1	Antibiotic resistance genes variation pattern and microbial community succession
2	during swine manure composting under different aeration strategies
3	
4	Submitted to the NAXOS 2018
5	6th International Conference on Sustainable Solid Waste Management
6	May 2018
7	
8	
9	Jiali Chang <sup>a</sup> , Tao Jiang <sup>a,b*</sup> , Mengxin Zhao <sup>c</sup> , Juan Yang <sup>a</sup> , Zhiguo Wen <sup>a</sup> , Feng Yang <sup>a</sup> ,
10	Xuguang Ma <sup>a</sup> , Guoxue Li <sup>b*</sup>
11	
12	<sup>a</sup> College of Chemistry, Leshan Normal University, Leshan 614004, China
13	<sup>b</sup> College of Resources and Environmental Sciences, China Agricultural University,
14	Beijing 100193, China
15	<sup>c</sup> School of Environment, Tsinghua University, Beijing 100084, China

<sup>\*</sup>Corresponding author. Tel: +86 833 2270785, Fax: +86 +86 833 2270785, E-mail: composting@163.com

# 16 Abstract

17	This study adopted 3 common composting aeration methods, including forced
18	aeration (with 3 ventilation rate of $0.0/0.1/50.25 \text{ L} \cdot \text{kg}^{-1}\text{DM} \cdot \text{min}^{-1}$ respectively), turn
19	windrow and static pile; with the aim to clarify the actual effects of aeration methods on
20	the removal of antibiotic resistance genes (ARGs) in swine manure composting.
21	Representative ARGs abundances and microbial communities were determined via
22	qPCR and 16S rRNA gene high-throughput sequencing technology. Results showed that
23	most of ARGs (including tetW/tetO/tetH, qnrS, ermB and bla <sub>TEM</sub> ) in swine manure
24	could be significantly reduced to low levels after composting for 77 days, but
25	sulfonamide resistance genes (sull/sulII) and integrase gene (intI) were highly enriched,
26	closely correlated with the increase of genera belonged to Actinobacteria,
27	Proteobacteria and Bacteroidetes and the decrease of genera affiliated with Firmicutes.
28	Temporal variations of ARGs were consistent under different aeration treatments.
29	Further study should pay more attention to the recalcitrant ARGs in compost.
30	
31	Keywords: Compost; Swine manure; Antibiotic resistance gene; Microbial community;
32	Aeration methods.

### 34 **1. Introduction**

The world is undergoing the antibiotic crisis and entering a "post-antibiotic" era. The bloom of antibiotic resistance is threatening human health [1, 2]. China, as one of the largest antibiotic production and consumption country, the situation is even more serious due to the widespread overuse of antibiotics in clinic and agriculture [3]. As the vital initiator of the antibiotic crisis, the antibiotic resistance genes (ARGs) have been recently identified as emerging environmental contaminants [4], subsequently arousing a burst of attentions in related research fields.

In agriculture, the fate and dissemination pathways of ARGs have been largely investigated [5-7]. Animal manure has been reported as one of the significant reservoirs of ARGs due to the abuse of veterinary antibiotics as disease treatment or prevention and growth promotion agent. Among the multifarious types of livestock breed, swine manure was especially obtained much more researches in terms of ARGs occurrence state. Compared with other manures, swine manure harbored relatively more diverse ARGs with higher abundances and mobility that was positively correlated with antibiotics and heavy metal [8, 9], which was deserved further studies on its ARGs removal.

48 Composting process has been declared to reduce the diversity and quantity of ARGs in swine manure 49 to certain extent [10]. Numerous studies focused on enhancement of ARGs attenuation via composting 50 with versatile methods, e.g., adding biochar [11], surfactants [12], natural zeolite [13], superabsorbent 51 polymers [14], co-composting with Chinese medicinal herbal residues [15], vermicomposting using 52 housefly larvae [16] and continuous thermophilic composting [17]. However, controversy about the actual 53 effect of composting to ARGs were still existed. For instance, it was found that efflux pump tetracycline 54 ARGs, sulfonamide ARGs and integrase gene *intl* generally increased significantly after manure 55 composting for 32 days with 3.5 days of thermophilic periods [18]. Another study also indicated that 56 composting process enriched ARGs instead of elimination using sewage sludge as raw materials [19]. 57 Hence, it should be necessary to further explore the composting method itself in terms of monitoring the

58 fate of ARGs.

