

Biodegradation of Phenolic Effluent of Producer Gas Plant using *Scenedesmus* sp.

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Highlights

- Degradation of phenolic effluent using the microalgae *Scenedesmus* sp.
- *Scenedesmus* sp. degrades more than 90 % of phenol in the effluent.
- Degradation study was performed in the ambient conditions.

Abstract

Phenol is a pollutant, usually found in effluents produced from industries such as coal processing plants, oil refineries, pulp and paper manufacturing plants, resins and coke manufacturing, steel industries etc. During last two decades, extensive attention has been paid on the management of environmental pollution caused by hazardous materials. A number of methods have been developed for removal of such substances like precipitation, evaporation, ion-exchange etc. However these methods have several disadvantages. The present work highlights the alternative biological agent abundantly present in nature i.e microalgae (*Scenedesmus* sp.) as a potential biological agent for removal of such toxic substances from the surrounding. The fresh water microalgae *Scenedesmus* sp. was investigated for the ability to degrade the phenolic compounds present in the effluent of producer gas plant. The effluent has 1024 ppm of initial phenol concentration. The biodegradation study was performed in two batches (with and without nutrients) both having four different concentrations of *Scenedesmus* sp. mixed with effluent. It was found that wet biomass of *Scenedesmus* sp., with nutrients, was able to reduce the phenol concentration (C₆H₅OH) present in the effluent by more than 90% for algal concentration of 2000 ppm (mg of wet biomass/l of effluent) and above. While for samples without nutrients the percentage reduction of total phenol was observed upto a maximum of 46%.

Key words

Phenol, Phycoremediation, Wastewater treatment, Phenolic effluent, Biodegradation, *Scenedesmus* sp., Microalgae

1.0 Introduction

Phycoremediation is a biodegradation method in which, either the microalgae or macro algae are used as a biological agent. Phycoremediation offers an interesting step for wastewater treatments, because they provide a tertiary bio-treatment coupled with the production of potentially valuable biomass, which can be used for several purposes. While degradation of phenol and its derivatives by conventional methods, releases toxic byproducts, degradation by microalgae is advantageous, as it eliminates toxic byproduct formation and favors useful byproducts, biomass production and CO₂ absorption [7]. Among the various industrial sectors, the coal and the petroleum industries generate significant quantities of wastewater containing large amount of phenolic compounds which may have a detrimental impact when exposed into the environment without any prior treatment ([6],[9]). Phenol or its derivatives are toxic to aquatic life if its concentration exceeds 2 ppm. Therefore, its effective decomposition and analysis of byproducts formed are necessary

43 ([8], [13]). Algal strains can survive at higher concentration of phenol and its types [1]. Algae
44 degrade organic matter present in the effluent by oxidation using the oxygen produced via
45 photosynthesis. Algae utilize the released carbon dioxide and nutrients during aerobic oxidation
46 of organic matter for their growth. In comparison to common treatment systems, oxidation pond
47 supports the growth of some other species. High pH in algal ponds leads to pathogen disinfection
48 other than algal species, so the degradation of organic matter by the specific algal species is
49 achieved [4].

50 Growing algae in wastewater offers numerous economic and environmental merits, providing
51 one of the most sustainable ways to produce biodiesel derived from microalgae. Wastewater
52 usage eliminates competition for fresh water, saves cost of nutrients supplement since nutrients
53 are in abundance in wastewater, provides the treatment of the wastewater by assimilating organic
54 and inorganic pollutants into their cells, apart from eliminating the CO₂ emissions associated
55 with wastewater treatment [17]. Algal strains are capable of metabolizing phenol in the
56 environment. In the dynamic energy budget model proposed by Lika and Papadakis [14] for
57 aerobic degradation, inhibition of phenol reduction was reported due to the presence of glucose
58 (carbon source for algal growth).

59 *Chlorella sp.*, *Scenedesmus obliquus* and *Spirulina maxima* degraded variety of Phenol and its
60 types viz. 2,4-dimethylphenol, 2,4-dinitrophenol, 2,4-dichlorophenol, 2-chlorophenol, in which
61 phenol was completely degraded by all the algal cultures up to 1000mg/l. *Scenedesmus obliquus*
62 was able to degrade 2,4-dinitrophenol completely at a concentration of about 190 mg/l. *Chlorella*
63 sp. dechlorinate 2-chlorophenol at a maximum concentration of 200 mg/l [1]. Pinto et al.,
64 2002[4] reported that the two micro algae, *Ankistrodesmus braunii* and *Scenedesmus*
65 *quadricauda* degraded more than 70% of olive mill wastewater having 400mg/l of phenol within
66 5 days.

67 Degradation of phenol by prokaryotic and eukaryotic organisms requires the presence of
68 molecular oxygen to initiate enzymatic attack of the aromatic rings ([2] & [3]), hence the phenol
69 degradation by aerobic oxidation is one of the most suitable methods. A typical pathway for
70 metabolizing phenol is to hydroxylate the ring to form catechol and then to open the ring through
71 ortho-or meta-oxidation ([3]&[11]).

72 Semple et al., 1996 [2] reported that the *O.danica* is able to metabolize phenol completely; and
73 the algae accumulate the carbon in the nucleic acid and in the lipid content of the cells.

