRR=2:1	5.45±0.15	0.000	3054.5±416.3	7.14±0.08	4401.4±315.5

### 3.4 Soluble organic matters

#### 3.4.1 Soluble carbohydrate, soluble proteins and ammonium

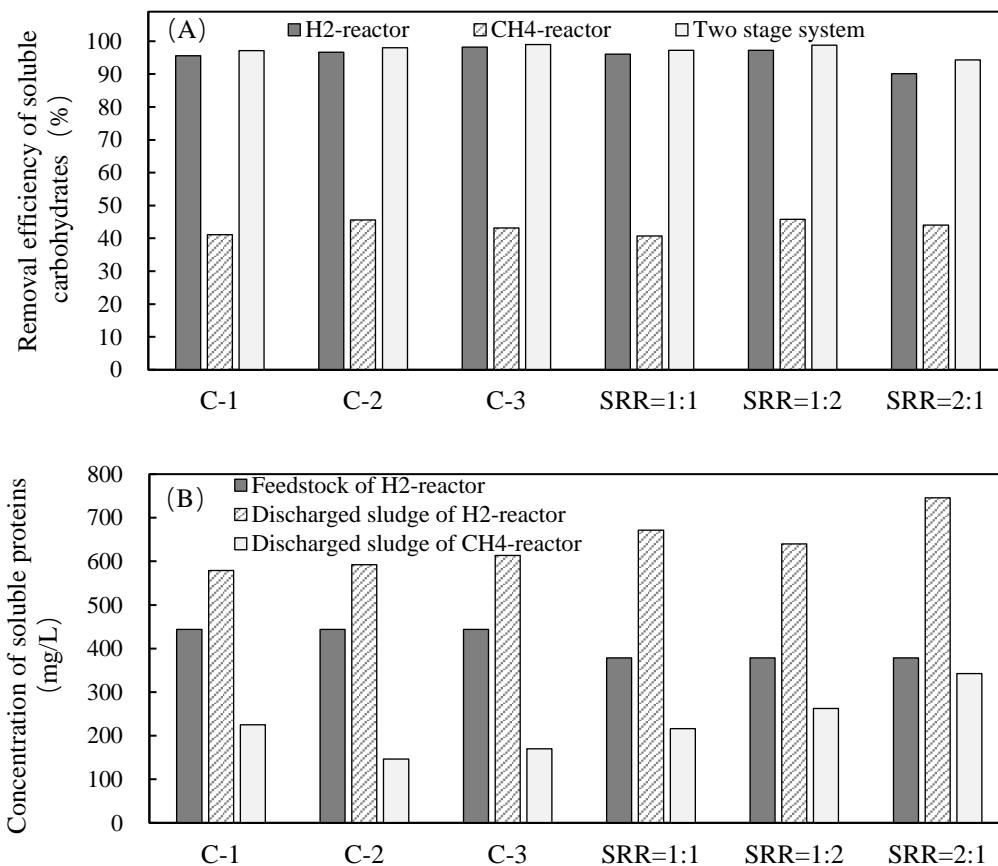
Carbohydrates and proteins are the main biodegradable organic matters in substrates for anaerobic digestion. Removal efficiency of soluble carbohydrates, concentration of soluble proteins and ammonia nitrogen under different operating conditions were shown in Fig. 5.

Most soluble carbohydrates were removed in  $H_2$ -reactor with removal efficiency of 95.6%-98.2%. The methane in  $CH_4$ -reactor mainly came from the degradation of VFAs as there is few soluble carbohydrates flowed into  $CH_4$ -reactor. The slightly increase of removal efficiency of soluble carbohydrates in  $H_2$ -reactor with higher M-SRT indicated that with lower stability of  $H_2$  production, the acidogenesis process in  $H_2$ -reactor was still promoted. Therefore, removal efficiency of soluble carbohydrates in the two-stage system was promoted with higher M-HRT. Of all the operating conditions, the minimum removal efficiency of soluble carbohydrates in  $H_2$ -reactor was observed with SRR of 2:1. Considering the stage transfer in  $H_2$ -reactor, it seemed that the utilization of soluble carbohydrates mainly occurred in acidogenic stage in the two-stage system.