Mounting evidences verified that microbial community structure was the core determinants of ARGs variation pattern during composting [12, 15-16]. Considering that the environmental factors correlated with microbial growth and metabolism were distinct under different composting methods, this study adopted 5 composting methods (including static pile, turn windrow and forced aeration with 3 different ventilation rate of 0.05/0.1/0.25 L·kg<sup>-1</sup>DM·min<sup>-1</sup>) and mainly focused on the temporal variation of ARGs and also the succession of microbial community structures, with the aim to answer questions below: 1)

- 65 Whether ARGs were eliminated or enriched after composting? 2) Which composting method was best in
- ARGs removal? 3) What was the relationship between ARGs pattern and microbial community structure?

#### 67 2. Materials and methods

#### 68 2.1. Raw materials and composting instrument

- 69 Swine manure and cornstalk were collected as mentioned previously [20]. Rotting boxes
- 70 (1.08×0.8×1.4 m) were used to simulate central parts of composting piles in practice. The detail
- 71 information was shown in Fig. 1.



- 4. Bottom board with aeration holes;
- 78 5. Compost materials.

79

# 80 2.2. Experimental design and sampling

In total, 5 different composting methods (including static pile, turn windrow and forced aeration with 3 ventilation rate of 0.05/0.1/0.25 L·kg<sup>-1</sup>DM·min<sup>-1</sup>) were conducted for each rotting box. The static pile was under no artificial interference, neither forced ventilation nor turning occurred during the whole process. The turning windrow and forced aeration treatments were all turned weekly. All piles were composted for 77 days. Sampling was carried out on 3 time points of day 21, day 49 and day 77 before turning. 9 samples of

87 each composting system were randomly collected through the sampling holes on the front wooden boards,

- 88 3 of which were thoroughly mixed as one sample. Bulk cornstalk residues were picked out to avoid the
- 89 disturbance introduced by plant. Then the left fresh composting samples were stored in -80 °C for further
- 90 molecular experiment.

### 91 **2.3. DNA extraction**

92 In total, 16 groups of samples (including the original swine manure and compost samples at day 21,

49 and 77 of the five treatments) in triplicate were ready to extract DNA. Before the extraction, samples

94 were pre-washed three times by phosphate buffer solution (pH 8.0) and 1.7% polyvinylpyrrolidone K30

to remove impurity like humic acids. Then DNA was extracted from 0.2 g of each sample via the

96 Fast-DNA Spin Kit for Soil (MP Biomedicals, Santa Ana, CA) according to the manufacturer's

97 instructions. DNA concentrations and purity were determined via the NanoDrop® Spectrophotometer

98 ND-1000 (Thermo Fisher Scientific, MA, USA).

### 99 2.4. Quantitative PCR (qPCR)

100 In this study, 8 ARGs including resistence genes of tetracycline (*tetW/tetO/tetH*), sulphonamide

101 (*sull/sull1*), fluoroquinolone (*qnrS*), macrolide (*ermB*), beta-lactam (*bla<sub>TEM</sub>*) and one integrase gene (*int1*)

102 were determined by qPCR. Besides, the total bacteria was quantified through the general 16S rRNA gene.

103 The detail information of the primers targeted the above genes were shown in Table 1.

104

Table 1 Detail information of primers and qPCR condition used in this study

Target	Prime	rime Sequence(5'-3')		Product	qPCR
Gene			temperature (°C)	length (bp)	efficiency (%)
16S	F1048	GTGSTGCAYGGYTGTCGTCA	60	146	87.4
rRNA	R1194	ACGTCRTCCMCACCTTCCTC			
sulI	sul(I)-FX	CGCACCGGAAACATCGCTGCAC	55	163	95.1
	sul(I)-RX	TGAAGTTCCGCCGCAAGGCTCG			
sulII	sul(II)-FX	TCCGGTGGAGGCCGGTATCTGG	58	190	94.6
	sul(II)-RX	CGGGAATGCCATCTGCCTTGAG			
tetW	tet(W)-FX	GAGAGCCTGCTATATGCCAGC	60	168	100.2
	tet(W)-FV	GGGCGTATCCACAATGTTAAC			
tetO	tetO-F	ACGGARAGTTTATTGTATACC	57	171	84.1
	tetO-R	TGGCGTATCTATAATGTTGAC			
tetH	tetH-F	CAGTGAAAATTCACTGGCAAC	60	185	105.7