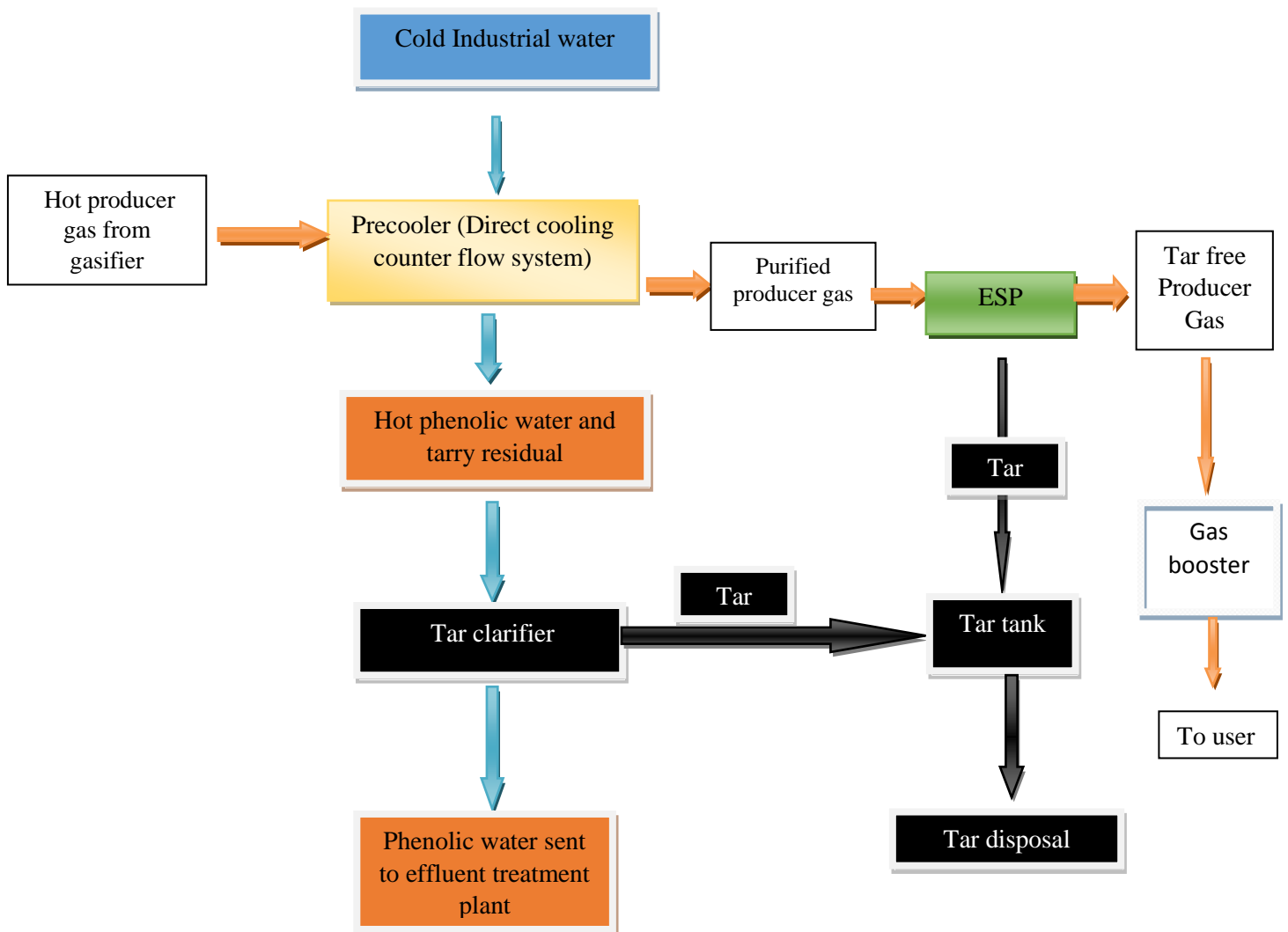
74 Many researchers demonstrated the biodegradation of synthetic phenol and its derivatives by
75 algae; only few researchers reported degradation experiments on real effluents. The present study
76 focuses on the degradation of phenol rich effluent of producer gas plant by the fresh water micro
77 algae *Scenedesmus sp.*

78 **2.0 Source of effluent**

79 The effluent was collected from a producer gas plant located nearby. Coal is gasified using air
80 and steam in this producer gas plant. The gasifiers used for the production of gas are of single
81 stage with rotating bottom parts, water cooled jacket, coal charging and ash removal system. The
82 water used for cooling and scrubbing the hot raw gas comes out from the bottom water seal of
83 the pre-cooler. The schematic flow diagram of producer gas plant is depicted in Fig 1. The water
84 released from the pre-cooler is contaminated with phenolic compounds. Due to stripping, various

85 phenolic compounds get dissolved in water, phenol being the major constituent. In addition, the
86 effluent contains large amount of oil and grease. Due to the possibilities of enormous hazardous
87 effects of phenolic compounds over the flora and fauna around, it becomes essential to
88 characterize and degrade the effluent discharged by the industry.

Fig 1- Flow diagram of existing producer gas plant



89 **3.0 Materials and Methods**

90 The chemicals required are laboratory grade Bromate- Bromide solution, Starch powder,
91 Potassium iodide, Sodium thiosulfate, Conc. HCl (M), Chloroform, Mercury (II) sulphate (in
92 Sulphuric acid), Potassium dichromate (in Sulphuric acid), Ethyl acetate, Methanol, Urea and
93 Potassium Bicarbonate (Nutrient source for the growth of *Scenedesmus* sp.)

94 **3.1 Culture of *Scenedesmus* sp.**

95 A culture of targeted algal species was procured from the Bioenergy lab, in the Department of
96 Energy and environment at the National Institute of Technology, Trichy, India. The *Scenedesmus*
97 sp. was cultivated in optimized culture medium comprising potassium bicarbonate and urea in
98 the ratio of 2:1. The culture condition was atmospheric as it was kept in a closed room with
99 sufficient openings via windows and doors for light and air. No artificial lamps were used.
100 Temperature range varies from 30 -35° C. The pH of the culture was found out to be 7-8. The
101 growth of microalgae was monitored by observing optical density at 600 nm and for the
102 identification of contamination, the algae was monitored daily in fluorescent inverted microscope
103 (Nikon DS-Fi2).

104 **3.2 Characterization of phenolic effluent**

105 The conventional parameters like pH, color, Chemical Oxygen Demand (COD), Biological
106 Oxygen Demand (BOD), Total Dissolved Solids (TDS) and Total Suspended Solids (TSS) were
107 analyzed by using standard procedures listed in the Table 1. The concentration of total phenol
108 present in effluent was analyzed by using Indian Standard, IS 3025 part no 43 method, the types
109 of phenolic compounds present in the effluent was analyzed using gas chromatography. Total
110 organic content present in the effluent was analyzed by TOC Analyzer (AJ- Analyzer multi N/C
111 3100; multiwin 4.09).

