# Ammonium uptake over analcime and its soil enhancer potential

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**Abstract:** Leaching of ammonium-nitrogen cause's eutrophication in water bodies and it evaporates easily as ammonia gas when applied as fertilizer. The use of industrial process based analcime to uptake ammonium (NH<sub>4</sub><sup>+</sup>) from synthetic and authentic wastewaters for soil enhancer purposes was studied in this paper. Sorption occurred during the first 20 minutes of contact time. NH<sub>4</sub><sup>+</sup> removal improved with increasing analcime dose and removal decreased when initial NH<sub>4</sub><sup>+</sup> concentration increased above 150 mg/L. Temperature did not have a clear impact on the NH<sub>4</sub><sup>+</sup> removal up to temperatures of 90°C. However, at 120°C, the NH<sub>4</sub><sup>+</sup> uptake increased considerably and the NH<sub>4</sub><sup>+</sup> concentration in analcime was 2.85 m%. This was explained by the formation of ammonioleucite. With real agricultural slurry the removal percentages were small. Results show that analcime releases NH<sub>4</sub><sup>+</sup> slowly during a long time period. It could be mixed directly to soil together with NH<sub>4</sub><sup>+</sup> containing fertilizer to reduce ammonia losses.

Keywords: analcime; zeolite; ammonium; sorption; soil enhancer

#### 1. Introduction

Zeolites are crystalline, hydrated aluminosilicate minerals with a three-dimensional framework, which consists of  $SiO_4$  and  $AIO_4$  tetrahedra linked with shared oxygen atoms. Zeolites possess large surface area and high cation exchange capacity and consequently they have been frequently applied as sorbents in water and wastewater treatment [1]. The most common types of zeolites are clinoptilolite, mordenite, dachiardite, analcime, phillipsite and heulandite. In addition to natural mineral sources, analcime  $[Na_{16}(Al_{16}Si_{32}O_{96}) \cdot 16H_2O]$  is formed as a by-product when lithium carbonate is produced from spodumene (LiAlSi<sub>2</sub>O<sub>6</sub>) using sodium pressure leaching process [2]. Zeolites have been reported to be effective in removing variable size cations from wastewaters, such as dyes [3,4], organics [5,6],  $NH_4^+$  [7,8], and metal ions [9,10]. Also analcime has been reported to be a sorbent in the wastewater treatment [11–13].

Nitrogen rotation in ecosystems is one of the most critical factors globally for both the environment and the food supply. Nitrogen is one of the main nutrients in wastewaters promoting the growth of organic matter and algae, which cause eutrophication in water bodies. In fact, ammonium-nitrogen (NH<sub>4</sub><sup>+</sup>) is more severe in term of eutrophication than the other common nitrogen species nitrate (NO<sub>3</sub><sup>-</sup>) [14]. Discharges of nitrogen are known e.g. for agriculture and for biogasification reject waters. Nitrogen is often present in the form of NH<sub>4</sub><sup>+</sup>, the ion which is not easy to precipitate, and therefore it is removed in the biological process by nitrification and denitrification. One challenge is also pH as NH<sub>4</sub><sup>+</sup>-ammonia equilibrium is pH-dependent and NH<sub>4</sub><sup>+</sup> volatilizes easily as ammonia gas at alkaline conditions [15]. Also, nitrogen fertilizers are produced mainly by using the energy-intensive Haber–Bosch reaction, which causes large amounts of greenhouse gas emissions, as production consumes high amounts of natural gas [16]. The challenge of adequacy of nutrients in use. However, if agricultural sludges and manures are applied to fields as such, up to 30% of the NH<sub>4</sub><sup>+</sup> is volatilized as ammonia [17]. Ammonia losses could be reduced for example by binding NH<sub>4</sub><sup>+</sup> to zeolites. Zeolites release NH<sub>4</sub><sup>+</sup> into the soil slowly and gradually not only in the first year of the vegetation period but also in the following years. Clinoptilolite is already used commercially in agriculture as an ammonium carrier [18]. Also natural analcime has shown potential in NH<sub>4</sub><sup>+</sup> uptake [19].