	tetH-R	ATCCAAAGTGTGGTTGAGAAT			
qnrS	qnrSrtF11	GACGTGCTAACTTGCGTGAT	62	240	91.4
	qnrSrtR11	TGGCATTGTTGGAAACTTG			
ermB	erm(B)-91f	GATACCGTTTACGAAATTGG	58	364	90.0
	erm(B)-454r	GAATCGAGACTTGAGTGTGC			
bla <sub>TEM</sub>	RTblaTEMFX	GCKGCCAACTTACTTCTGACAACG	60	247	93.8
	RTblaTEMFR	CTTTATCCGCCTCCATCCAGTCTA			
intI	intI-F	GGCTTCGTGATGCCTGCTT	55	145	80.1
	intI-R	CATTCCTGGCCGTGGTTCT			

105 Before qPCR detection, DNA standards for the absolute quantification were prepared from purified 106 plasmid DNA containing the targeted gene ranging from  $1.0 \times 10^2$  to  $1.0 \times 10^8$  copies  $ul^{-1}$ . The standard 107 plasmid was constructed as follows. Firstly, targeted gene was amplified using the above extracted DNA 108 as template. Then PCR products were purified and ligated into the pMD19-T vector (TaKaRa) according 109 to the manufacturer's instructions. The recombinant plasmids were then transformed into Escherichia coli 110 JM109 competent cells (TaKaRa) and the selected positive clones were sequenced using ABI 3730 XL 111 sequencing platform (Majorbio, Shanghai, China). After phylogenetic identification of the targeted gene, 112 the plasmids were re-extracted from positive clones. The initial copy numbers of plasmid DNA were calculated as follows:  $6.02 \times 10^{23} \times (\text{plasmid DNA concentration, ng/µl}) \times 10^{-9} (\text{plasmid DNA length, bp})$ 113  $\times$ 660. Then gradient diluted from  $1.0 \times 10^8$  to  $1.0 \times 10^2$  copies  $\cdot$  ul<sup>-1</sup>, ready for the next qPCR. 114 115 qPCR were carried out in 7500 real-time PCR system (Applied Biosystems) using primer pairs listed

in Table 1. Specifically, it was performed in a total volume of 25 µl, containing 12.5 µl of 2×Power
SYBRs Green PCR Master mix kit (Applied Biosystems), 0.5 µl of each primer (10 µM), 0.5 µl of bovine
serum albumin (10 mg/ml) and 2 µl of DNA (20:1 dilution). Each measurement was performed in three
replicates.

#### 120 2.5. 16S rRNA gene high-throughput sequencing

121 The 16S rRNA gene sequencing was performed at Majorbio Technology (Shanghai, China) using the

122 Illumina Miseq platform. Microbial community structures were determined using primers 338F

123 (5'-barcode-ACTCCTACGGGAGGCAGCA-3') and 806R (5'-GGACTACHVGGGTWTCTAAT-3')

124 targeted the V3-V4 region of 16S rRNA gene [21]. The raw paired-end reads were merged by Flash [22].

125 Then the obtained data were analyzed using QIIME pipeline [23]. The representative sequences of each

126 OTU were picked using uclust method. The taxonomy was annotated with the RDP classifier.

# 127 **2.6.** Data analysis

128 The redundancy analysis (RDA) of ARGs and its relationship with physicochemical factors

- 129 (temperature, pH, moisture and O<sub>2</sub>) were performed using CANOCO5 (Microcomputer Power, Ithaca,
- 130 NY). The heatmap of variation pattern of representative ARGs was based on the ratio of absolute quantity
- 131 of each ARG with the total bacterial copy numbers. Both the heatmaps of ARGs and microbial
- 132 community structure at genus level were performed via RStudio (RStudio Inc., Boston, Massachusetts).
- 133 The network analysis was performed using Cytoscape 3.4.0. Only correlations with coefficient (p-value)
- above 0.6 and a significance level (P-value) below 0.05 were displayed in the networks.
- 135 **3. Results and discussion**

136 3.1. Variation pattern of representative ARGs during composting under different aeration strategies

- 137 In this study, all of the 8 representative ARGs (including *bla<sub>TEM</sub>*, *ermB*, *qnrS*, *tetH/tetO/tetW* and
- 138 sull/sull) and integrase gene intl were detected in the original swine manure, with the absolute quantities
- ranged from  $1.2 \times 10^8$  to  $9.84 \times 10^{10}$  copies per gram of dry compost (Fig. 2), which was at the same level of

7



ARGs content in previous studies [14, 17].