112 **3.3 Analysis of phenolic compounds through gas chromatography**

113 The types of phenolic compounds present in effluent were characterized by gas chromatography
114 (GC 450 Varian).

115 **3.3.1 Extraction Method**

116 The phenolic compounds were extracted from the effluent using ethyl acetate as a solvent, dried
117 over anhydrous sodium sulphate and analyzed by gas chromatography [16]. Extracted samples
118 were analyzed in GC 450 Varian with area normalization.

119 **3.3.2 Gas chromatography operating conditions**

120 Capillary column VF-1ms (15m x 0.25 mm) and 0.25 μ m, Injector: 250°C, injection volume: 1 μ l
121 (split), Flame Ionization Detector: 250°C, carrier gas: N₂, oven temperature: 100°C (hold 2 min.)
122 to 230°C at 20°C/ min (hold 35min.)[16].

123 To identify the types of phenol present in the effluent, 11 phenolic standards were selected based
124 on list enumerated by Environmental Protection Agency (EPA) [18]. The standard phenolic

125 compounds are Phenol, 2-Chlorophenol, 2-Methyl phenol, 2-Nitrophenol, 2,4-Dimethyl phenol,
126 2,4-Dichlorophenol, 4-Chloro-3-methyl phenol, 2,4,6-trichlorophenol, 2,4-Dinitrophenol 4-
127 Nitrophenol, Pentachlorophenol. Out of which phenol constitutes 50.4% and 2,4,6-
128 trichlorophenol constitutes 2.3%.

129 **3.4 Experimentation**

130 A 15 day old algal culture (*Scenedesmus* sp.) was centrifuged at 7000 g for 10 min to separate
131 the biomass without disturbing the cells. The cells were washed with distilled water to remove
132 the salts deposited on the cells. This step was repeated three times to accomplish complete
133 removal of salts. The ability of *Scenedesmus* sp. on phenol degradation was analyzed by
134 introducing four different concentration of algae viz. 1g, 2g, 3g and 4g per liter of raw effluent
135 (without prior treatment) with nutrients (Potassium bicarbonate: Urea in the ratio of 2:1) (Batch I)
136 and without nutrients (Batch II). The experiments were performed in 1000ml Erlenmayer flasks
137 with cotton plugs. The two batches along with algal biomass were maintained at ambient
138 condition same as maintained for the cultivation of *Scenedesmus* sp. over a period of 7 days. The
139 effluent devoid of algae with and without nutrients served as a control and maintained as the
140 same condition as maintained for the inoculum of *Scenedesmus* sp. The cultures were mixed
141 three times a day to accomplish complete mixing. The growth of the algal species was analyzed
142 by observing absorbance at 600 nm at UV-Vis spectrophotometer (Spectroquant pharo 300
143 IMERCK). Contamination and physical changes in the cells of *Scenedesmus* sp. due to the high
144 phenolic stress was observed daily using fluorescent inverted microscope. The pH of the samples
145 was analyzed by digital pH meter (Duralab).

146 **3.5 Analyses**

147 For the analysis of phenol, *Scenedesmus* sp. was separated from the culture broth through
148 centrifugation at 10,000 g for 5 min. The supernatant was filtered through a 0.45 µm PTFE
149 membrane filter and the filtrate was further analyzed in GC for identifying the phenolic
150 compounds using the same method followed for the effluent characterization. The pellet was
151 extracted with 1 mL ethyl acetate and then centrifuged at 10,000 g for 5 min. The ethyl acetate
152 extract was filtered through a 0.45 µm PTFE filter and this was used for GC analysis in order to
153 check for any possible phenol adsorbed. Analysis was performed using the same operating
154 condition adopted for effluent analysis. GC analysis were performed using capillary column Vf-
155 1ms, 15m, 0.25mm, 0.25µm, (Varian 450 GC), Detection was achieved by Flame Ionization
156 Detector. Variation in Total Organic Content (TOC) was analyzed by using TOC analyzer
157 (Analytikgena (multi N/C) 3100).

158 **4.0 Results and Discussion**

159 **4.1 Initial characteristics of raw effluent**

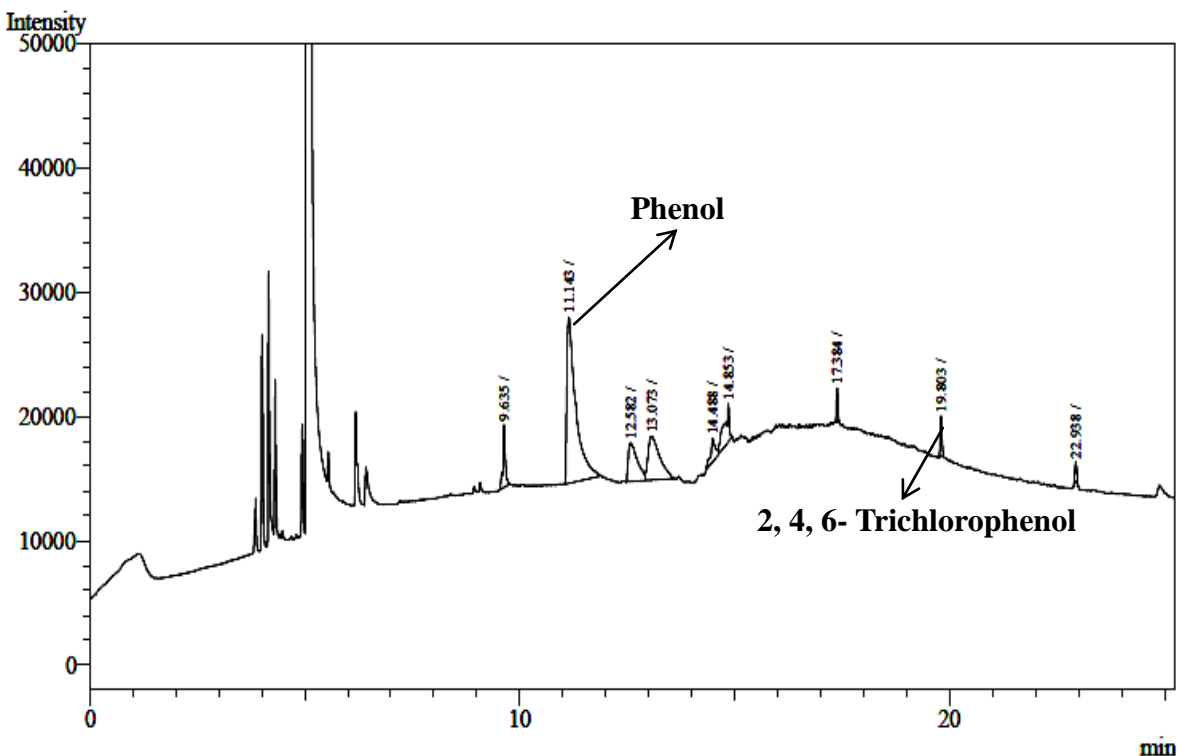
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161 Conventional parameters like Colour, pH, Chemical Oxygen Demand (COD), Biological
162 Demand (BOD), Total dissolved solids (TDS), Total Suspended Solids (TSS), Total Phenol
163 concentration were analyzed using the standard procedure and the results were tabulated in Table
164 2. The GC Chromatogram for the raw effluent (Fig 2) showed that the raw effluent before
165 treatment contains 9 different compounds. The results were compared with GC chromatogram
166 for the known standard phenolic compounds. Out of the different compounds, Phenol (C₆H₅OH)