In the present study, the use of industrial process based analcime to uptake  $NH_4^+$  -nitrogen via sorption for soil enhancer purposes was studied. The effect of analcime dosage, contact time, temperature, and initial  $NH_4^+$  concentration on the  $NH_4^+$  uptake potential of analcime was studied. Also, the leaching of  $NH_4^+$  ions and the solubility of harmful elements from the analcime was studied to assess the potential of analcime to be applied as a soil improver.

# 2. Materials and methods

Analcime was received from a Finnish mining company. It was a side stream from lithium carbonate production. Analcime was dried in a hot air oven for 24–48 h at  $105 \pm 2^{\circ}$ C. Chemical composition of the analcime is presented in Table 1 and XRD diffractogram in Fig. 1. Main components were SiO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub> and Na<sub>2</sub>O. All the main spikes in the XRD diffractogram were associated with analcime.

Table 1 Chemical composition of analcime measured by XRF



Fig. 1 XRD diffractogram of the analcime. A = analcime

NH<sub>4</sub><sup>+</sup> uptake experiments were conducted in batch mode with synthetic NH<sub>4</sub><sup>+</sup> solution and agricultural slurry from a Finnish farm. Synthetic NH<sub>4</sub><sup>+</sup> solution was prepared from NH<sub>4</sub>Cl salt. Agricultural slurry contained 176 mg/L NH<sub>4</sub><sup>+</sup>, 25 mg/L phosphate and the pH value was 8.95. Analcime dosage was 5–500 g/L, concentration of synthetic NH<sub>4</sub><sup>+</sup> solution was 50–1000 mg/L, volume of NH<sub>4</sub><sup>+</sup> solution was 50 mL (150 mL in autoclave), and initial pH 2 (adjusted with HCl). Used temperatures were  $8 \pm 1^{\circ}$ C,  $22 \pm 1^{\circ}$ C,  $40 \pm 1^{\circ}$ C,  $60 \pm 1^{\circ}$ C,  $80 \pm 1^{\circ}$ C in water bath and 90  $\pm 1^{\circ}$ C and 120  $\pm 1^{\circ}$ C in autoclave. Samples were shaken or stirred during contact time. All NH<sub>4</sub><sup>+</sup> uptake experiments were performed as duplicate. Solids were separated from NH<sub>4</sub><sup>+</sup> solution by centrifuging before measuring NH<sub>4</sub><sup>+</sup> concentration with a Hach Lange IntelliCal NH<sub>4</sub><sup>+</sup> ISE- electrode.

The harmful element concentrations were characterized by aqua regia digestion according to standard ISO 11466 [20] and analyzed with ICP-OES technique (Perkin Elmer Optima 5300 DV). The 4-step sequential leaching procedure [21,22] used in this study is presented in Table 2. The eluates from each step were separated from the residue by centrifuging for 20 min (4500 rpm) and analyzed with ICP-OES.

Step	Fraction	Extractant	Experimental conditions			
F1	Water-soluble	$40 \text{ cm}^3 \text{ H}_2\text{O}$ at pH 4 (with HNO <sub>3</sub> )	16 h at $22 \pm 1^{\circ}$ C, constant shaking			
F2	Exchangeable and acid soluble	40 cm <sup>3</sup> HOAc 0.11 mol/L	16 h at $22 \pm 1^{\circ}$ C, constant shaking			
F3	Reducible	40 cm <sup>3</sup> NH <sub>2</sub> OH·HCl 0.5 mol/L	16 h at $22 \pm 1^{\circ}$ C, constant shaking			
		at pH 1.5 (with HNO <sub>3</sub> )				
F4	Oxidizable	10 cm <sup>3</sup> H <sub>2</sub> O <sub>2</sub> 300 g/L	1 h at $22 \pm 1^{\circ}$ C, occasional manual shaking, then 1 h 85 $\pm 2^{\circ}$ C. Reduce the volume to less than 3 cm <sup>3</sup>			
		$10 \text{ cm}^3 \text{ H}_2\text{O}_2 300 \text{ g/L}$	1 h at $85 \pm 2^{\circ}$ C. Reduce the volume to 1 cm <sup>3</sup>			
		50 cm <sup>3</sup> NH <sub>4</sub> OAc 1 mol/L at pH 2 (with HNO <sub>3</sub> )	16 h at 22 ± 1°C			

 Table 2 4-step sequential leaching procedure [21,22]

Neutralizing values were determined according to European standard EN 12945 [23]. The composition of the analcime was characterized using a sequential XRF (S4 Pioneer, Bruker AXS) and XRD (Rigaku SmartLab 9 kW with PDXL2 software). Nitrogen contents in the analcime after  $NH_4^+$  uptake were measured by using a Perkin Elmer 2400 series II CHNS/O analyser.