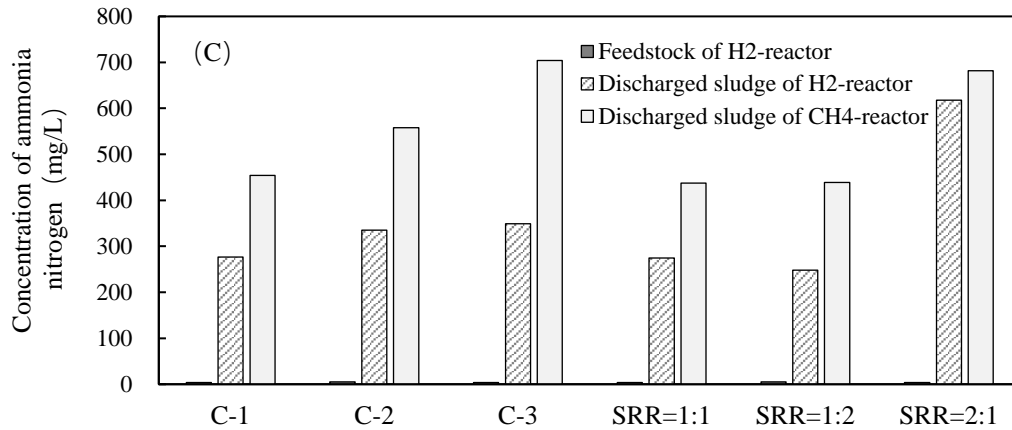
Generally, the hydrolysis of macromolecule organics in substrates were considered the rate-limiting step in anaerobic fermentation process. The protein can only be fermented by hydrogen-producing bacteria when it is hydrolysed into amino acids and the utilization depends on its hydrolysis degree[18]. Removal efficiency of soluble proteins in the two-stage system were 40.4%-61.4%, proved that the carbohydrates were more easily biodegraded substrates for anaerobic microorganisms. The concentration of soluble proteins in discharged sludge was much higher than that in feedstock of  $H_2$ -reactor, whereas an opposite trend was observed in  $CH_4$ -reactor. It is obvious that dissolution rate of proteins from solid phase was much higher than its degradation rate in acidogenic stage. The increasing concentration of soluble proteins in discharged sludge of  $H_2$ -reactor with higher M-SRT suggested worse degradation or higher dissolution of soluble proteins. Then the microorganisms in  $CH_4$ -reactor broke down the residual proteins which cannot be consumed in

acidogenic stage[19]. Contrary to soluble carbohydrates, the utilization of soluble proteins might be promoted in methanogenic stage rather than acidogenic stage. There was no significant change in removal efficiency of soluble proteins in CH<sub>4</sub>-reactor with M-SRT of 8 d and 12 d, suggested a limited ability to degrade soluble proteins in CH<sub>4</sub>-reactor. In the SRR optimization experiments, the highest removal efficiency of soluble proteins in CH<sub>4</sub>-reactor was obtained with SRR of 1:1. As mentioned above, high SRR of 2:1 caused stage transfer and thus increased the removal efficiency of soluble proteins in H<sub>2</sub>-reactor. Therefore, moderate increase of SRR might promote hydrolysis of proteins in CH<sub>4</sub>-reactor. But too high SRR would inhibit it and degradation of proteins was inhibited in the two-stage system eventually.

During anaerobic digestion, some organic nitrogen (mainly protein and urea) is converted to ammonia[20]. Ammonia with a proper concentration provides well buffer capacity of methanogenic medium, while high concentration of ammonia may inhibit the activity of microbial and thus disturb the anaerobic process performance[21]. Synopsis of various studies on ammonia inhibition during anaerobic digestion of organic feedstock showed that the inhibition threshold of total ammonia nitrogen (TAN) ranged from 800 mg/L to 14000 mg/L[22]. Therefore, the concentration of ammonia nitrogen was below the inhibition threshold in this study. The higher concentration of ammonia nitrogen in CH<sub>4</sub>-reactor with higher M-SRT may be partly influenced by the accumulation of ammonia nitrogen during the whole operating process. Meanwhile, the degradation rate of proteins in CH<sub>4</sub>-reactor increased with higher M-SRT and produced more ammonia nitrogen[23]. The concentration of ammonia nitrogen in the two-stage system accumulated obviously with SRR of 2:1. Meanwhile, the concentration of ammonia nitrogen in H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor were very close, indicating increased hydrolysis of proteins in H<sub>2</sub>-reactor when the acidogenic stage transferred to methanogenic stage.