167 and 2,4,6-Trichlorophenol were confirmed by having the same retention times at the same
168 operating conditions. According to the results, the effluent contains only two phenolic
169 compounds listed by EPA. Phenol (C_6H_5OH) alone comprises more than 50% of raw effluent.

170

171 **Fig-2 Characterization of types phenol present in the effluent using gas chromatography**

172



173

174

175 4.2 Effect of phenol on the growth of microalgae

176

177 Growth of *Scenedesmus* sp. was monitored daily by analyzing optical density at 600 nm using
178 UV –Vis spectrophotometer (Spectroquant) (Fig- 3,4,5,6) The increase in optical density was
179 observed in both the batches. The increase in optical density is the measure of identification of
180 algal growth. Increased growth rate for the *Scenedesmus* sp. was observed in batch I compared to
181 batch-II. Optical density was gradually increased in all the samples of batch-I. Sudden lag in
182 growth curve was observed in batch-I (with nutrient) due to the removal pollutant by the
183 biological agent, which is involved in the degradation [10]. In case of batch-II (without nutrients)
184 increased lag phase was observed because the *Scenedesmus* sp. requires adaptation period when
185 it is exposed to high phenolic environment as reported in [1]. Increase in optical density was only
186 observed after 2 days for batch II except for the sample having algal concentration of 4000 ppm.

Fig 3-6 Effect of phenolic effluent on the growth of *Scenedesmus* sp. (wet biomass of algae were inoculated at various concentration)

Fig 3 Growth of *Scenedesmus* sp. in 1000 PPM (1g of algae/1l of effluent)

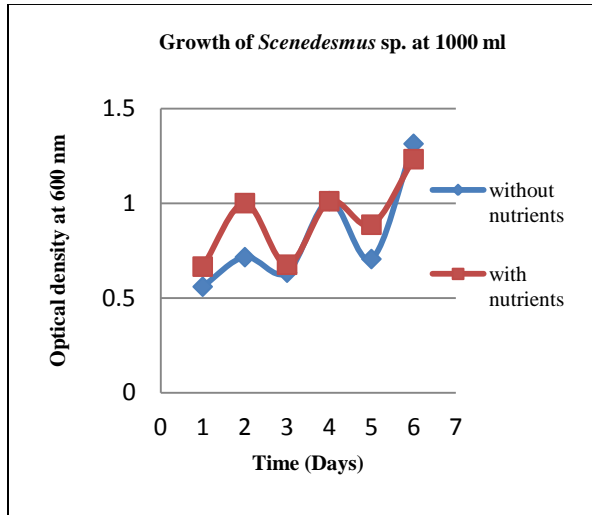


Fig 5 Growth of *Scenedesmus* sp. in 3000 ppm (2g of algae/1l of effluent)

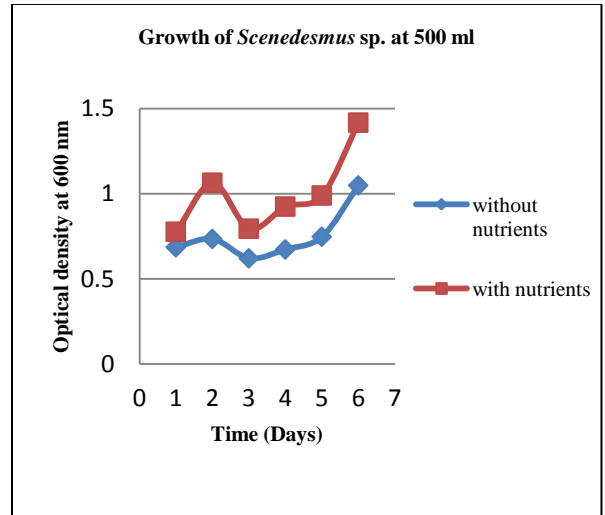


Fig 4 Growth of *Scenedesmus* sp. in 2000ppm (2g of algae/1l of effluent)