Release of  $NH_4^+$  ions from the analcime was tested after  $NH_4^+$  uptake experiments (1000 mg/L  $NH_4^+$  solution, 4 h contact time, temperature  $22 \pm 1^{\circ}C$ , and analcime dose 5 g/L) by shaking the analcime after ammonium uptake for 24 h at room temperature  $22 \pm 1^{\circ}C$ , either in distilled water (stages 1–2) or in 0.1 M NaCl solution (stages 3–10). Liquid was separated from the solid matter by centrifuging (20 min., 4500 rpm) and  $NH_4^+$  concentration was measured with a Hach Lange IntelliCal  $NH_4^+$  ISE- electrode. Procedure was repeated ten times (10 stages in total).

# 3. Results and discussion

# 3.1 NH<sub>4</sub><sup>+</sup> uptake experiments

# Effect of analcime dose

Effect of analcime dose (5-500 g/L) on the removal percentage of  $NH_4^+$  is presented in Fig. 2.  $NH_4^+$  model solution concentrations were 50 mg/L or 1000 mg/L with treatment time 4 h at room temperature ( $22 \pm 1^{\circ}C$ ). The  $NH_4^+$  removal percentage increased as the analcime dose increased. The  $NH_4^+$  removal was 37.4% as analcime dose was 200 g/L and 70.8% as analcime dose was 500 g/L. However, when analcime dose was 500 g/L, the pH increased to 9.5, which could partly explain the quite high removal percentage. With the 50 mg/L  $NH_4^+$  solution, the  $NH_4^+$  removal increased up to 53.1% (analcime dose 200 g/L). The sorption capacity (q) decreased with increasing analcime dose. The highest sorption capacity (3.55 mg/g) was achieved with analcime dose 20 g/L when initial  $NH_4^+$  concentration was 1000 mg/L. Final pH of the solution was below 7.5 up to analcime dose of 200 g/L, indicating that no losses of ammonia occurred during experiments.



Fig. 2 Effect of analcime dose on the removal percentage (a) and sorption capacity (b) of  $NH_4^+$ , and the effect of analcime dose on the final pH value (c).  $NH_4^+$  model solution concentrations were 50 mg/L or 1000 mg/L, treatment time 4 h at room temperature ( $22 \pm 1^{\circ}C$ )

#### Effect of initial NH4<sup>+</sup> concentration

Effect of initial NH<sub>4</sub><sup>+</sup> concentration (25-1000 mg/L) on the removal percentage of NH<sub>4</sub><sup>+</sup> is presented in Fig. 3. Analcime dose was 200 g/L with 4 h treatment time at room temperature ( $22 \pm 1^{\circ}$ C). The highest removal percentage (62.4%) was achieved as the initial NH<sub>4</sub><sup>+</sup> concentration was 150 mg/L and the lowest removal efficiencies (33.1–37.4%) were achieved when the initial NH<sub>4</sub><sup>+</sup> concentration was 800–1000 mg/L. However, the sorption capacity (q) increased with increasing NH<sub>4</sub><sup>+</sup> concentration (from q: 0.04 mg/g (25 mg/L) to 1.4 mg/g (1000 mg/L)). The increase of sorption capacity as a function of increasing initial NH<sub>4</sub><sup>+</sup> concentration indicate that suitable specific sorption sites are occupied at first [24]. The final pH of the solution decreased from 9.1 to 7.6 as initial NH<sub>4</sub><sup>+</sup> concentration increased from 25 mg/L to 1000 mg/L apparently because the increasing ionic strength caused by NH<sub>4</sub>Cl causes some buffering.