**Fig. 2** Variation patterns in terms of absolute quantities of each ARG during the whole composting process among five different managements. Fig. 2a-f was corresponding to *bla*<sub>TEM</sub> (a), *ermB* (b), *tetW/O/H* (c), *qnrS* (d), *sull/II* (e) and *intI* (f), respectively. Bars represent standard deviations (n = 3). Note: origin, original swine manure; the number followed "Aeration" indicated the ventilation rate; Turn, turn

157 windrow management; Static, static pile management.

158 The relative abundance of *qnrS* was the most dominant 159 (31%) in swine manure, followed by *sull* (27%), *intl* (16%),

160 *ermB* (9%), *sulII* (7%) and *tetO/tetW* (8% / 3%), while that



161	of $bla_{TEM}$ and $tetH$ were less than 0.5% of the total ARGs (Fig. 3). Even though the detection of such high
162	content of qnrS was uncommon in other studies, it was indeed exhibited that antibiotics of
163	fluoroquinolone, sulfonamide, macrolide and tetracycline routinely feeding pig with therapeutic or
164	subtherapeutic concentration could enrich diverse and abundant ARGs in manure, posing potential risk
165	for direct application of manure into soil.
166	
167	
168	Fig. 3 The diversity patterns of the representative ARGs during the whole composting process among 5 different aeration methods.
169	Data were shown in average. Note: origin, original swine manure; the number followed "Aeration" indicated the ventilation rate;
170	Turn, turn windrow management; Static, static pile management.
171	
172	
173	
174	Similar with prior study [10], the situation turned to be better monitored after composting. The
175	diversity of ARGs in manure significantly decreased via composting, with the only overwhelming
176	dominance of sulfonamide resistance genes <i>sull/sull1</i> and integrase gene <i>int1</i> (Fig. 3), fluctuating around
177	$\sim 10^{11}$ copies per gram of dry compost (Fig. 2e&f). However, inconsistent with previous study reporting
178	the undetectable of most of ARGs at the end of composting, the seemingly disappearing ARGs were still
179	above the detection limitation during the whole composting process. Specifically, the original dominant
180	qnrS, ermB and tetO/W nearly dropped two orders of magnitude from $\sim 10^{10}$ to $\sim 10^8$ copies per gram of
181	dry compost, while the less abundant $bla_{TEM}$ and <i>tetH</i> were even detected with variable low levels of ~10 <sup>6</sup>

182 to  $\sim 10^8$  copies per gram of dry compost (Fig. 2a-d). Considering that ARGs could exist for long term,

183 either confined as genetic materials within a cell of the microbe or aggregated with humus in the form of

184 naked nucleotide acid fragments exempt from degradation by extracellular enzymes, it was rational to

185 deduce that composting could not completely remove ARGs from manures [24]. Nevertheless, compared

186 with spreading manure to crops directly, composting was still an efficient measurements in terms of

187 controlling ARGs dissemination, since the decline of most of ARGs instead of elimination could be

available to weaken the transmission possibility of ARGs in the environment [25, 26].

189 It had been reported that high temperature was key factor to control ARGs through killing or190 inhibiting microbes during the thermophilic phase of composting [17]. The thermophilic period with

- 191 temperature above 55 °C extended to almost 6 weeks in this study (data not shown) and even persisted
- along the whole process in the static pile treatment, especially longer than prior studies with high
- temperature less than 3 weeks [10, 18]. Here, ARGs contents were determined at 3 time points of the 5
- treatments, corresponding to mid-stage of thermophilic phase (day 21), early stage of maturing phase (day
- 49) and end state of experiment (day 77). Unexpectedly, the scenario of ARGs variation pattern seemed
- 196 complicated, neither exhibiting regular temporal dynamic nor clustering with composting aeration
- 197 methods (Fig. 4). Compared with original status, most of the representative ARGs including qnrS, ermB,
- 198 *tetO/W/H* and *bla<sub>TEM</sub>* decreased and kept at low levels among the 3 time points of different treatments,
- 199 while *sull/sull1* and *int1* were highly abundant all along the experiment (Fig. 3&4). It had been found that
- 200 continuous thermophilic composting with 40 days could effectively reducing the abundances of most of
- 201 ARGs [17], which reoccurred in our study with extraordinarily long-term of high temperature. However,
- the decrease of ARGs were not strengthened in static pile with longer thermophilic phase of 77 days.
- 203 Moreover, no obvious clue correlated aeration rate and aerobic status with the decline of ARGs. The
- 204 removal patterns of each ARG in different aeration methods were consistent, hence it was tricky to screen
- 205 out the best methods in terms of ARGs controlling.