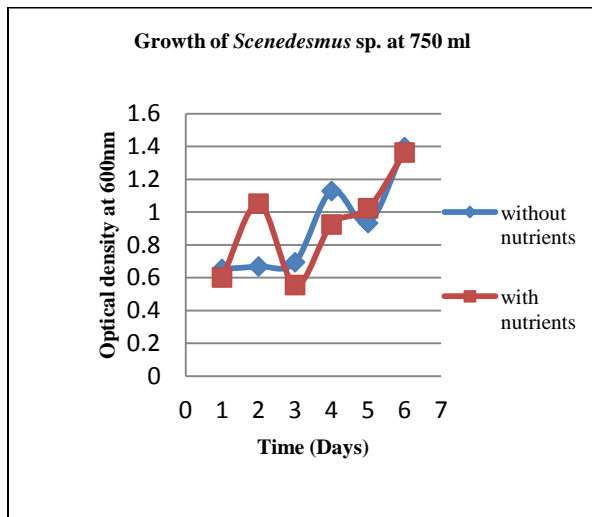
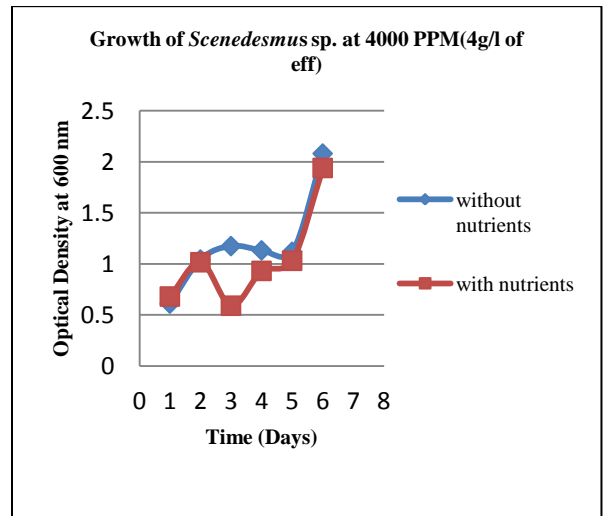


Fig 6 Growth of *Scenedesmus* sp. in 4000 ppm (4g of algae/1l of effluent)



187 4.3 Effect of potassium bicarbonate and urea on the removal of phenol

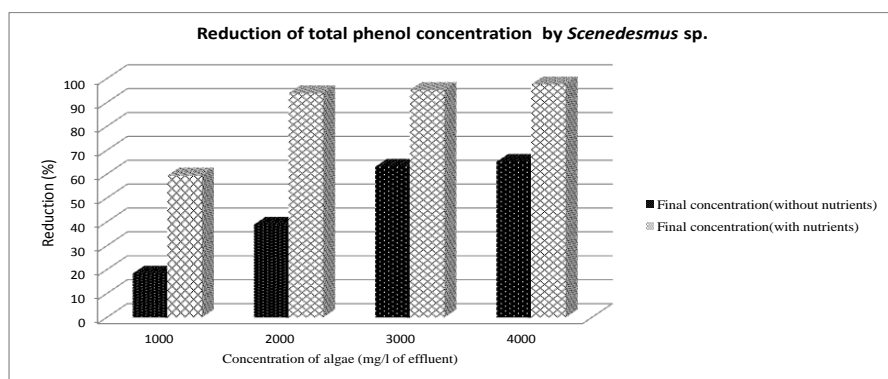
188
189 The effect of media components used in this study increases the growth of microalgae and also
190 increases the removal of phenolic content present in the effluent by the microalgae. Petroustos et
191 al., 2007[12] reported that the increase in initial concentration of NaHCO_3 , leads to increased
192 biomass production and p-Chlorophenol removal by the microalga *Tetraselmis marina*. The
193 meagre reduction of inorganic carbon observed through TOC analysis, reveals that the added
194 salts are not involved in chemical degradation of phenol.

196 4.4 Removal of phenol by *Scenedesmus* sp.

197
198 In order to analyze the ability of phenol removal by the *Scenedesmus* sp., the algae was grown in
199 two different batches. Each batch has four different concentration of algae viz. 1000, 2000, 3000
200 and 4000 ppm (mg of wet biomass/ liter of effluent). Potassium bicarbonate and urea were used
201 as nutrient source in batch-I. All the experiments were conducted three times and the results were
202 averaged. The algal cells were separated from the effluent by centrifuging the sample at 10000 g
203 for 10 min and the phenol was extracted by the same method that was used for raw effluent
204 characterization. And the total phenol concentration was analyzed using the IS 3025 part no 43
205 method and for phenol ($\text{C}_6\text{H}_5\text{OH}$) gas chromatography was used. The *Scenedesmus* sp. shows
206 high phenol removal in batch-I having concentration of algae 4000 ppm (mg of wet biomass/l of
207 effluent). No bioaccumulated (intracellular or cell surface adsorbed) phenol was detected with
208 GC analysis of the micro algal pellet ethyl acetate extract. The degradation ability of the
209 *Scenedesmus* sp. was comparatively identified by analyzing the batch without medium.

210 Reduction of total phenol observed using the IS 3025 part no 43 method was high at 4000 ppm
211 (4g of wet biomass/ 1 liter of effluent) in batch-I and average of 87 % reduction was
212 observed(Fig-7)) within 7 days and for without nutrients average of 46% reduction.

Fig-7 Comparison of concentration of total phenol presents in the effluent after treatment with nutrients and without nutrients samples



213
214 The reduction of phenol ($\text{C}_6\text{H}_5\text{OH}$) was determined by gas chromatography (GC 450 Varian)
215 analysis with area normalization method (Fig 8-11 (batch-I)). The reduction of phenol ($\text{C}_6\text{H}_5\text{OH}$)

216 by the *Scenedesmus* sp. was high at 99 % at the algal concentration of 4000 ppm (mg/l of
217 effluent) with nutrient batch (Fig 15). From the results, the degradation of phenol should be
218 accomplished by the oxidation of phenol by the oxygen evolved during degradation. No change
219 in the concentration of phenol was observed in the controls.