**Fig. 3** Effect of initial  $NH_4^+$  concentration on the removal percentage (a) and sorption capacity (b) of  $NH_4^+$ , and on the final pH value (c). Sorbent dose was 200 g/L, treatment time 4 h at room temperature ( $22 \pm 1^{\circ}C$ )

#### Effect of contact time

Effect of contact time (0-360 min.) on the removal percentage of  $NH_4^+$  is presented in Fig. 4. Removal percentage rose to around 40% during the first 20 minutes and settled there until 180 minutes. At the end of the experiment, the removal percentage was 56%. However, results showed that total sorption equilibrium was not reached within 360 minutes and therefore it is possible that some  $NH_4^+$  evaporated as ammonia gas, which would explain why removal percentage keeps rising after 180 minutes.



Fig. 4 Effect of time on the removal percentage of  $NH_4^+$ . Sorbent dose was 20 g/L and initial  $NH_4^+$  concentration was 100 mg/L at room temperature ( $22 \pm 1^{\circ}C$ )

### Effect of temperature

Effect of temperature (8-80°C) on the removal percentage of  $NH_4^+$  is presented in Fig. 5. The sorbent dose used in the experiments was either 20 g/L or 50 g/L, treatment time was 4 h and initial  $NH_4^+$  concentration 1000 mg/L. Experiments were conducted with several sorbent doses (20, 50, 100, 200, 500 g/L) but temperature did not have a clear effect on the removal percentage. The main conclusion is that  $NH_4^+$  removal does not improve as temperature is increased up to  $80 \pm 1^{\circ}$ C. However, Yuan et al. (2016) reported that the  $NH_4^+$  sorption capacity of analcime increased considerably as temperature was between 110–130 °C. Therefore, experiments were conducted also at 90 ± 1°C and 120 ± 1°C.



**Fig. 5** Effect of temperature on the removal percentage (a) and sorption capacity (b) of NH<sub>4</sub><sup>+</sup>, and on the final pH value (c). Sorbent dose was 20 g/l or 50 g/L, initial NH<sub>4</sub><sup>+</sup> concentration 1000 mg/L, and treatment time 4 h

Concentration in analcime after ammonium uptake at different temperatures is presented in Fig. 6. Even though some ammonium was sorbed, the concentration was at most 0.5 m% at temperatures up to  $90 \pm 1^{\circ}$ C. However, at  $120 \pm 1^{\circ}$ C the NH<sub>4</sub><sup>+</sup> sorption increased considerably up to 2.85 m% when sorbent dose was 10 g/L.



Fig. 6 Ammonium concentration in analcime after sorption at different temperatures and sorbent doses. Treatment time was 4 h and initial  $NH_4^+$  concentration 1000 mg/L at 22 ± 1°C and 60 ± 1°C and 2 h at 90 ± 1°C and 120 ± 1°C

#### Experiments with agricultural slurry

Sorption experiment results with agricultural slurry are presented in Fig. 7. In practice, no  $NH_{4^+}$  removal was achieved at 8 ± 1°C. The best removal percentage (around 30%) was achieved at room temperature with sorbent dose 500 g/L. Sorption capacity values (q) were small (below 3 mg/g) and they decreased as the sorbent dose increased. The pH remained below 7 in all cases, therefore no  $NH_{4^+}$  losses should have occurred [25].



**Fig. 7** Removal of NH<sub>4</sub><sup>+</sup> from agricultural slurry using different doses of analcime (50 g/L, 100 g/L, 200 g/L, and 500 g/L) at  $8 \pm 1^{\circ}$ C and  $22 \pm 1^{\circ}$ C. Removal percentage of NH<sub>4</sub><sup>+</sup> (a); sorption capacity of NH<sub>4</sub><sup>+</sup> (b), and final pH (c). Treatment time was 4 h and initial NH<sub>4</sub><sup>+</sup> concentration was 176 mg/L