Fig. 4 The heatmap of ARGs variation pattern during the whole composting process among five different managements. Data were analyzed based on the ratio of absolute quantities of ARGs and 16S rRNA gene. Note: origin, original swine manure; the number followed "Aeration" indicated the ventilation rate; Turn, turn windrow management; Static, static pile management.

- 216
- 217
- 218 Even though 6/9 genes detected in this study exhibited removal trend via composting, it was
- 219 noteworthy that sulphonamide resistance genes *sull/sulII* and integrase gene *intI* were enriched instead of
- 220 elimination via composting. With regard to the frequent detection of *sull/sull1* and *int1* in high abundance
- 221 from manure and compost samples, further study should be carried out to focus on the removal of these

222 recalcitrant genes and relevant evaluation of the risk it might be imposed to the accepted environment like
223 soil and water.

# 224 3.2. Microbial community succession under different aeration methods

After normalization, 46788 sequences of each sample were randomly selected to further analyze the

Samples	ACE	Chao1	Shannon	Simpson	Coverage
Origin	868	887	4.38	0.04	99%

- 226 microbial community structure. In total, 1373 OTUs were obtained from the 11 samples with 0.97
- similarity level. The good's coverages were all above 99% (Table 2), indicating the sequencing depth of
- this study was reliable to reflecting the microbial community structure. The richness and diversity indexes
- 229 were shown in table 2. Compared with original raw material, the microbial richness conveyed via ACE
- and Chao1 was decreased after composting, especially in the mature ending point. However, the diversity
- did not change a lot among different aeration methods except the turn windrow treatment at day 77 with a
- slight reduction.
- 233

Table 2 Microbial richness and diversity indexes

Aeration-0.25-day21	855	853	4.50	0.04	99%
Aeration-0.1-day21	674	701	4.57	0.02	99%
Aeration-0.05-day21	754	779	4.53	0.03	99%
Turn -day21	753	754	4.35	0.03	99%
Static-day21	718	695	4.56	0.02	99%
Aeration-0.25-day77	604	600	4.08	0.05	100%
Aeration-0.1-day77	605	610	4.57	0.02	99%
Aeration-0.05-day77	611	606	4.51	0.03	99%
Turn-day77	539	536	3.88	0.07	100%
Static-day77	676	656	4.37	0.03	99%

234 The microbial community composition at phylum level kept steady and consistent across the whole 235 process and among different aeration modes, with the main dominant members containing Firmicutes, 236 Actinobacteria, Proteobacteria and Bacteroidetes, accounting for 89-97% of the total sequences (Fig. 5a), 237 which had also been frequently detected with high abundances in previous studies [14-17]. Moreover, the 238 microbial community structures were undergone similar temporal successions among different aeration 239 treatments. Specifically, compared with the original swine manure, the relative abundance of Firmicutes 240 was increased at high temperature stage (day 21) but decreased at the end of the experiment (day 77), owing to the fact that most of members belonged to Firmicutes were thermophile. Actinobacteria, 241 242 Proteobacteria and Bacteroidetes were especially promoted at maturity stage in all treatments, which 243 might be correlated with the enrichment of *sull/sull1* and *int1*. The dominance of Actinobacteria at day 77 244 marked the maturation status of composting [27]. However, Actinobacteria was also well-known for its 245 capability to produce antibiotics and harbor diverse ARGs [28]. Hence, it should be cautious to evaluate 246 the compost quality with the abundance of Actinobacteria. 247 Temporal succession of microbial community was prominent at genus level. According to the 248 heatmap shown in Fig. 5b, the core microbial community was significantly distinct among and clustered 249 with the composting stage, but not the aeration modes. Specifically, cluster 1 including Ruminofilibacter 250 (8%, Bacteroidetes), Ignatzschineria (3%, Proteobacteria) and Streptococcus / Lactobacillus /