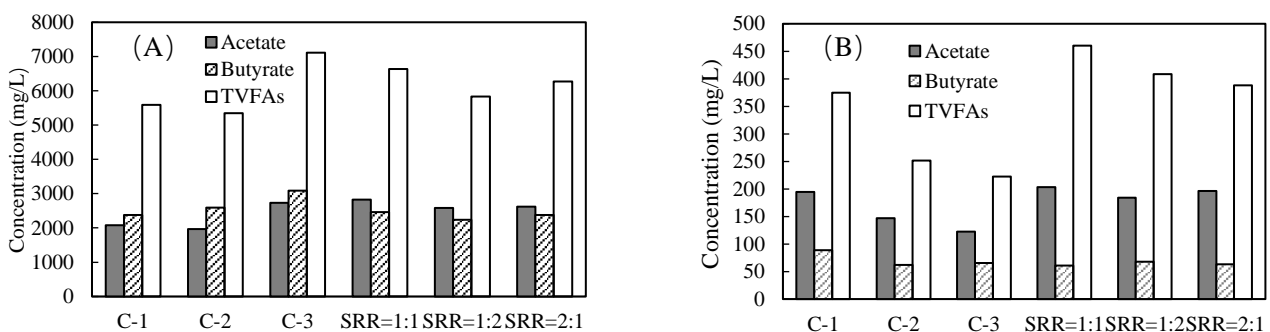




**Fig. 5.** Removal efficiency of soluble carbohydrates (A), concentration of soluble proteins (B) and ammonia nitrogen (C) under different operating conditions.

### 3.4.2 VFAs

The concentration of acetate, butyrate and TVFAs in sludge of H<sub>2</sub>-reactor and CH<sub>4</sub>-reactor under different operating conditions are shown in Fig.6. The main components of TVFAs in sludge of H<sub>2</sub>-reactor mainly include acetate and butyrate, implied that acetate or butyrate production pathway were dominant in H<sub>2</sub>-reactor instead of ethanol production pathway[24]. The degradation products of butyrate are acetate and hydrogen[25]. Compared with VFAs in H<sub>2</sub>-reactor under condition C-1, more butyrate and less acetate as well as improved CH<sub>4</sub> production under condition C-2 implied better metabolism of acetate to CH<sub>4</sub> than butyrate to acetate. The concentration of VFAs in H<sub>2</sub>-reactor grew higher under condition C-3. The existence of methanogens in acidogenic stage can enhance the acetate production by consuming H<sub>2</sub> without interfering with the acidogenesis process[12]. Considering worse stability of H<sub>2</sub> production, possibly it was hydrogenotrophic methanogenesis that consumed hydrogen and enhanced acidogenic metabolic pathways. This also caused lower pH and more alkali dosage. The accumulation of both VFAs and alkalinity affected stability of hydrogen producing stage. Most of the VFAs were removed in CH<sub>4</sub>-reactor. With higher M-SRT, the VFAs were degraded better and CH<sub>4</sub> production increased correspondingly. Within a certain range, the operation of the two-stage system was enhanced by raising M-SRT. However, higher M-SRT might influence the balance between bacteria and methanogens, and thus reduce the stability of H<sub>2</sub> production. The VS of feedstock in H<sub>2</sub>-reactor decreased with lower SRR, led to lower concentration of VFA in H<sub>2</sub>-reactor when SRR decreased from 1:1 to 1:2. With SRR of 2:1, the stage transfer in H<sub>2</sub>-reactor promoted degradation of VFAs by methanogens and thus the concentration of VFA was lower than that with SRR of 1:1.



**Fig. 6.** Concentration of acetate, butyrate and TVFAs of hydrogen (A) and methane (B) producing stage under different operating conditions.