#### 3.2 XRD analysis

Formation of ammonioleucite was confirmed by the XRD analysis of analcime after  $NH_4^+$  uptake at 120 °C (Fig. 8). This also explains why the effect of temperature was so high on the concentration of  $NH_4^+$  in the used analcime. Previous studies have indicated that the (Si,Al)O<sub>4</sub> framework of analcime is identical with the framework of ammonioleucite [26,27]. Since the ionic radius of  $NH_4^+$  is 1.43 Å but the effective tunnel size of analcime is only 2.6 Å,  $NH_4^+$  cannot pass through the tunnels at room temperature. Yamada et al. [26] suggested that the effective inside diameter of analcime is increased at high temperatures (above 100 °C) as a result of lattice vibration, which enables the  $NH_4^+$  to pass through the tunnel. Also, the increase of temperature causes the crystal water to escape from analcime and the  $NH_4^+$  ion in the formed ammonioleucite takes the same site as water molecule in analcime structure [19,26]. These findings are confirmed also in our study. However, the results clearly indicate that some  $NH_4^+$  is sorbed onto analcime also at lower temperatures but the presence of ammonioleucite was only confirmed after sorption at high temperatures.



Fig. 8 XRD diffractogram of the analcime after sorption. A = analcime; B = ammonioleucite

#### 3.3 Soil enhancer properties

Sequential leaching and aqua regia digestion results of the analcime before and after  $NH_4^+$  uptake experiments are presented in Table 3. The leaching of Cd, Cr, Cu, Ni, Pb, and Zn was very low in the sequential leaching, which means that they are not bioavailable. However, As is quite soluble in the first fraction indicating it is bioavailable but it seems that the sorption reaction reduces the bioavailability of As. Still, all concentrations (aqua regia digestion) of the harmful elements (As, Cd, Cr, Cu, Ni, Pb, Zn) are well below the limit values of Finnish fertilizer decree [28]. Therefore, it would be possible to use analcime as a soil improver in Finland. However, also the nutrient concentrations (Ca, K, P) are low. The neutralizing value of the analcime before sorption was 5.5 Ca%, which is in line with the Ca concentration. Therefore, the ability of analcime to raise the pH of acidic soil is low compared to lime (neutralizing value typically between 30–40 Ca%).

	Ca	K	Р	As	Cd	Cr	Cu	Ni	Pb	Zn
Before sorption										
F1	660	460	80	9.0	< 0.08	< 0.4	1.9	< 0.2	<0.6	1.7
F2	560	300	30	1.1	< 0.08	< 0.4	0.5	< 0.2	<0.6	0.4
F3	1000	460	410	1.4	< 0.08	0.7	1.3	< 0.2	0.9	1.9
F4	450	480	210	< 0.8	< 0.10	< 0.5	1.7	< 0.3	< 0.8	1.3
Aqua regia	6850	4700	850	16	< 0.3	60	30	21	3.8	70
After sorption										
F1	100	140	28	3.8	< 0.08	< 0.4	1.3	0.1	<0.6	0.9
F2	730	390	5.1	0.5	0.2	< 0.4	1.1	0.4	<0.6	0.3
F3	2200	970	730	4.1	0.4	< 0.4	3.1	0.3	<0.6	2.4
F4	690	650	170	< 0.8	< 0.10	0.8	0.1	0.6	< 0.8	1.1
Aqua regia	6270	3280	1060	15	1.4	68	45	19	4.1	100
Limit value (mg/kg) field fertilizers/forest fertilizers	-	-	-	25/40	1.5/25	300	600/700	100/150	100/150	1500/4500

 Table 3 Sequential leaching and aqua regia digestion results of the analcime before and after NH4<sup>+</sup> uptake (mg/kg) and a comparison to the limit values of Finnish Fertilizer decree [28]

The release of  $NH_{4^+}$  from analcime was studied by shaking the sample in distilled water (stages 1–2) and in 0.1 M NaCl solution (stages 3–10). Release of  $NH_{4^+}$  was around 315 mg/kg in first stage, which then reduced to less than 75 mg/kg in stage 2 (Fig 9.). With NaCl, the release of  $NH_{4^+}$  was fast at first (around 1200 mg/kg), decreased constantly in stages 4-7, and then settled to around 70–100 mg/kg in stages 8–10. Results confirm that the  $NH_{4^+}$  uptake was mostly caused by sorption and even though the conditions used here do not represent chemical composition of soil solution, they do confirm that analcime releases  $NH_{4^+}$  slowly during a long period of time. Because of the low  $NH_{4^+}$  concentration (Fig. 6), analcime does not have much value as an ammonium fertilizer as such but it could be used as a component in fertilizers to control the ammonia release in soil as Souza et al. [29] suggested for clinoptilolite. A reduction of up to 50% to the ammonia volatilization from urea in soil was achieved as clinoptilolite was pelletized with urea. As the enzymes in soil decompose urea, the part of the  $NH_{4^+}$  that is not instantly needed for plant growth is captured by clinoptilolite and analcime could be used in as similar way. In this case, analcime could be mixed directly to soil without the previous  $NH_{4^+}$  sorption. As urea or other  $NH_{4^+}$  containing fertilizer was applied on the field, analcime could capture part of the  $NH_{4^+}$  and release it gradually.