Fig (8-11) Gas chromatographic analysis of samples in batch I

Fig-8 Gas Chromatogram of effluent having the concentration of 1000 ppm of algae with nutrients

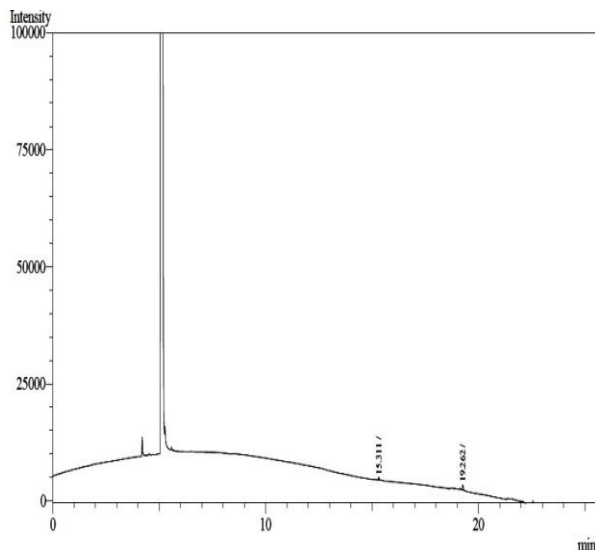


Fig-9 Gas Chromatogram of effluent having the concentration of 2000 ppm of algae with nutrients

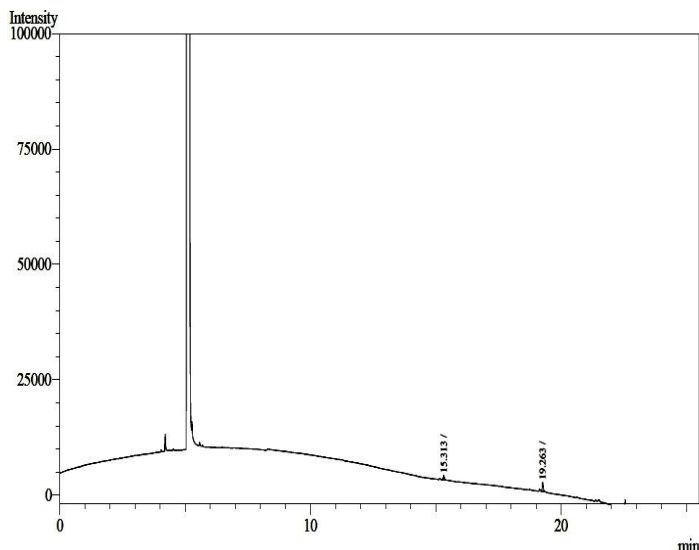


Fig-10 Gas Chromatogram of effluent having the concentration of 3000 ppm of algae with nutrients

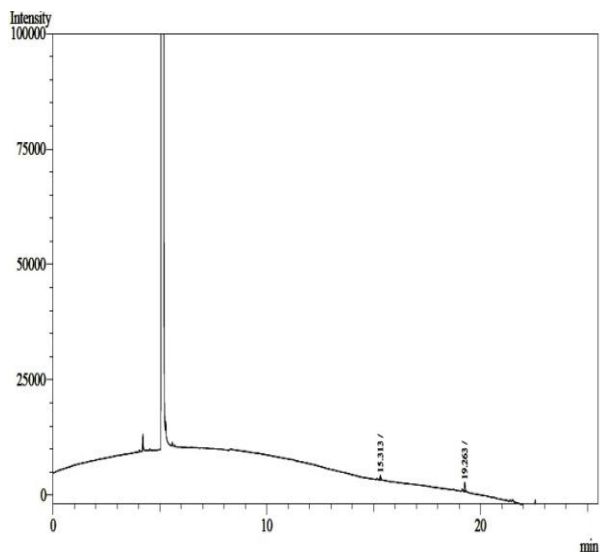


Fig-11 Gas Chromatogram of effluent having the concentration of 4000 ppm of algae with nutrients

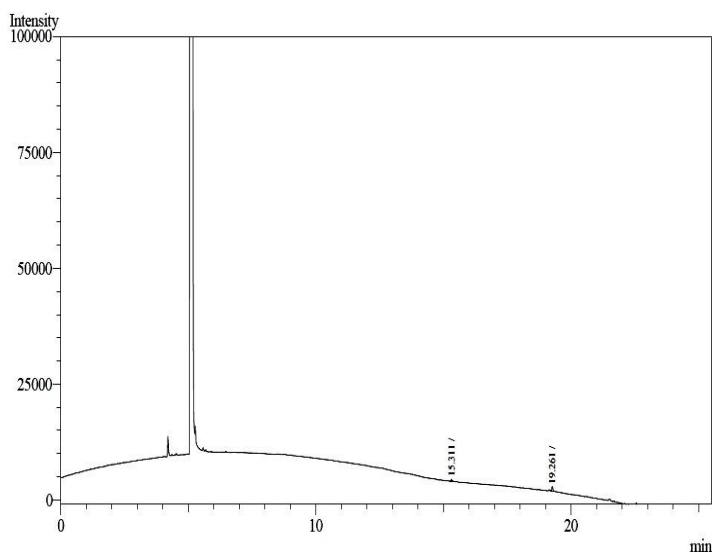


Fig (12-15) Gas chromatographic analysis of samples in batch II

Fig-12 Gas Chromatogram of effluent having the algal concentration of 1000(ppm) without nutrients

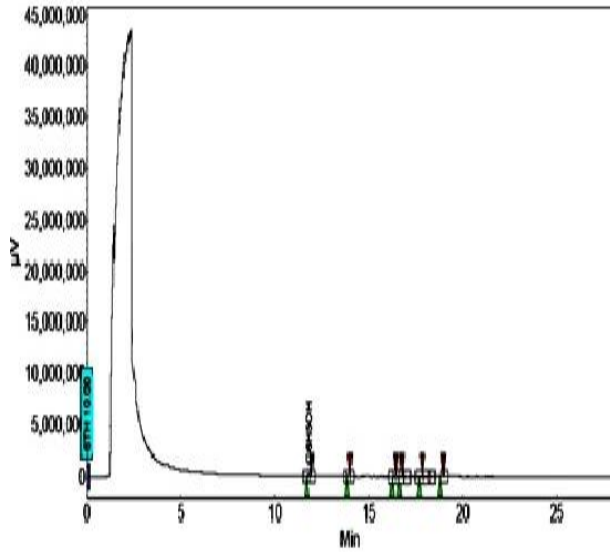


Fig-14 Gas chromatogram of effluent having the algal concentration of 3000(ppm) without nutrients