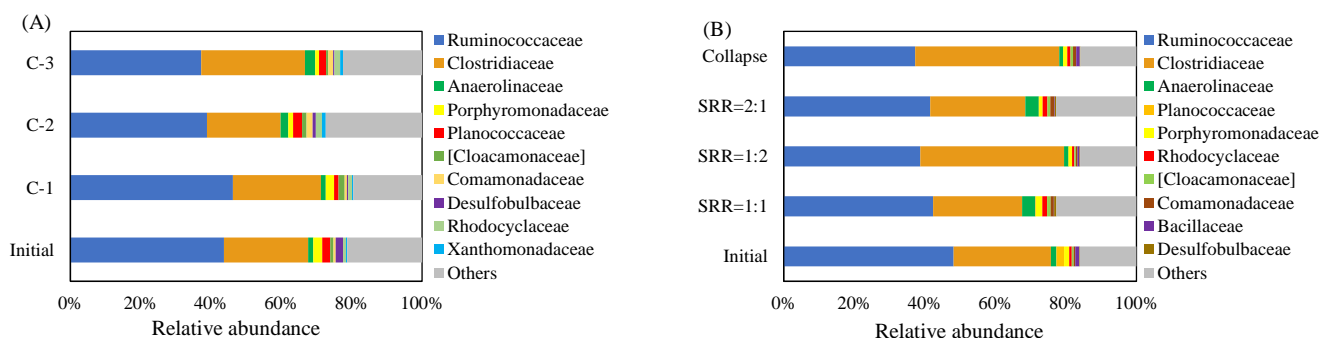
### 3.5 Microbial analysis

Relative sequence abundances of discharged sludge samples from H<sub>2</sub>-reactor at the family level under different operating conditions were shown in Fig. 7. In all the periods, the dominant bacterial families were *Ruminococcaceae* and *Clostridiaceae*, both of which are *Clostridia*. Clostridia usually involve in anaerobic digestion process and perform different essential metabolic pathways, including substrate hydrolysis, fermentation and acetogenesis. They conduct macromolecules hydrolysis[26], such as polysaccharides and oligosaccharides[27], cellulolytic[28] and proteins[29].

Some species of Clostridia also grow in co-culture with methanogens[30] and conduct syntrophic acetate oxidation coupled with hydrogenotrophic methanogenesis[31].

Some microorganisms in the family *Ruminococcaceae* are cellulose degrader able to use various sugars, including xylose as a source of carbon[32,33]. *Ruminococcaceae* are also positively correlated with VFAs and metabolism of amino acid and glycan[34]. In particular, *Ruminococcus albus* 7 has played a key role in the development of the concept of interspecies hydrogen transfer[35]. The relative abundance of *Ruminococcaceae* were relatively high with higher SRR, indicated the syntrophic association with methanogens. It is suggested that *Ruminococcaceae* are mainly involved in acetate and lactate production/consumption but indirectly contributed to butyrate production via the generation of acetate, lactate, and succinate for interconversion reactions[36].

Studies show that *Clostridiaceae* are related to butyrate[37] and H<sub>2</sub> production[38,39], such as produce H<sub>2</sub> from sugars[40] and lactate[41]. Some species of *Clostridiaceae* are also related to hydrolysis of keratin[42]. The lower relative abundance of *Ruminococcaceae* with higher SRR seemed to suggest the relation between those bacteria and hydrogen production.



**Fig. 7.** Relative sequence abundances of discharged sludge samples from H<sub>2</sub>-reactor at the family level under different M-SRT (A) and SRR (B).

## CONCLUSIONS

In this study, a temperature-separated two-stage anaerobic fermentation process of waste activated sludge and food waste was performed to optimize the M-SRT and SRR. Though the stability of H<sub>2</sub>-reactor was affected by higher M-SRT, the removal of organics and methane production was promoted for the two-stage system. Proper SRR could complement alkalinity for acidogenic stage while excessive SRR might cause stage transfer and thus inhibit the running of whole system. Therefore, the optimized M-SRT and SRR in this study were 12 d and 1:1 respectively.