**Fig. 9** NH<sub>4</sub><sup>+</sup> release during leaching test. Distilled water was used in stages 1–2 and 0.1 M sodium chloride in stages 3–10

### 4. Conclusions

The use of industrial process based analcime to uptake  $NH_4^+$  from synthetic wastewater and real agricultural slurry via sorption for soil enhancer purposes was studied in this paper.  $NH_4^+$  sorption reaction was fast but the concentration of  $NH_4^+$  in analcime was low (0.5 m%) at temperatures up to  $90 \pm 1^{\circ}C$ . However, results indicate that  $NH_4^+$  uptake was mostly caused by sorption and analcime releases  $NH_4^+$  slowly during a long period of time. Therefore, analcime does not have much value as ammonium fertilizer but it could be used to control the ammonia release in soil. In this case, analcime could be mixed directly to soil without  $NH_4^+$  sorption. As urea or other  $NH_4^+$  containing fertilizer was applied on the field, analcime could capture part of the  $NH_4^+$  and release it gradually. However, this should be confirmed with field tests.

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#### 5. References

- 1. Wang, S., Peng, Y.: Natural Zeolites as Effective Adsorbents in Water and Wastewater Treatment. Chem. Eng. J. 156, 11–24 (2010)
- Chen, Y., Tian, Q., Chen, B., Shi, X., Liao, T.: Preparation of Lithium Carbonate from Spodumene by a Sodium Carbonate Autoclave Process. Hydrometallurgy 109, 43–46 (2011)
- 3. Karadag, D., Akgul, E., Tok, S., Erturk, F., Kaya, M.A., Turan, M.: Basic and Reactive Dye Removal Using Natural and Modified Zeolites. J. Chem. Eng. Data 52, 2436–2441 (2007)
- 4. Wang, S., Zhu, Z.H.: Characterisation and Environmental Application of an Australian Natural Zeolite for Basic Dye Removal from Aqueous Solution. J. Hazard. Mater. 136, 946–952 (2006)
- Bowman, R.S.: Applications of Surfactant-Modified Zeolites to Environmental Remediation. Microporous Mesoporous Mater. 61, 43–56 (2003)
- 6. Huttenloch, P., Roehl, K.E., Czurda, K.: Sorption of Nonpolar Aromatic Contaminants by Chlorosilane Surface Modified Natural Minerals. Environ. Sci. Technol. 35, 4260–4264 (2001)
- 7. Lei, L., Li, X., Zhang, X.: Ammonium Removal from Aqueous Solutions Using Microwave-Treated Natural Chinese Zeolite. Sep. Purif. Technol. 58, 359–366 (2008)