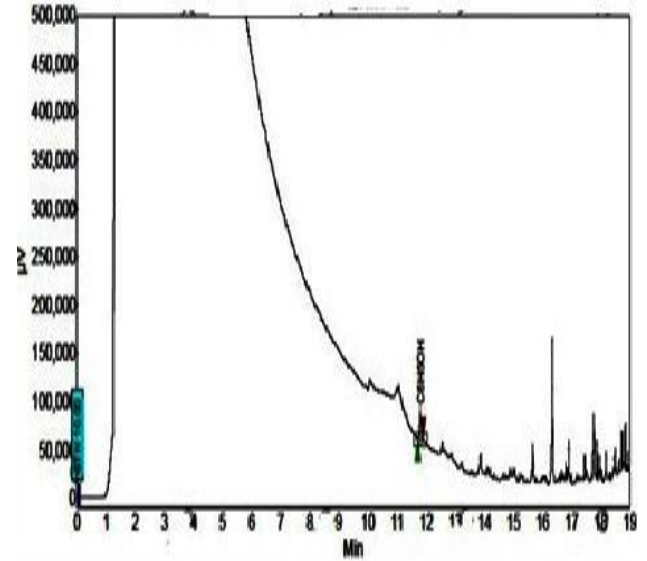


Fig-13 Gas Chromatogram of effluent having the algal concentration of 2000(ppm) without nutrients

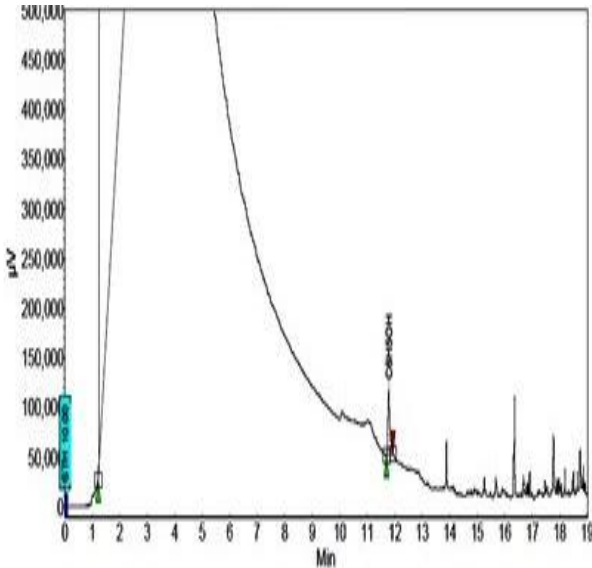
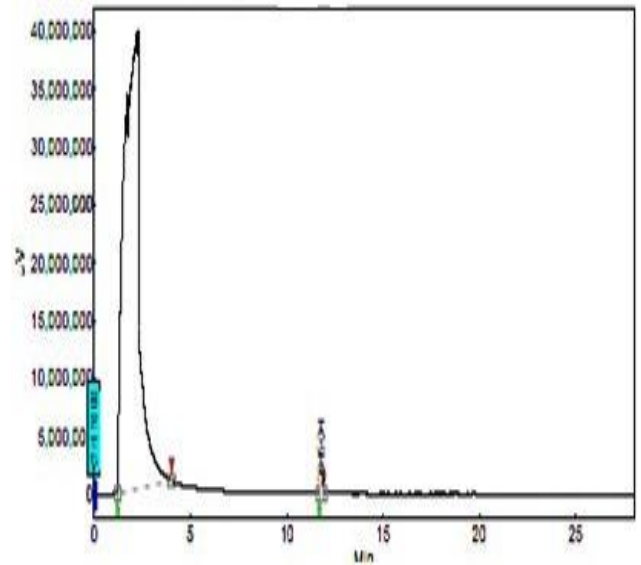


Fig-15 Gas Chromatogram of effluent having the algal concentration of 4000(ppm) without nutrients



220 The reduction of phenol (C₆H₅OH) by *Scenedesmus* sp. in the batch II also shows the same
 221 degradation observed in the batch I. Disappearance of peak corresponds to the phenol (C₆H₅OH)
 222 was observed in all samples in batch I. Interestingly in batch I the microalgae *Scenedesmus* sp.
 223 degrades the phenol (C₆H₅OH) as well as other contaminants present in the effluent. The gas
 224 chromatogram results of batch II showed that the reduction of phenol was less comparatively
 225 than batch I.

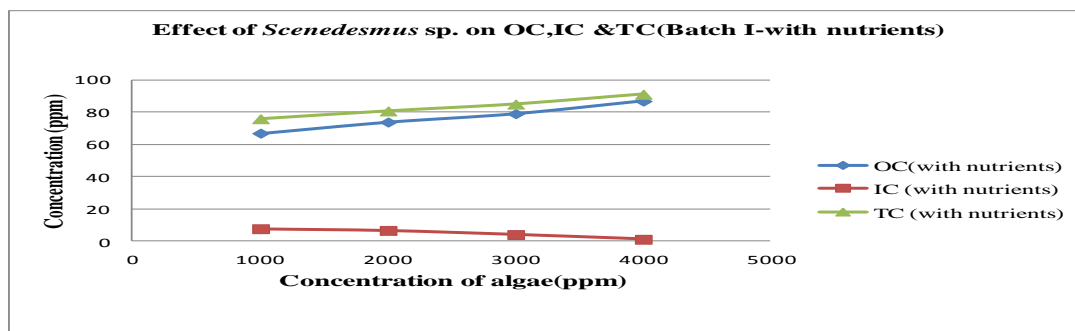
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227 **Effect of *Scenedesmus* sp. on total carbon (Organic (OC), Inorganic (IC) and Total (TC))**