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## REFERENCES

- Cheng, J., Ding, L.K., Lin, R.C., Yue, L.C., Liu, J.Z., Zhou, J.H., Cen, K.F.: Fermentative biohydrogen and biomethane co-production from mixture of food waste and sewage sludge: Effects of physiochemical properties and mix ratios on fermentation performance. *Appl. Energy* **184**, 1-8 (2016). doi:10.1016/j.apenergy.2016.10.003
- Ren, Y.Y., Yu, M., Wu, C.F., Wang, Q.H., Gao, M., Huang, Q.Q., Liu, Y.: A comprehensive review on food waste anaerobic digestion: Research updates and tendencies. *Bioresour. Technol.* **247**, 1069-1076 (2018). doi:10.1016/j.biortech.2017.09.109
- Li, Y.Y., Jin, Y.Y., Borrion, A., Li, H.L.: Current status of food waste generation and management in China. *Bioresour. Technol.* **273**, 654-665 (2019). doi:10.1016/j.biortech.2018.10.083
- Chen, X., Yuan, H.R., Zou, D.X., Liu, Y.P., Zhu, B.N., Chufo, A., Jaffar, M., Li, X.J.: Improving biomethane yield by controlling fermentation type of acidogenic phase in two-phase anaerobic co-digestion of food waste and rice straw. *Chem. Eng. J.* **273**, 254-260 (2015). doi:10.1016/j.cej.2015.03.067
- Zhu, H.G., Beland, M.: Evaluation of alternative methods of preparing hydrogen producing seeds from digested wastewater sludge. *Int. J. Hydrog. Energy* **31**(14), 1980-1988 (2006). doi:10.1016/j.ijhydene.2006.01.019
- Liu, X.Y., Li, R.Y., Ji, M., Han, L.: Hydrogen and methane production by co-digestion of waste activated sludge and food waste in the two-stage fermentation process: Substrate conversion and energy yield. *Bioresour. Technol.* **146**, 317-323 (2013). doi:10.1016/j.biortech.2013.07.096
- Gaudy, A.F.: *Colorimetric Determination of Protein and Carbohydrate*, vol. 7. (1962)
- Lowry, O.H., Rosebrough, N.J., Farr, A.L., Randall, R.J.: Protein measurement with the Folin phenol reagent. *The*

Journal of biological chemistry **193**(1), 265-275 (1951).