251 Sporosarcina (4% / 2% / 2%, Firmicutes) were mainly emerged in original swine manure, but declined to

- lower abundances with less than 0.5% after composting (Fig. 5b). The succession of these bacteria might
- contribute to the decrease of diversity pattern of ARGs via composting (Fig. 3). During the composting
- process, the core microbial community shifted dramatically from cluster 2 of thermophilic phase to
- cluster 3 of maturity state. In cluster 2, Bacillus, Clostridium sensu stricto 1, Oceanobacillus,
- Paucisalibacillus, Sinibacillus, Terrisporobacter, Turicibacter, uncultured Thermoactinomycetaceae and
- OPB54 with no rank accounted for the increase of *Firmicutes* at high temperature stage, mostly due to
- their property of thermo-tolerance. While distinct genus affiliated with Actinobacteria were abundant in
- day 21 and day 77 respectively, with frequent detection of Corynebacterium, Dietzia, Georgenia and
- Saccharomonospora in cluster 2 but Actinomadura, Longispora, Thermobifida and uncultured

Acidimicrobiales in cluster 3. Such shifts were also reported in prior researches, which was attributed to

- temperature selection [29,30].





















# 292 3.3. Correlation of representative ARGs and microbial community structure

Numerous studies had highlighted the role of microbial community on ARGs pattern in manure and compost samples [13,15-16]. In this study, diverse microbes were significantly negative or positive correlated with most of the detected ARGs except the less abundant *tetH* gene (P < 0.05). It was noticeable that different genera under the same phylum correlated with distinct ARGs. As shown in Fig. 6, 2 of 12 genera belonged to Bacteroidetes (Ruminofilibacter and Petrimonas) were potential carriers of qnrS, ermB, tetO and tetW, while the other genera were mainly positive correlated with sulII and intI, among which intI were co-occurred with sulII via the connection of Flavobacterium and unclassified Flavobacteriaceae. Moreover, the decrease of different genera of Firmicutes could lead to the decline of *qnrS/ermB/tetO/tetW* and the enrichment of *sulII*, respectively. In terms of *sulI*, there was only one negative correlation established with Hahella affiliated with Proteobacteria. Such weak relationship with microbial community was probably due to the redundant existence of *sull* highly widespread in diverse microbes, rendering it to persist for long term undisturbed by the variation of microbial community [26]. Generally, the enrichment of sull/sulli/intl at the end of the composting was correlated with the increase of genera belonged to Actinobacteria, Proteobacteria and Bacteroidetes and the decrease of genera affiliated with Firmicutes. Given that the microbial succession exhibited similar trends among different treatments, it was well explained the consistent variation patterns of ARGs under different aeration strategies.



Actinobacteria
 Bacteroidetes
 Chloroflexi
 Firmicutes
 Proteobacteria
 Saccharibacteria
 resistant-gene

- 310 Fig. 6 Network analysis based on the co-occurrence of representative ARGs and their potential host bacteria. A connection
- 311 represents a significant positive correlation (P < 0.05) according to Spearman's rank analysis. The blue and red lines were
- 312 designated with negative and positive correlations, respectively.

# 313 4. Conclusion

- 314 Combined with the above results, the questions listed in the introduction could be answered as
- 315 follows. Firstly, composting was an efficient tool in terms of reducing most of ARGs in original swine
- 316 manure, but sulfonamide resistance gene (*sull/sulII*) and integrase gene (*intI*) were enriched instead of
- 317 removal. Secondly, since the aeration methods showed little influence on the variation patterns of ARGs,
- 318 especially the intractable stubborn ARGs like *sull/sulli/intl*, innovative methods should be further
- 319 developed for the elimination of these genes in compost. Lastly, the variation pattern of ARGs was highly
- 320 depended on the succession of microbial community.
- 321

# 322 Acknowledgments

- 323 This investigation was supported by China Agriculture Research System (CARS-39-19), and the
- 324 National Key Research and Development Program of China (2017YFD0800202).
- 325

# 326 References

327 [1] Davies, J., Davies, D. 2010. Origins and evolution of antibiotic resistance. *Microbiol Mol Biol Rev*,

**74**(3), 417-33.

- Reardon, S. 2014. Antibiotic resistance sweeping developing world: bacteria are increasingly dodging
   extermination as drug availability outpaces regulation. Nature, 509(7499), 141-2.
- [3] Hao, R., Zhao, R., Qiu, S., Wang, L., Song, H. 2015. Antibiotics crisis in China. Science, 348(6239),
  1100-1.
- 333 [4] Pruden, A., Pei, R., Storteboom, H., Carlson, K.H. 2006. Antibiotic resistance genes as emerging
- contaminants: studies in northern Colorado. Environmental Science & Technology, 40(23), 7445.
- 335 [5] Chee-Sanford, J.C., Mackie, R.I., Koike, S., Krapac, I.G., Lin, Y.F., Yannarell, A.C., Maxwell, S.,
- 336 Aminov, R.I. 2009. Fate and transport of antibiotic residues and antibiotic resistance genes following

land application of manure waste. J Environ Qual, 38(3), 1086-108.