228

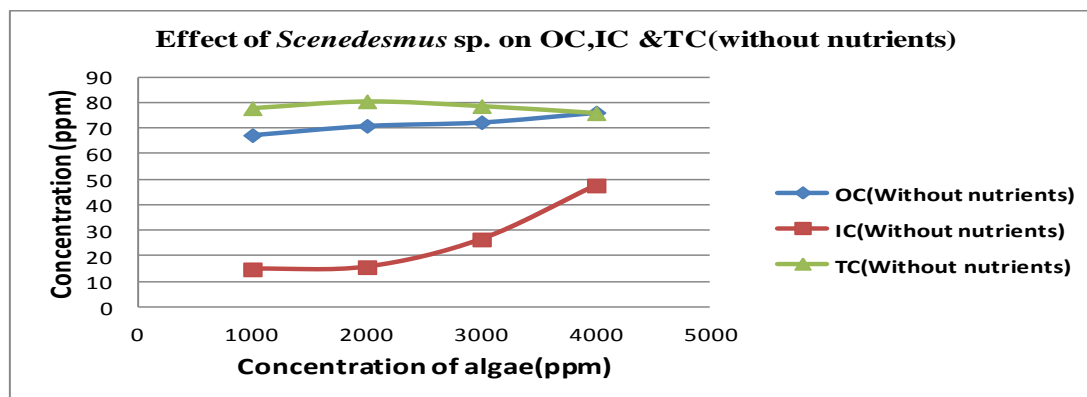
229 The variation in total organic content present in the effluent was analyzed using TOC Analyzer
 230 (Fig 16 & 17), in which the effluent with nutrients devoid of algae and the effluent without
 231 nutrients devoid of algae served as control to monitor the actual reduction of carbon content
 232 present in the effluent by the microalgae *Scenedesmus* sp. The reduction of organic and inorganic
 233 carbon has linear relationship with the concentration of algae in batch-I. The results showed that
 234 the *Scenedesmus* sp. was able to reduce the organic carbon content present in the effluent to a
 235 maximum of 91 % in batch- I(Fig-16) having 4000 ppm of algal concentration (mg of wet
 236 biomass/l of effluent) and the 80% in batch-II(Fig-17) having algal concentration of 4000 ppm
 237 within 7 days. No change the concentration of carbon (OC, IC and TC) was observed in the
 238 controls.

239 **Fig-16 Effect of *Scenedesmus* sp. on the carbon content (Organic Carbon (OC), Inorganic
 240 carbon(IC), Total carbon (TC)) of batch-I**



241

242 **Fig-17 Effect of *Scenedesmus* sp. on the carbon content (Organic Carbon (TOC), Inorganic
 243 carbon(IC), Total Carbon (TC)) of batch-II**



244

245 The reduction of inorganic carbon decrease with increase in organic carbon reduction was noted
246 in the degradation study carried out in batch-I (with nutrients) (Fig-15). But the reduction of
247 inorganic carbon present in the effluent is negligible in batch-I. The results demonstrate that the
248 algae require initially some amount of inorganic carbon for their growth [4]. The oxygen evolved
249 from the photosynthesis initiates the breakdown of phenol and makes the algae to utilize it
250 further. In case of external nutrient starved condition(batch-II), the reduction pattern was
251 reversed (Fig- 17); which clearly indicates that the utilization of inorganic carbon by the algae
252 enhanced when the species is in nutrient starved condition, since, the algal cells utilize the
253 carbon easily from inorganic carbon than from the organic carbon donor such as phenol.

254 On comparing both the batches, it can be observed that the fresh water microalgae *Scenedesmus*
255 sp. is able to utilize the organic carbon present in effluent as their carbon source for their growth
256 rather than the inorganic carbon supplemented externally. Further, the reduction was only by the
257 microalgae *Scenedesmus* sp. as confirmed by the results obtained for the controls.

258 **4.6 Discussion**

259 *Chlorella* sp., *Scenedesmus obliquus* and *Spirulina maxima* degraded about 1000 ppm of
260 synthetic phenol in less than 6 days. *Chlorella* and *Scenedesmus* needs an adaptation period of a
261 few days before starting the degradation beyond 400 ppm [1]. Wet biomass of fresh water
262 microalgae *Scenedesmus* sp. exhibits average of more than 95 % degradation of phenol
263 (C₆H₅OH) present in the effluent of producer gas plant in with nutrient batch having
264 concentration of algae 4g of wet biomass/liter of effluent within 7 days. The analysis of total
265 organic content indicates that the media constituents (carbonates and nitrate) added in the
266 effluent enhances the growth and enhances the phenol removal by the microalgae. Due to this,
267 the reduction of phenol increases with the addition of nutrients to the effluent during
268 degradation. The degradation of phenol requires molecular oxygen to initiate the aromatic
269 cleavage [2], [3]; that oxygen is gained by the phenol from the byproduct of photosynthesis of
270 microalgae. The addition of nutrient externally is required by the algae as confirmed by
271 analyzing the batch II in which the reduction of inorganic carbon is significantly more compared
272 to the nutrient rich batch. The increase of pH demonstrates that utilization of carbon dioxide
273 (CO₂) leads to disinfection [3]. There were no physical changes identified during the degradation
274 in the cells of *Scenedesmus* sp. Removal of contaminants present in the effluent other than
275 phenol could also be noticed when the *Scenedesmus* sp. was externally supplied with nutrients.

276

277 **5.0 Conclusion**

278 Phycoremediation of phenolic effluents is gaining attention globally. Compared to conventional
279 wastewater treatment (physical and chemical) processes, Phycoremediation can potentially
280 achieve nutrient removal in a less expensive and ecologically safer way with the added benefits
281 of resource recovery and recycling [15]. The unique properties of the algae in waste water
282 treatment integrated with conventional methods present great opportunities to the water and
283 wastewater treatment technologies. The present study proves that the *Scenedesmus* sp. is a
284 promising alternative to the conventional methods for phenol removal from the industries
285 producing phenol rich effluents without prior treatment.

286

287

288 6.0 Acknowledgement

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338 **Abbreviation**

- 339 ppm- parts per million
- 340 mg/l- milligram per liter
- 341 OC-Organic carbon
- 342 IC-Inorganic carbon
- 343 TC-Total carbon
- 344 TOC –Total organic carbon
- 345 TSS-Total dissolved solids
- 346 TDS-Total Dissolved solids
- 347 COD-Chemical Oxygen Demand
- 348 BOD-Biological Oxygen Demand