9. Bryant, M.P., Wolin, E.A., Wolin, M.J., Wolfe, R.S.: *Methanobacillus omelianskii*, a symbiotic association of two species of bacteria. *Archiv fur Mikrobiologie* **59**(1), 20-31 (1967). doi:10.1007/bf00406313
10. Shen, L., Zhao, Q.C., Wu, X.E., Li, X.Z., Li, Q.B., Wang, Y.P.: Interspecies electron transfer in syntrophic methanogenic consortia: From cultures to bioreactors. *Renew. Sust. Energ. Rev.* **54**, 1358-1367 (2016). doi:10.1016/j.rser.2015.10.102
11. Kraemer, J.T., Bagley, D.M.: Supersaturation of dissolved H<sub>2</sub> and CO<sub>2</sub> during fermentative hydrogen production with N<sub>2</sub> sparging. *Biotechnol. Lett.* **28**(18), 1485-1491 (2006). doi:10.1007/s10529-006-9114-7
12. Huang, W.H., Wang, Z.Y., Zhou, Y., Ng, W.J.: The role of hydrogenotrophic methanogens in an acidogenic reactor. *Chemosphere* **140**, 40-46 (2015). doi:10.1016/j.chemosphere.2014.10.047
13. DiStefano, T.D., Palomar, A.: Effect of anaerobic reactor process configuration on useful energy production. *Water Research* **44**(8), 2583-2591 (2010). doi:10.1016/j.watres.2010.01.010
14. Lay, J.J., Fan, K.S., Chang, J., Ku, C.H.: Influence of chemical nature of organic wastes on their conversion to hydrogen by heat-shock digested sludge. *Int. J. Hydrog. Energy* **28**(12), 1361-1367 (2003). doi:10.1016/s0360-3199(03)00027-2
15. Hwang, J.H., Choi, J.A., Abou-Shanab, R.A.I., Bhatnagar, A., Min, B., Song, H., Kumar, E., Choi, J., Lee, E.S., Kim, Y.J., Um, S., Lee, D.S., Jeon, B.H.: Effect of pH and sulfate concentration on hydrogen production using anaerobic mixed microflora. *Int. J. Hydrog. Energy* **34**(24), 9702-9710 (2009). doi:10.1016/j.ijhydene.2009.10.022
16. Leung, D.Y.C., Wang, J.: An overview on biogas generation from anaerobic digestion of food waste. *Int. J. Green Energy* **13**(2), 119-131 (2016). doi:10.1080/15435075.2014.909355
17. Lin, Y.Q., Wu, S.B., Wang, D.H.: Hydrogen-methane production from pulp & paper sludge and food waste by mesophilic-thermophilic anaerobic co-digestion. *Int. J. Hydrog. Energy* **38**(35), 15055-15062 (2013). doi:10.1016/j.ijhydene.2012.01.051
18. Spellman, D., McEvoy, E., O'Cuinn, G., FitzGerald, R.J.: Proteinase and exopeptidase hydrolysis of whey protein: Comparison of the TNBS, OPA and pH stat methods for quantification of degree of hydrolysis. *Int. Dairy J.* **13**(6), 447-453 (2003). doi:10.1016/s0958-6946(03)00053-0
19. Song, W.L., Cheng, J., Zhou, J.H., Xie, B.F., Su, H.B., Cen, K.F.: Cogeneration of hydrogen and methane from protein-mixed food waste by two-phase anaerobic process. *Int. J. Hydrog. Energy* **35**(7), 3141-3146 (2010). doi:10.1016/j.ijhydene.2009.09.102
20. Gonzalez-Fernandez, C., Garcia-Encina, P.A.: Impact of substrate to inoculum ratio in anaerobic digestion of swine slurry. *Biomass Bioenerg.* **33**(8), 1065-1069 (2009). doi:10.1016/j.biombioe.2009.03.008
21. Zhang, Y., Zamudio Canas, E.M., Zhu, Z.W., Linville, J.L., Chen, S., He, Q.: Robustness of archaeal populations in anaerobic co-digestion of dairy and poultry wastes. *Bioresour. Technol.* **102**(2), 779-785 (2011). doi:10.1016/j.biortech.2010.08.104
22. Rajagopal, R., Masse, D.I., Singh, G.: A critical review on inhibition of anaerobic digestion process by excess ammonia. *Bioresour. Technol.* **143**, 632-641 (2013). doi:10.1016/j.biortech.2013.06.030
23. Ding, H.B., Liu, X.Y., Stabnikova, O., Wang, J.Y.: Effect of protein on biohydrogen production from starch of food waste. *Water Sci. Technol.* **57**(7), 1031-1036 (2008). doi:10.2166/wst.2008.080
24. Lay, C.H., Wu, J.H., Hsiao, C.L., Chang, J.J., Chen, C.C., Lin, C.Y.: Biohydrogen production from soluble condensed molasses fermentation using anaerobic fermentation. *Int. J. Hydrog. Energy* **35**(24), 13445-13451 (2010). doi:10.1016/j.ijhydene.2009.11.128
25. Yang, Y., Chen, Q., Guo, J.L., Hu, Z.Q.: Kinetics and methane gas yields of selected C1 to C5 organic acids in anaerobic digestion. *Water Res.* **87**, 112-118 (2015). doi:10.1016/j.watres.2015.09.012
26. Tracy, B.P., Jones, S.W., Fast, A.G., Indurthi, D.C., Papoutsakis, E.T.: Clostridia: the importance of their exceptional substrate and metabolite diversity for biofuel and biorefinery applications. *Curr. Opin. Biotechnol.* **23**(3), 364-381 (2012). doi:10.1016/j.copbio.2011.10.008
27. Krause, L., Diaz, N.N., Edwards, R.A., Gartemann, K.H., Kromeke, H., Neuweger, H., Puhler, A., Runte, K.J., Schluter, A., Stoye, J., Szczepanowski, R., Tauch, A., Goesmann, A.: Taxonomic composition and gene content of a methane-producing microbial community isolated from a biogas reactor. *J. Biotechnol.* **136**(1-2), 91-101 (2008). doi:10.1016/j.jbiotec.2008.06.003
28. Tomazetto, G., Hahnke, S., Koeck, D.E., Wibberg, D., Maus, I., Puhler, A., Klocke, M., Schluter, A.: Complete genome analysis of *Clostridium bornimense* strain M2/40(T): A new acidogenic *Clostridium* species isolated from a mesophilic two-phase laboratory-scale biogas reactor. *J. Biotechnol.* **232**, 38-49 (2016). doi:10.1016/j.jbiotec.2015.08.001
29. Pagliano, G., Vantorino, V., Panico, A., Pepe, O.: Integrated systems for biopolymers and bioenergy production from organic waste and by-products: a review of microbial processes. *Biotechnol. Biofuels* **10**, 24 (2017). doi:10.1186/s13068-017-0802-4
30. McInerney, M.J., Bryant, M.P., Hespell, R.B., Costerton, J.W.: *Syntrophomonas wolfei* gen. nov. sp. nov., an Anaerobic, Syntrophic, Fatty Acid-Oxidizing Bacterium. *Applied and environmental microbiology* **41**(4),

1029-1039 (1981).