- 338 [6] Heuer, H., Schmitt, H., Smalla, K. 2011. Antibiotic resistance gene spread due to manure application
- 339 on agricultural fields. Current Opinion in Microbiology, 14(3), 236.
- [7] Youngquist, C.P., Mitchell, S.M., Cogger, C.G. 2016. Fate of Antibiotics and Antibiotic Resistance
   during Digestion and Composting: A Review. Journal of Environmental Quality, 45(2), 537.
- 342 [8] Zhu, Y.G., Johnson, T.A., Su, J.Q., Qiao, M., Guo, G.X., Stedtfeld, R.D., Hashsham, S.A., Tiedje,
- J.M. 2013. Diverse and abundant antibiotic resistance genes in Chinese swine farms. Proc Natl Acad
  Sci U S A, 110(9), 3435-40.
- 345 [9] Qian, X., Gu, J., Sun, W., Wang, X.J., Su, J.Q., Stedfeld, R. 2018. Diversity, abundance, and
- persistence of antibiotic resistance genes in various types of animal manure following industrial
   composting. J Hazard Mater, 344, 716-722.
- 348 [10] Selvam, A., Xu, D., Zhao, Z., Wong, J.W. 2012. Fate of tetracycline, sulfonamide and

349 fluoroquinolone resistance genes and the changes in bacterial diversity during composting of swine

- 350 manure. Bioresour Technol, 126, 383-90.
- 351 [11] Cui, E., Ying, W., Zuo, Y., Hong, C. 2016. Effect of different biochars on antibiotic resistance genes
- and bacterial community during chicken manure composting. Bioresource Technology, 203, 11-17.
- 353 [12] Zhang, Y., Li, H., Jie, G., Xun, Q., Yin, Y., Yang, L., Zhang, R., Wang, X. 2016. Effects of adding
- different surfactants on antibiotic resistance genes and intI1 during chicken manure composting.

- Bioresource Technology, 219, 545-551.
- [13] Zhang, J., Chen, M., Sui, Q., Tong, J., Jiang, C., Lu, X., Zhang, Y., Wei, Y. 2016. Impacts of
  addition of natural zeolite or a nitrification inhibitor on antibiotic resistance genes during sludge
  composting. Water Research, 91(9), 1286-91.
- 359 [14] Guo, A., Gu, J., Wang, X., Zhang, R., Yin, Y., Sun, W., Tuo, X., Zhang, L. 2017. Effects of
- superabsorbent polymers on the abundances of antibiotic resistance genes, mobile genetic elements,
   and the bacterial community during swine manure composting. Bioresour Technol, 244(Pt 1),
- and the bacterial community during swine manure composting. Bioresour Technol, 244(Pt 1),
  658-663.
- [15] Zhang, L., Gu, J., Wang, X., Sun, W., Yin, Y., Sun, Y., Guo, A., Tuo, X. 2017. Behavior of
  antibiotic resistance genes during co-composting of swine manure with Chinese medicinal herbal
  residues. Bioresour Technol, 244(Pt 1), 252-260.
- 366 [16] Wang, H., Sangwan, N., Li, H.Y., Su, J.Q., Oyang, W.Y., Zhang, Z.J., Gilbert, J.A., Zhu, Y.G., Ping,
- F., Zhang, H.L. 2017. The antibiotic resistome of swine manure is significantly altered by association
  with the Musca domestica larvae gut microbiome. Isme Journal, 11(1), 100-111.
- 369 [17] Qian, X., Sun, W., Gu, J., Wang, X.J., Sun, J.J., Yin, Y.N., Duan, M.L. 2016. Variable effects of
- oxytetracycline on antibiotic resistance gene abundance and the bacterial community during aerobic
   composting of cow manure. Journal of Hazardous Materials, 315, 61-69.
- 372 [18] Wang, J., Ben, W., Zhang, Y., Yang, M., Qiang, Z. 2015. Effects of thermophilic composting on
- 373 oxytetracycline, sulfamethazine, and their corresponding resistance genes in swine manure.
- Environmental Science Processes & Impacts, 17(9), 1654.
- 375 [19] Su, J.Q., Wei, B., Ou-Yang, W.Y., Huang, F.Y., Zhao, Y., Xu, H.J., Zhu, Y.G. 2015. Antibiotic

376 resistome and its association with bacterial communities during sewage sludge composting.