- 8. Luukkonen, T., Sarkkinen, M., Kemppainen, K., Rämö, J., Lassi, U.: Metakaolin Geopolymer Characterization and Application for Ammonium Removal from Model Solutions and Landfill Leachate. Appl. Clay Sci. 119, 266–276 (2016)
- 9. Cincotti, A., Mameli, A., Locci, A.M., Orrù, R., Cao, G.: Heavy Metals Uptake by Sardinian Natural Zeolites: Experiment and Modeling. Ind. Eng. Chem. Res. 45, 1074–1084 (2006)
- 10. Oter, O., Akcay, H.: Use of Natural Clinoptilolite to Improve Water Quality: Sorption and Selectivity Studies of Lead(II), Copper(II), Zinc(II), and Nickel(II). Water Environ. Res. 79, 329–335 (2007)
- Atta, A.Y., Jibril, B.Y., Aderemi, B.O., Adefila, S.S.: Preparation of Analcime from Local Kaolin and Rice Husk Ash. Appl. Clay Sci. 61, 8–13 (2012)
- 12. Mallah, M.H., Soorchi, H., Jooybari, T.F.: Development of Empirical Equation for Analcime in the Treatment of Nuclear Waste. Ann. Nucl. Energy 47, 140–145 (2012)
- Montalvo, S., Guerrero, L., Borja, R., Sánchez, E., Milán, Z., Cortés, I., Angeles de la la Rubia, M.: Application of Natural Zeolites in Anaerobic Digestion Processes: A Review. Appl. Clay Sci. 58, 125–133 (2012)
- 14. Wetzel, R.G.: The Nitrogen Cycle. In: Limnology, pp. 205–237. Academic Press, San Diego (2001)
- 15. Rumble, J.: CRC Handbook of Chemistry and Physics. CRC Press, Boca Raton (2018)
- 16. Elvers, B.: Ullmann's Encyclopedia of Industrial Chemistry: Wiley and Sons, Hoboken (2014)
- 17. Sommer, S.G., Olesen, J.E.: Modelling Ammonia Volatilization from Animal Slurry Applied with Trail Hoses to Cereals. Atmos. Environ. 34, 2361–2372 (2000)
- Reháková, M., Čuvanová, S., Dzivák, M., Rimár, J., Gaval'ová, Z.: Agricultural and Agrochemical Uses of Natural Zeolite of the Clinoptilolite Type. Curr. Opin. Solid State Mater. Sci. 8, 397–404 (2004)
- 19. Yuan, J., Yang, J., Ma, H., Liu, C.: Crystal Structural Transformation and Kinetics of /Na+ Ion-Exchange in Analcime. Microporous Mesoporous Mater. 222, 202–208 (2016)
- 20. ISO 11466 Soil Quality Extraction of Trace Elements Soluble in Aqua Regia. International Organization of Standardization, Geneva (1995)
- Rauret, G., López-Sánchez, J.F., Sahuquillo, A., Rubio, R., Davidson, C., Ure, A., Quevauviller, P.: Improvement of the BCR Three Step Sequential Extraction Procedure Prior to the Certification of New Sediment and Soil Reference Materials. J. Environ. Monit. JEM 1, 57–61 (1999)
- 22. Nurmesniemi, H., Pöykiö, R., Kuokkanen, T., Perämäki, P., Välimäki, I.: Sequential Leaching of Trace Elements in Bottom Ash from a Fluidized Bed Co-Combustion Boiler at a Pulp and Paper Mill Complex. J. Solid Waste Technol. Manag. 31, 115–121 (2005)
- 23. EN 12945 Liming Materials. Determination of Neutralizing Value. Titrimetric Methods. European Committee for Standardization, Brussels (2014)
- 24. Bhatnagar, A., Kumar, E., Sillanpää, M.: Nitrate Removal from Water by Nano-Alumina: Characterization and Sorption Studies. Chem. Eng. J. 163, 317–323 (2010)
- Zhao, Y., Niu, Y., Hu, X., Xi, B., Peng, X., Liu, W., Guan, W., Wang, L.: Removal of Ammonium Ions from Aqueous Solutions Using Zeolite Synthesized from Red Mud. Desalination Water Treat. 57, 4720– 4731 (2016)
- 26. Yamada, M., Miyawaki, R., Nakai, I., Izumi, F., Nagashima, K. A.: Rietveld Analysis of the Crystal Structure of Ammonioleucite. Mineral. J. 20, 105–112 (1998)
- 27. Hori, H., Nagashima, K., Yamada, M., Miyawaki, R., Marubashi, T.: Ammonioleucite, a new mineral from Tatarazawa, Fujioka, Japan. Am. Min. 71, 1022–1027 (1986)
- 28. Finnish Ministry of Agriculture and Forestry: Ministry of Agriculture and Forestry Decree on Fertilizer Products 24/11 (2011)
- Souza, I.M.S., Gurgel, G.C.S., Medeiros, A.M., Zonta, E., Ruiz, J.A.C., Paskocimas, C.A., Motta, F.V., Bomio, M.R.D.: The Use of Clinoptilolite as Carrier of Nitrogened Fertilizer with Controlled Release. J. Environ. Chem. Eng. 6, 4171–4177 (2018)