31. Mosbaek, F., Kjeldal, H., Mulat, D.G., Albertsen, M., Ward, A.J., Feilberg, A., Nielsen, J.L.: Identification of syntrophic acetate-oxidizing bacteria in anaerobic digesters by combined protein-based stable isotope probing and metagenomics. *ISME J.* **10**(10), 2405-2418 (2016). doi:10.1038/ismej.2016.39
32. Weiss, S., Zankel, A., Leubhn, M., Petrak, S., Somitsch, W., Guebitz, G.M.: Investigation of microorganisms colonising activated zeolites during anaerobic biogas production from grass silage. *Bioresour. Technol.* **102**(6), 4353-4359 (2011). doi:10.1016/j.biortech.2010.12.076
33. Jia, Y.Y., Wilkins, D., Lu, H.Y., Cai, M.W., Lee, P.K.H.: Long-Term Enrichment on Cellulose or Xylan Causes Functional and Taxonomic Convergence of Microbial Communities from Anaerobic Digesters. *Applied and Environmental Microbiology* **82**(5), 1519-1529 (2016). doi:10.1128/aem.03360-15
34. Zhang, L., Wu, W.D., Lee, Y.K., Xie, J.J., Zhang, H.F.: Spatial Heterogeneity and Co-occurrence of Mucosal and Luminal Microbiome across Swine Intestinal Tract. *Front. Microbiol.* **9**, 14 (2018). doi:10.3389/fmicb.2018.00048
35. Zheng, Y.N., Kahnt, J., Kwon, I.H., Mackie, R.I., Thauer, R.K.: Hydrogen Formation and Its Regulation in *Ruminococcus albus*: Involvement of an Electron-Bifurcating FeFe -Hydrogenase, of a Non-Electron-Bifurcating FeFe -Hydrogenase, and of a Putative Hydrogen-Sensing FeFe -Hydrogenase. *J. Bacteriol.* **196**(22), 3840-3852 (2014). doi:10.1128/jb.02070-14
36. Esquivel-Elizondo, S., Ilhan, Z.E., Garcia-Pena, E.I., Krajmalnik-Brown, R.: Insights into Butyrate Production in a Controlled Fermentation System via Gene Predictions. *mSystems* **2**(4) (2017). doi:10.1128/mSystems.00051-17
37. Ayudthaya, S.P.N., van de Weijer, A.H.P., van Gelder, A.H., Stams, A.J.M., de Vos, W.M., Plugge, C.M.: Organic acid production from potato starch waste fermentation by rumen microbial communities from Dutch and Thai dairy cows. *Biotechnol. Biofuels* **11**, 15 (2018). doi:10.1186/s13068-018-1012-4
38. Navarro-Diaz, M., Valdez-Vazquez, I., Escalante, A.E.: Ecological perspectives of hydrogen fermentation by microbial consortia: What we have learned and the way forward. *Int. J. Hydrog. Energy* **41**(39), 17297-17308 (2016). doi:10.1016/j.ijhydene.2016.08.027
39. Alexandropoulou, M., Antonopoulou, G., Trably, E., Carrere, H., Lyberatos, G.: Continuous biohydrogen production from a food industry waste: Influence of operational parameters and microbial community analysis. *J. Clean Prod.* **174**, 1054-1063 (2018). doi:10.1016/j.jclepro.2017.11.078
40. Jiang, L., Wu, Q., Xu, Q., Zhu, L.Y., Huang, H.: Fermentative hydrogen production from Jerusalem artichoke by *Clostridium tyrobutyricum* expressing exo-inulinase gene. *Sci Rep* **7**, 10 (2017). doi:10.1038/s41598-017-07207-7
41. Noblecourt, A., Christophe, G., Larroche, C., Fontanille, P.: Hydrogen production by dark fermentation from pre-fermented depackaging food wastes. *Bioresour. Technol.* **247**, 864-870 (2018). doi:10.1016/j.biortech.2017.09.199
42. Xia, Y., Wang, D.K., Kong, Y., Ungerfeld, E.M., Seviour, R., Masse, D.I.: Anaerobic digestibility of beef hooves with swine manure or slaughterhouse sludge. *Waste Manage.* **38**, 443-448 (2015). doi:10.1016/j.wasman.2014.12.017