- Environmental Science & Technology, 49(12), 7356-63.
- 378 [20] Jiang, T., Ma, X., Tang, Q., Yang, J., Li, G., Schuchardt, F. 2016. Combined use of nitrification
- inhibitor and struvite crystallization to reduce the NH3 and N2O emissions during composting.
  Bioresource Technology, 217, 210-218.
- 381 [21] Dennis, K.L., Wang, Y., Blatner, N.R., Wang, S., Saadalla, A., Trudeau, E., Roers, A., Weaver, C.T.,
- 382 Lee, J.J., Gilbert, J.A., Chang, E.B., Khazaie, K. 2013. Adenomatous polyps are driven by

- microbe-instigated focal inflammation and are controlled by IL-10-producing T cells. Cancer Res,
  73(19), 5905-13.
- [22] Magoc, T., Salzberg, S.L. 2011. FLASH: fast length adjustment of short reads to improve genome
   assemblies. Bioinformatics, 27(21), 2957-63.
- 387 [23] Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K., Fierer,
- 388 N., Pena, A.G., Goodrich, J.K., Gordon, J.I., Huttley, G.A., Kelley, S.T., Knights, D., Koenig, J.E.,
- 389 Ley, R.E., Lozupone, C.A., McDonald, D., Muegge, B.D., Pirrung, M., Reeder, J., Sevinsky, J.R.,
- 390 Turnbaugh, P.J., Walters, W.A., Widmann, J., Yatsunenko, T., Zaneveld, J., Knight, R. 2010. QIIME
- allows analysis of high-throughput community sequencing data. Nat Methods, 7(5), 335-6.
- 392 [24] Forsberg, K.J., Patel, S., Gibson, M.K., Lauber, C.L., Knight, R., Fierer, N., Dantas, G. 2014.
- Bacterial phylogeny structures soil resistomes across habitats. *Nature*, **509**(7502), 612-6.
- 394 [25] Berendonk, T.U., Manaia, C.M., Merlin, C., Fatta-Kassinos, D., Cytryn, E., Walsh, F., Burgmann,
- H., Sorum, H., Norstrom, M., Pons, M.N., Kreuzinger, N., Huovinen, P., Stefani, S., Schwartz, T.,
- 396 Kisand, V., Baquero, F., Martinez, J.L. 2015. Tackling antibiotic resistance: the environmental
- 397 framework. *Nat Rev Microbiol*, **13**(5), 310-7.
- [26] Heuer, H., Smalla, K. 2007. Manure and sulfadiazine synergistically increased bacterial antibiotic
  resistance in soil over at least two months. Environmental Microbiology, 9(3), 657–666.
- 400 [27] Xiao, Y., Zeng, G.M., Yang, Z.H., Ma, Y.H., Huang, C., Xu, Z.Y., Huang, J., Fan, C.Z. 2011.
- 401 Changes in the actinomycetal communities during continuous thermophilic composting as revealed by
- 402 denaturing gradient gel electrophoresis and quantitative PCR. *Bioresour Technol*, **102**(2), 1383-8.
- 403 [28] Allen, H.K., Donato, J., Wang, H.H., Cloud-Hansen, K.A., Davies, J., Handelsman, J. 2010. Call of
- 404 the wild: antibiotic resistance genes in natural environments. *Nat Rev Microbiol*, **8**(4), 251-9.
- 405 [29] Hayakawa, M., Yamamura, H., Nakagawa, Y., Kawa, Y., Hayashi, Y., Misonou, T., Kaneko, H.,
- 406 Kikushima, N., Takahashi, T., Yamasaki, S. 2010. Taxonomic Diversity of Actinomycetes Isolated from
- 407 Swine Manure Compost. *Actinomycetologica*, **24**(2), 58-62.
- 408 [30] Steger, K., Jarvis, A., Vasara, T., Romantschuk, M., Sundh, I. 2007. Effects of differing temperature
- 409 management on development of Actinobacteria populations during composting. Res Microbiol, 158(7),
- 410 617-24.